

Integrating biogeochemistry with multiomic sequence information in a model oxygen minimum zone

Stilianos Louca^a, Alyse K. Hawley^b, Sergei Katsev^{c,d}, Monica Torres-Beltran^b, Maya P. Bhatia^{b,e}, Sam Kheirandish^b, Céline C. Michiels^b, David Capelle^f, Gaute Lavik^g, Michael Doebeli^{h,i}, Sean A. Crowe^{b,f,j,1}, and Steven J. Hallam^{b,e,j,k,l,1}

^aInstitute of Applied Mathematics, University of British Columbia, Vancouver, BC, Canada V6T1Z2; ^bDepartment of Microbiology and Immunology, University of British Columbia, Vancouver, BC, Canada V6T1Z3; ^cLarge Lakes Observatory, University of Minnesota Duluth, Duluth, MN 55812; ^dDepartment of Physics and Astronomy, University of Minnesota Duluth, Duluth, MN 55812; ^eCanadian Institute for Advanced Research Program in Integrated Microbial Biodiversity, Canadian Institute for Advanced Research, Toronto, ON, Canada M5G1Z8; ^fDepartment of Earth, Ocean, and Atmospheric Sciences, University of British Columbia, Vancouver, BC, Canada V6T1Z4; ^gBiogeochemistry Group, Max Planck Institute for Marine Microbiology, Bremen D-28359, Germany; ^hDepartment of Zoology, University of British Columbia, Vancouver, BC, Canada V6T1Z4; ⁱDepartment of Mathematics, University of British Columbia, Vancouver, BC, Canada V6T1Z4; ^jEcosystem Services, Commercialization Platforms, and Entrepreneurship (ECOSCOPE) Training Program, University of British Columbia, Vancouver, BC, Canada V6T1Z3; ^kGraduate Program in Bioinformatics, University of British Columbia, Vancouver, BC, Canada V6T1Z3; and ^lPeter Wall Institute for Advanced Studies, University of British Columbia, Vancouver, BC, Canada V6T1Z2

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Microorganisms are the most abundant lifeform on Earth, mediating global fluxes of matter and energy. Over the past decade, high-throughput molecular techniques generating multiomic sequence information (DNA, mRNA, and protein) have transformed our perception of this microcosmos, conceptually linking microorganisms at the individual, population, and community levels to a wide range of ecosystem functions and services. Here, we develop a biogeochemical model that describes metabolic coupling along the redox gradient in Saanich Inlet—a seasonally anoxic fjord with biogeochemistry analogous to oxygen minimum zones (OMZs). The model reproduces measured biogeochemical process rates as well as DNA, mRNA, and protein concentration profiles across the redox gradient. Simulations make predictions about the role of ubiquitous OMZ microorganisms in mediating carbon, nitrogen, and sulfur cycling. For example, nitrite “leakage” during incomplete sulfide-driven denitrification by SUP05 Gammaproteobacteria is predicted to support inorganic carbon fixation and intense nitrogen loss via anaerobic ammonium oxidation. This coupling creates a metabolic niche for nitrous oxide reduction that completes denitrification by currently unidentified community members. These results quantitatively improve previous conceptual models describing microbial metabolic networks in OMZs. Beyond OMZ-specific predictions, model results indicate that geochemical fluxes are robust indicators of microbial community structure and reciprocally, that gene abundances and geochemical conditions largely determine gene expression patterns. The integration of real observational data, including geochemical profiles and process rate measurements as well as metagenomic, metatranscriptomic and metaproteomic sequence data, into a biogeochemical model, as shown here, enables holistic insight into the microbial metabolic network driving nutrient and energy flow at ecosystem scales.

metagenomics | metatranscriptomics | metaproteomics | biogeochemical | gene-centric model

Microbial communities catalyze Earth’s biogeochemical cycles through metabolic pathways that couple fluxes of matter and energy to biological growth (1). These pathways are encoded in evolving genes that, over time, spread across microbial lineages and today shape the conditions for life on Earth. High-throughput sequencing and mass-spectrometry platforms are generating multiomic sequence information (DNA, mRNA, and protein) that is transforming our perception of this microcosmos, but the vast majority of environmental sequencing studies lack a mechanistic link to geochemical processes. At the same time, mathematical models are increasingly used to describe local- and global-scale biogeochemical processes or predict future changes in global elemental cycling and climate (2, 3). Although these models typically incorporate the catalytic properties of cells, they fail to

integrate the information flow from DNA to mRNA, proteins, and process rates as described by the central dogma of molecular biology (4). Hence, a mechanistic framework integrating multiomic data with geochemical information has remained elusive.

Recent work based on metagenomics and quantitative PCR (qPCR) suggests that biogeochemical processes may be described by models focusing on the population dynamics of individual genes (5, 6). In such gene-centric models, genes are used as proxies for particular metabolic pathways, with gene production rates being determined solely by the Gibbs free energy

Significance

Modern molecular sequencing is beginning to provide great insight into microbial community structure and function at ecosystem scales. However, the quantitative integration of multiomic sequence information (DNA, mRNA, and protein) and geochemical processes has so far been elusive. Here, we develop a biogeochemical model that integrates geochemistry and multiomic sequence information to explain key metabolic processes in the oxygen-starved waters of Saanich Inlet, a model ecosystem for studying microbial community responses to oxygen minimum zone expansion. Our model largely explains DNA, mRNA, and protein distributions and sheds light on the metabolic networks coupling carbon, sulfur, and nitrogen transformations across a redox gradient. Our approach is extensible to other biogeochemical models incorporating feedbacks of global change on ecosystem functions.

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Data deposition: All sequencing data have been deposited in publicly accessible databases, as follows. The raw metagenomic sequence data reported in this paper have been deposited in the Joint Genome Institute (JGI) Genomes Online Database (GOLD) (Project ID nos. Gp0052414 and Gp0052384–Gp0052387). The metagenome assemblies reported in this paper have been deposited in the JGI Integrated Microbial Genomes and Microbiomes Samples Database (Taxon Object ID nos. 3300000198, 3300000213, 3300000255, 3300000256, and 3300000257). The raw metatranscriptomic sequence data reported in this paper have been deposited in the JGI GOLD (GOLD Project ID nos. Gp0055274–Gp0055277 and Gp0055281). The protein sequences reported in this paper have been deposited in the Mass Spectrometry Interactive Virtual Environment (MassIVE) accession no. MSV000079878; proteome exchange ID no. PXD004493). The KEGG (Kyoto Encyclopedia of Genes and Genomes) annotations of metagenomic contigs as well as detailed mass spectrometry run information are available as [Dataset S1](#) and [Dataset S2](#), respectively.

¹To whom correspondence may be addressed. Email: sean.crowe@ubc.ca or shallam@mail.ubc.ca.

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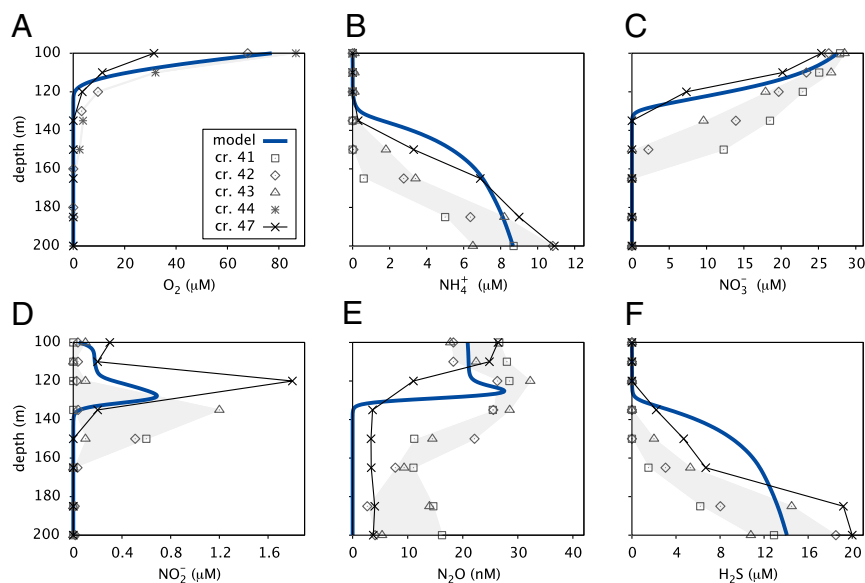


Fig. 2. Measured and predicted geochemical profiles. (A) Oxygen, (B) ammonium, (C) nitrate, (D) nitrite, (E) nitrous oxide, and (F) hydrogen sulfide concentrations, as predicted by the calibrated model at steady state (thick blue curves). Dots indicate data used for the calibration measured during cruise 41 (cr. 41) on January 13, 2010 (SI041_01/13/10; □), cr. 42 (SI042_02/10/10; ◇), and cr. 43 (SI043_03/10/10; △). Oxygen profiles were not available for cr. 41 and cr. 43; hence, data from cr. 44 (SI044_04/07/10; *) were used instead. Thin black curves indicate data measured during cr. 47 (SI047_07/07/10), shortly before deep water renewal. Details on data acquisition are in *SI Appendix, Section S1*.

during which the SNTZ continued to migrate upward in the water column (Fig. 2).

DNA Profiles and Process Rates

The calibrated model makes predictions about gene abundance and process rates, which can be validated using metagenomic sequence data and N process rate measurements from the same location and period as the geochemical calibration data (Fig. 3). Consistent with metagenomic depth profiles, the model predicts a redox-driven partitioning of pathways across the water column. Genes associated with ROM (ABC transporters mapped to dominant aerobic heterotrophs), aerobic ammonium oxidation to nitrite [ammonia monooxygenase (*amo*)], and aerobic nitrite oxidation to nitrate [nitrite oxidoreductase (*nxr*)] decline precipitously in deep basin waters, whereas genes associated with partial denitrification to nitrous oxide [PDNO; represented by nitric oxide reductase (*norBC*)], nitrous oxide reduction [nitrous oxide reductase (*nosZ*)], and anammox [hydrazine oxidoreductase (*hzo*)] are most abundant in the SNTZ (Fig. 3).

The similarity of the PDNO, *nosZ*, and *hzo* gene profiles is indicative of their strong metabolic interaction (Fig. 3A). In particular, the co-occurring peaks of PDNO and *nosZ* gene abundances in the absence of N_2O accumulation (Fig. 2E) reflect a quantitative coupling between the two denitrification steps and imply that both steps support extensive microbial growth at the SNTZ. This coupling is intriguing, because genomic reconstructions from both uncultivated and cultivated SUP05, the dominant denitrifier in Saanich Inlet, have not identified the *nosZ* gene (14, 17, 23). The absence of *nosZ* from SUP05 suggests that incomplete nitrate reduction by SUP05 and reduction of nitrous oxide by unidentified community members constitute separate and complementary metabolic niches in Saanich Inlet under low-oxygen and anoxic conditions (24).

The superposition of electron donor–acceptor pairs in redox transition zones supports chemical energy transfer in stratified water columns (25, 26), and previous studies have revealed relatively high cell abundances and chemoautotrophic activity in such zones (12, 27, 28). At the SNTZ in Saanich Inlet, the simultaneous availability of NO_3^- and H_2S fuels chemoautotrophic

nitrate reduction coupled to sulfide oxidation, in turn supplying anammox with NO_2^- via “leaky” denitrification (up to 88% of NO_2^- supplied by PDNO) (*SI Appendix, Section S2.11*). Most of the NH_4^+ used by anammox ($0.3 \text{ mmol} \cdot \text{m}^{-2} \cdot \text{d}^{-1}$), however, is predicted to originate from the underlying sediments and reach the SNTZ via eddy diffusion. Accordingly, both anammox and denitrification rates are predicted to peak around the SNTZ and lead to production of N_2 . This prediction is consistent with process rate measurements from discrete depth intervals during subsequent cruises in 2010 (Fig. 3B) as well as elevated SUP05 abundances at the SNTZ (Fig. 3A) (estimated via qPCR). In fact, the good agreement between predicted PDNO gene counts and observed SUP05 abundances suggests that energy fluxes associated with denitrification can be accurately translated to denitrifier growth rates. Predicted peak sulfide-driven denitrification rates are somewhat higher than peak anammox rates, although depth-integrated nitrogen loss rates are comparable for both pathways ($\sim 0.3 \text{ mmol} - N_2 \cdot \text{m}^{-2} \cdot \text{d}^{-1}$). These predictions are partly consistent with rate estimates derived directly from the geochemical profiles using inverse linear transport modeling (ILTM) (details in *Materials and Methods* and Fig. 3B). Hence, near steady-state conditions, coupled sulfide-driven denitrification and anammox can concurrently drive significant nitrogen loss in the water column.

The fraction of NO_2^- leaked during denitrification, compared with the total NO_3^- consumed ($L_{\text{PDNO}} = 0.352$) (*SI Appendix, Section S2.3*), was calibrated as a free model parameter based on the observed geochemical profiles. Such a high NO_2^- leakage may result from an optimization of energy yield under electron donor limitation. Additional experimental work is needed to determine the mechanisms controlling this leakage by SUP05. Heterotrophic denitrification and nitrification are conventionally thought of as the primary sources of both nitrite and ammonium for anammox in OMZs (20, 29), and so far evidence for a direct coupling between sulfide-driven denitrification and anammox has been scarce (28). Our results indicate that incomplete sulfide-driven denitrification can be an important precursor for anammox, particularly under conditions of organic carbon limitation (30). This coupling and the benthic supply of ammonium lead to a substantial departure of the fraction of total nitrogen

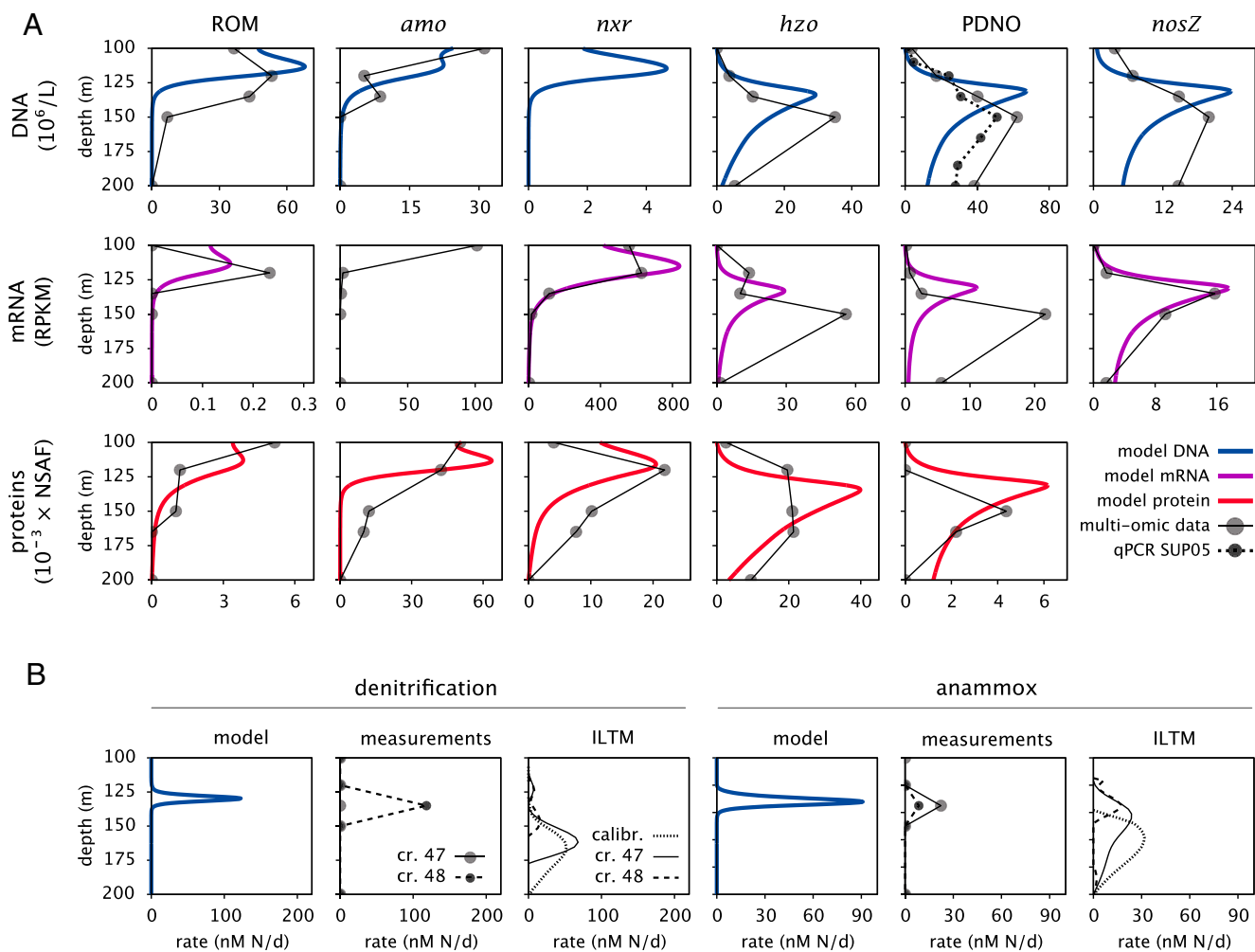


Fig. 3. Molecular and rate profiles. (A) Predicted DNA, mRNA, and protein concentrations for ROM, *amo*, *nxr*, *norBC*, *hzo*, and *nosZ* genes (thick curves) compared with corresponding metagenomic, metatranscriptomic, and metaproteomic data (circles; February 10, 2010). The dashed curve under PDNO genes (row 1, column 4) shows concurrent qPCR-based abundance estimates for SUP05, the dominant denitrifier in Saanich Inlet. (B) Denitrification and anammox rates predicted by the model (thick blue curves) compared with rate measurements (circles) during cruises (cr.) 47 (SI047_07/07/10) and cr. 48 (SI048_08/11/10) as well as rates estimated from geochemical concentration profiles using ILTM (*SI Appendix, Section S5*). The ILTM estimates calibration (calibr.) in columns 3 and 6 is based on the same geochemical data as used for model calibration (Fig. 2).

loss via anammox in Saanich Inlet ($\sim 50\%$; predicted at steady state) from previous predictions based on labile organic matter stoichiometry ($\sim 28\%$) (31).

Steady-state gene production rates for chemoautotrophic pathways are predicted to peak around the SNTZ, reaching $\sim 3.2 \times 10^6$ genes $\cdot L^{-1} \cdot d^{-1}$. This gene production rate corresponds to a dark carbon assimilation (DCA) rate of ~ 60 nmol $- C \cdot L^{-1} \cdot d^{-1}$, assuming a carbon:dry weight ratio of 0.45 (32) and a dry cell mass of $m = 5 \times 10^{-13}$ g (22). Previously measured peak DCA rates in the Saanich Inlet OMZ reached $2 \mu\text{mol} - C \cdot L^{-1} \cdot d^{-1}$ (16), which is significantly higher than the values predicted here. The potential activity of pathways not considered here, such as heterotrophic or mixotrophic inorganic carbon assimilation (33), may explain some of these differences. Moreover, the model assumes steady-state conditions, whereas redox conditions were far from steady state during previous DCA measurements (16). Transient dynamics in Saanich Inlet can exhibit significantly higher nitrogen fluxes and chemoautotrophic activity (34), which might further explain discrepancies between the model and measured DCA rates. Accounting for chemoautotrophic productivity based on oxidant and reductant

supply in redox transition zones is generally difficult because of limited knowledge on active pathways, the possibility of cryptic nutrient cycling, and potential lateral substrate intrusions, and discrepancies similar to our study are frequently reported for other OMZs (35–38). Hence, fully accounting for measured DCA rates remains an unresolved problem with important implications for carbon cycling in OMZs.

Previously detected amino acid motifs similar to those found in proteins catalyzing DNRA suggested that SUP05 may also be providing NH_4^+ to anammox through DNRA (14). DNRA, not included in the model, is known to fuel anammox in anoxic sediments and water columns (20, 39). So far, incubation experiments have not revealed any DNRA activity in Saanich Inlet, and measured ammonium profiles do not indicate a significant ammonium source at or below the SNTZ (Fig. 2B). Nevertheless, DNRA could be active in Saanich Inlet and remain undetected because of rapid ammonium consumption by anammox (39). An extension of the model that included DNRA as an additional pathway, which we calibrated to the same geochemical data (January to March 2010), predicted negligible DNRA rates compared with denitrification and anammox and consistently converged

to the same predictions as the simpler model. These results suggest that DNRA may be absent from the Saanich Inlet water column—at least near steady-state conditions in late spring—and that the hydroxylamine-oxidoreductase homolog encoded by SUP05 plays an alternative role in energy metabolism (17, 24).

DNA concentration profiles of anammox and denitrification genes appear more diffuse and are skewed toward deep basin waters compared with their corresponding rate profiles and the SNTZ (Fig. 3). The model explains this apparent discrepancy based on turbulent diffusion and sinking, which transport genes away from their replication origin. Hence, community composition at any depth is the combined result of local as well as nonlocal population dynamics. Metabolic flexibility encoded in the genomes of microorganisms mediating these processes may also contribute to broader distributions of individual genes than their predicted activity range (30). This disconnect between local metabolic potential and activity needs to be considered when interpreting metagenomic profiles in a functional context, especially in environments with strong redox gradients, such as OMZs (9) or hydrothermal vents (6).

The concentration maxima of anammox and denitrification genes are predicted at shallower depths than measured (Fig. 3A). This observation is consistent with the upward offset of the predicted SNTZ and highlights an important limitation of steady-state models applied to dynamic ecosystems. Indeed, process rate maxima predicted via ILTM at multiple time points continue to move upward beyond the time interval used for model calibration (Fig. 3B). In reality, an electron donor/electron acceptor interface as narrow as predicted by the model would only develop after sufficient time for transport processes and microbial activity to reach a true steady state. Such narrow interfaces do appear in permanently stratified meromictic lakes (40) or the Baltic Sea, where stagnation periods can persist for many years (30).

mRNA and Protein Profiles

Metatranscriptomics and metaproteomics present powerful means to assess community metabolic activity—rather than just metabolic potential—and each method comes with its own set of advantages (14, 41). For example, although transcripts represent immediate proxies for gene up-regulation (e.g., in response to changing redox conditions), proteins reflect the immediate catalytic potential of a community, and *in vitro* characterization of enzyme kinetics can facilitate the projection of protein abundances to *in situ* process kinetics (42). Transcript abundances need not always correlate strongly with protein abundances (for example, in cases of translational control or protein instability) (42, 43), and hence, metatranscriptomics and metaproteomics provide different perspectives on community activity. Hence, a systematic evaluation of the consistency between these alternative layers of information in real ecosystems is warranted. In fact, a unifying mechanistic model describing the processes that control environmental mRNA and protein distributions is crucial for the correct interpretation of multiomic data in relation to biogeochemical processes (41).

Although DNA replication and process rates are predicted by our gene-centric model merely based on environmental redox conditions, it is uncertain to what extent intermediate stages of gene expression (transcription and translation) can be explained based on such a paradigm. For example, environmental mRNA concentrations measured via qPCR have previously been directly compared with predicted reaction rates (5), but such a heuristic comparison ignores other mechanisms controlling environmental biomolecule distributions, such as physical transport processes. Here, in an attempt to mechanistically describe mRNA and protein dynamics at ecosystem scales, we hypothesized that both mRNA and protein production rates at a particular depth are proportional to the total reaction rate at that depth (calculated using the calibrated model). This premise is motivated by

observations of elevated transcription and translation rates during high metabolic activity or growth (44–46). Furthermore, we assumed that mRNA and protein molecules are subject to the same hydrodynamic dispersal processes as DNA, while decaying exponentially with time postsynthesis. The decay time of each molecule as well as the proportionality factor between the reaction rate and synthesis were estimated statistically using metatranscriptomic and metaproteomic data (*SI Appendix, Section S2.10*).

The general agreement between this model and the molecular data (Fig. 3A) suggests that the production–degradation dynamics of several of these molecules is, at the ecosystem level, dominated by the mechanisms described above. The best fit (in terms of the coefficient of determination) (*SI Appendix, Table S5*) is achieved for *nosZ* and *nrx* mRNA as well as *amo*, *norBC*, *ROM*, and *nrx* proteins. The greater number of protein over mRNA profiles that can be explained by the model suggests that the proteins considered here are, indeed, simply produced on demand and slowly degrade over time, whereas mRNA dynamics are subject to more complex regulatory mechanisms (41, 47). In particular, the decay times of some transcripts and proteins were estimated to be as high as several weeks (*SI Appendix, Table S5*). For proteins, these estimates fall within known ranges (47); however, for transcripts, these estimates are much higher than decay times determined experimentally in cells (48). One reason for this discrepancy seems to be the underestimation of the SNTZ depth range by the model, which in turn leads to longer estimates for mRNA decay times needed to explain the detection of these molecules outside of the SNTZ. Alternatively, transcripts and proteins might persist in the cells in inactive states for a significant period, even after dispersal into areas with low substrate concentrations. For example, stable but silent transcripts have been found in bacteria after several days of starvation (49, 50). Furthermore, gene expression may not change immediately in response to external stimulus (14). For example, for some prokaryotic transcription cascades, the basic time unit may be the cell doubling time (which can reach several weeks in anoxic environments) (51) because of regulation by long-lived transcription factors (52). Hence, the decay times estimated here may reflect a hysteresis in gene down-regulation after nutrient depletion, perhaps in anticipation of potential future opportunities for growth (53, 54). Overall, these observations suggest that future metatranscriptomic sequencing efforts and models for environmental mRNA dynamics would benefit from a consideration of additional control mechanisms (for example, derived from cell-centric transcription models) (55, 56).

Consequences for Geobiology

Gene-centric models have the potential to integrate biogeochemical processes with microbial population dynamics (5, 6). According to the central dogma of molecular biology (4), gene transcription and translation are intermediate steps that regulate metabolic processes in individual cells, but the appropriate projection of the central dogma to ecosystem scales remained unresolved, because transcription and translation were not explicitly considered in previous models (5, 6). We have developed a biogeochemical model that explicitly incorporates multiomic sequence information and predicts pathway expression and growth in relation to geochemical conditions. In particular, when mRNA and protein dynamics are omitted, the gene-centric model only includes four calibrated parameters and yet is able to largely reproduce geochemical profiles (Fig. 2), relative metagenomic profiles (Fig. 3A), and SUP05 cell abundances, indicating that the good agreement between the model and the data is unlikely caused by overfitting. In fact, as we refined and calibrated our model to the geochemical profiles, we observed that the metagenomic profiles were well-reproduced as soon as the model's geochemical predictions roughly aligned with the data, even if the parameters being calibrated had not converged

yet. This observation reinforces the interpretation that fluxes of matter and energy are robust predictors of microbial productivity and functional community structure.

Our model successfully explains a large fraction of environmental mRNA and protein distributions based on DNA concentration profiles and biochemical reaction rates in the OMZ. These results are consistent with the idea that DNA is a robust descriptor of an ecosystem's biotic component (57, 58), which in conjunction with the geochemical background, determines pathway expression and process rates (59). This idea implies that the co-occurrence of a metabolic niche with cells able to exploit it is sufficient to predict microbial activity. Under this paradigm, DNA may be regarded as directly interacting with the geochemical background, whereas the production of mRNA and proteins is an inevitable consequence of the biologically catalyzed flow of matter and energy. This interpretation is supported by the overall consistency between the metatranscriptomic and metaproteomic profiles for N and S cycling genes (Fig. 3A). Hence, mRNA and proteins may each be adequate proxies for pathway activity in Saanich Inlet, although as discussed above, questions regarding the relevant time scales remain. Additional work is needed to test this paradigm in other ecosystems, especially under non-steady-state conditions. Nevertheless, many aquatic ecosystems are permanently or almost permanently anoxic (8, 40, 60), and hence, our approach and conclusions are expected to be particularly applicable to these systems.

In addition to providing a systematic calibration and validation of the model, we identified processes that need to be considered when interpreting multiomic data. Conventional proxies for activity, such as mRNA and proteins, are themselves subject to complex population dynamics that include production and active or passive degradation as well as physical transport processes. Consequently, the close association between process rates and biomolecule production suggested above does not imply that biomolecule distributions, per se, are equivalent to local microbial activity rates. In Saanich Inlet, for example, the wide distribution of DNA, mRNA, and proteins across the OMZ, in contrast to a relatively narrow metabolic “hot zone” at the SNTZ, is predicated on a balance between spatially confined production and dispersal across the water column. This “mass effect” (61) means that geochemical or biochemical information is needed to assign actual activity to genes or pathways identified in multiomic data, especially for components mediating energy metabolism. This conclusion is generalizable and should be applied to other ecosystems exhibiting dispersal across strong environmental gradients, such as estuaries (62) or hydrothermal vents (6). Moreover, in dynamic ecosystems with rapidly changing geochemical conditions, past population growth rates can influence future community structure and biomolecular patterns, and hence cross-sectional community profiles may not reflect current dynamics (63). In such systems, an incorporation of multiple layers of geochemical and biological information into a mechanistic model—as shown here—will be crucial for disentangling the multitude of processes underlying experimental observations.

The gene-centric model constructed here, although evaluated at steady state, is in fact a spatiotemporal model that could, in principle, predict gene population dynamics and process rates over time. A spatiotemporal analysis of the model would require multiomic time series coverage and knowledge of nonstationary physical processes, such as seasonal patterns in surface productivity and hydrodynamics during deep water renewal events. Multiomic time series would be especially useful for improving the mRNA and protein models introduced here because of the high number of currently unknown parameters. For example, integrating metatranscriptomic, metaproteomic, and geochemical time series during rapid environmental changes into our model would allow for a more direct determination of in situ

transcriptional and translational responses and biomolecule decay times.

The multiomic profiles that we used to validate our model are only given in terms of relative—rather than absolute—biomolecule concentrations. Hence, the observed abundance of each biomolecule may be influenced by the abundances of other biomolecules, which could explain some of the discrepancies between the model and the multiomic data. Unfortunately, this limitation is currently pervasive across environmental shotgun sequencing studies, largely because of technical challenges in estimating in situ DNA, mRNA, and protein concentrations (64). These challenges will likely be overcome in the future (65, 66). Given this current caveat, the general agreement of the model with the shape of the multiomic profiles (Fig. 3A) is remarkable and suggests that the spatial structuring of the metabolic network is well-captured by the model. In fact, our qPCR-based estimates for absolute SUP05 abundances are consistent with absolute PDNO gene concentrations predicted by the model as well as the shape of the PDNO abundance profiles in the metagenomes (Fig. 3A). This double agreement suggests that—at least for PDNO—both our model and our metagenomic datasets (Datasets S1 and S2) reflect the actual gene distributions.

Conclusions

Most major metabolic pathways driving global biogeochemical cycles are encoded by a core set of genes, many of which are distributed across distant microbial clades (1). These genes are expressed and proliferate in response to specific redox conditions and in turn, shape Earth's surface chemistry. Here, we have shown that the population dynamics of genes representative of specific metabolic pathways, their expression, and their catalytic activity at ecosystem scales can all be integrated into a mechanistic framework for understanding coupled carbon, nitrogen, and sulfur cycling in OMZs. This framework largely explained DNA, mRNA, and protein concentration profiles and resolved several previous uncertainties in metabolic network structure in Saanich Inlet, including a direct coupling of sulfide-driven denitrification and anammox through leaky nitrite production by SUP05 as well as the presence of a metabolic niche for nitrous oxide reduction contributing to nitrogen loss. Beyond OMZ-specific predictions, model results indicate that geochemical fluxes are robust indicators of microbial community structure and reciprocally, that gene abundances and geochemical conditions largely determine gene expression patterns. Such integrated modeling approaches offer insight into microbial community metabolic networks and allow prediction of elemental cycling in a changing world.

Materials and Methods

Model Overview. The core model is a set of differential equations for the concentrations of eight metabolites and six proxy genes (DNA) across depth (100–200 m) and time. Each gene is a proxy for a particular energy-yielding pathway, which couples the oxidation of an external electron donor to the reduction of an external electron acceptor. Each gene is considered as a replicating unit that is independent of other genes. This simplifying assumption corresponds to the case where each cell occupies a single metabolic niche associated with one of the modeled pathways (14, 67). Gene-specific reaction rates depend on the concentrations of metabolites according to first- or second-order (Michaelis–Menten) kinetics (5, 21) (SI Appendix, Section S2.4). In turn, the production or depletion of metabolites at any depth is determined by the reaction rates at that depth, taking into account reaction stoichiometry (SI Appendix, Section S2.3) and diffusive transport across the water column. The production of genes at any depth is driven by the release of energy from their catalyzed reactions and is proportional to the Gibbs free energy multiplied by the reaction rate (22) (SI Appendix, Section S2.5). In addition, gene populations are subject to exponential decay rates, diffusive transport, and sinking.

Mathematical Model Structure. The DNA concentration for gene r (Γ_r ; copies per volume) exhibits the dynamics

$$\frac{\partial \Gamma_r}{\partial t} = -q_r \Gamma_r + \frac{1}{c} Z_r H_r \Gamma_r - v \frac{\partial \Gamma_r}{\partial z} + \frac{\partial}{\partial z} \left(K \frac{\partial \Gamma_r}{\partial z} \right), \quad [1]$$

whereas the concentration of the m th metabolite (C_m ; mole per volume) follows

$$\frac{\partial C_m}{\partial t} = \sum_r S_{mr} H_r \Gamma_r + \frac{\partial}{\partial z} \left(K \frac{\partial C_m}{\partial z} \right). \quad [2]$$

Both the gene concentrations Γ_r and metabolite concentrations C_m depend on time t and depth z . The first term on the right-hand side of Eq. 1 corresponds to cell death, with q_r being the exponential death rate in the absence of any metabolic activity for pathway r . The second term corresponds to gene production, with H_r being the per-gene reaction rate as a function of metabolite concentrations (SI Appendix, Section S2.4). The biomass production coefficient Z_r is a linear function of the Gibbs free energy of reaction r (SI Appendix, Section S2.5); c is the average dry cell mass, which is used to convert biomass production into cell production. The third term corresponds to cell sinking at speed v . The last term in Eqs. 1. and 2 corresponds to diffusive transport, with K being the vertical eddy diffusion coefficient. In Eq. 2, S_{mr} is the stoichiometric coefficient of metabolite m in reaction r (SI Appendix, Section S2.3). The sum on the right-hand side of Eq. 2 iterates through all reactions and accounts for microbial metabolic fluxes. Eqs. 1. and 2 specify the rates of change for the DNA and chemical concentration profiles. Steady-state profiles were obtained after long simulations when all profiles had eventually stabilized.

Considered Pathways. Redox pathways occurring in a single cell require at least two enzymes: one involved in the oxidation of the initial electron donor and one involved in the reduction of the final electron acceptor. In the model, such pathways are represented by single proxy genes, chosen such that ambiguities in their functional role are minimized. For example, nitrous oxide reduction using nosZ coupled to sulfide oxidation is identified with nosZ, because many sulfur oxidizing enzymes are reversible. Other pathways considered in the model are partial denitrification of nitrate to nitrous oxide coupled to sulfide oxidation (PDNO), aerobic ammonium oxidation using amo, aerobic nitrite oxidation to nitrate using nxr, and anammox (i.e., the anammox involving hzo) as well as aerobic remineralization of (dissolved) organic matter (ROM). PDNO comprises three denitrification steps that are thought to be predominantly performed by the same microorganisms in the SUP05 clade (14, 17): dissimilatory nitrate reduction to nitrite by membrane-bound respiratory nitrate reductases (narGHIJ) or periplasmic dissimilatory nitrate reductases (napAB), nitrite reduction to nitric oxide using nitrite reductases (nirKS), and nitric oxide reduction to nitrous oxide using norBC. The first denitrification step was assumed to be leaky, so that a small fraction of nitrite is released into the extracellular environment (20). We used norBC as a proxy for PDNO when interpreting molecular data (Fig. 3A shows coverage of the dissimilatory sulfide oxidation pathway, and SI Appendix, Fig. S4 D–F shows narGHIJ, napAB, and nirKS multimolecular data). ROM is associated with the release of ammonium and sulfate (SO_4^{2-}) at ratios corresponding to marine bacterial biomass stoichiometry (32). The choice of redox pathways in the model follows the hypotheses put forward by Hawley et al. (14) based on molecular depth profiles as well as reports of nitrous oxide reduction coupled to hydrogen sulfide oxidation in Saanich Inlet (68).

Hydrogen sulfide is assumed to originate via diffusion from the sediments, where intense sulfate reduction occurs (15) (SI Appendix, Section S3.1). Sulfate reduction was omitted from our model, because both our molecular as well as chemical data suggest that sulfate reduction in the water column is negligible compared with the oxidation of sulfur compounds (a detailed discussion is in SI Appendix, Section S3.1). In fact, when we included sulfate reduction in preliminary tests of our model, the agreement between the model and the H_2S profiles decreased dramatically, providing additional evidence that H_2S is largely supplied from the bottom rather than produced in the water column.

Aerobic H_2S oxidation was omitted from the model based on extensive previous work that points toward NO_3^- and other nitrogen compounds as dominant electron acceptors for H_2S oxidation during periods of strong stratification (18, 16, 12–14). For example, as shown in Fig. 2B, the upper boundary of H_2S concentrations closely follows the lower boundary of NO_3^- —rather than O_2 —over time, especially during the period considered here (early 2010). We mention that, during renewal events in the fall, O_2 can become an important electron acceptor for H_2S oxidation (12); however, this

possibility does not affect this study, which focuses on a period of intense stratification near steady-state conditions. A more detailed discussion on the role of aerobic sulfide oxidation is provided in SI Appendix, Section S3.3.

Pathways for hydrogen (H_2) and methane (CH_4) metabolism are not included on grounds of parsimony, because these pathways are not directly linked to the other considered pathways (12) and because low hydrogen and methane fluxes into the OMZ suggest that hydrogen and methane pathways are of secondary importance (12, 69).

Model Calibration and Data. Unknown parameters of the basic gene-centric model (Eqs. 1. and 2) (ignoring mRNA and protein dynamics) were calibrated by comparing steady-state predictions with measured depth profiles of oxygen, ammonium, nitrate, nitrite, hydrogen sulfide, and nitrous oxide. Chemical calibration data were acquired on January 13, February 10, and March 10, 2010 (or February 10 and April 7 for oxygen) from a single location in Saanich Inlet (123° 30.30' W, 48° 35.50' N) (SI Appendix, Section S1.2). The calibrated parameters were the maximum cell-specific reaction rate V_{PDNO} and the first-order rate constants A_{ROM} and A_{nosZ} as well as the PDNO leakage fraction L_{PDNO} (SI Appendix, Table S3). Calibration was performed by maximizing the likelihood of a statistical model, in which the deterministic part (i.e., expectation) is given by the predictions of the gene-centric model and the stochastic part (i.e., error) is normally distributed (SI Appendix, Section S2.8). This calibration method is known as maximum likelihood estimation and is widespread in statistical regression and physics (70). Maximization of the likelihood was performed using the MATLAB function `fmincon`, which uses repeated simulations and gradual exploration of parameter space (71). The sensitivity of the model to parameter variation was assessed via local sensitivity analysis (72) as described in SI Appendix, Section S2.12. An overview of our workflow is shown in SI Appendix, Fig. S2.

Samples for molecular sequencing were collected on February 10, 2010 from the same geographical location as the geochemical data (SI Appendix, Sections S1.3 and S1.4). Because metaproteomes were missing at depth 100 m (the upper bound of our simulation domain) and to increase statistical power when evaluating our protein models, we used linear interpolation between depths 97 and 120 m (where metaproteomes were available) to estimate protein normalized abundance factor (NSAF) values at 100-m depth (“unit imputation”). Metagenomic profiles (a priori in relative units) were rescaled to match the model scales using maximum likelihood estimated factors (SI Appendix, Section S2.9). SUP05 abundances for February 10, 2010 were estimated via qPCR using SUP05-specific primers targeting the 519–1,048 region of the SUP05 16S rRNA gene following the protocol in ref. 73; 16S gene counts were corrected for the number of 16S rRNA gene copies per cell estimated using the Tax4Fun pipeline (74) (SI Appendix, Section S1.6). Denitrification and anammox rates were measured on cruises 47 (SI047_07/07/10) and 48 (SI048_08/11/10) via ex situ incubation experiments and subsequently corrected for differences between in situ and incubated substrate concentrations (SI Appendix, Section S1.5).

mRNA and Protein Models. As mentioned previously, following calibration of the gene-centric model to the geochemical profiles, we extended the model to describe mRNA (and similarly, protein) dynamics in the water column. Specifically, the production rate of an mRNA (transcripts produced per time and volume of seawater) at a particular depth was assumed to be proportional to the total reaction rate (moles per time and volume of seawater) at that depth. A linear relation, although only an approximation, can be justified by the fact that increased enzyme dilution rates at elevated cell division rates must be balanced (at the population level) by correspondingly increased translation—and hence, transcription—rates (55). We also assumed that mRNA molecules disperse via diffusion and sinking similar to genes (because they are hosted by the same cells) and decay exponentially with time. Thus, environmental mRNA concentrations satisfy the partial differential equation

$$\frac{\partial T_r}{\partial t} = -\frac{T_r}{\tau_r} + \frac{R_r}{\alpha_r} - v \frac{\partial T_r}{\partial z} + \frac{\partial}{\partial z} \left(K \frac{\partial T_r}{\partial z} \right), \quad [3]$$

where T_r is the mRNA concentration corresponding to the r th reaction, τ_r is the decay time of the mRNA molecule, $R_r = H_r \Gamma_r$ is the total reaction rate, and α_r is an unknown proportionality constant. We considered T_r in the same units as the multiomic data, i.e., reads per kilobase per million mapped reads (RPKM) for metatranscriptomes and NSAF for metaproteomes. Consequently, α_r is the ratio between the r th reaction rate (moles · day⁻¹ · liter⁻¹) and the corresponding RPKM (or NSAF) “production rate” (RPKM · day⁻¹), and thus it depends on not only the particular reaction but also on our sampling protocol and sequencing pipeline. The above model was evaluated at steady state, when mRNA production, dispersal, and decay are balanced at each depth ($\partial T_r / \partial t = 0$). The parameters of the mRNA and protein models

(proportionality factors and decay times) were calibrated by fitting to the metatranscriptomic and metaproteomic data, respectively (*SI Appendix, Section S2.10*). Calibration to metatranscriptomic data failed for *amo* mRNA. Metagenomic and metaproteomic data were not available for *nrx* and *nosZ*, respectively (*SI Appendix, Section S1.3*). For all other mRNAs and proteins, the iterative calibration converged rapidly to an optimum, and this optimum was robust against various starting values for the parameters.

ILTM. In addition to the model predictions and rate measurements, denitrification and anammox rates were also estimated directly from chemical concentration profiles via ILTM (*SI Appendix, Section S5*). ILTM provides an estimate for the metabolic fluxes in the OMZ based on the observed chemical concentration profiles. The exact shape of estimated rate profiles depends sensitively on measurement errors and the noise reduction method applied to the concentration profiles. Hence, ILTM only serves as a rough verification of the order of magnitude of rates predicted by the model or measured experimentally. ILTM fitting was applied separately to concentration profiles from cruises 47 and 48 as well as the chemical profiles used for model calibration (cruises 41–44) (Fig. 2) after averaging across replicates at each depth.

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- Falkowski PG, Fenchel T, Delong EF (2008) The microbial engines that drive Earth's biogeochemical cycles. *Science* 320(5879):1034–1039.
- Ridgwell A, et al. (2007) Marine geochemical data assimilation in an efficient earth system model of global biogeochemical cycling. *Biogeochemistry* 4(1):87–104.
- Galbraith E, Gnanadesikan A, Dunne J, Hiscock M (2010) Regional impacts of iron-light colimitation in a global biogeochemical model. *Biogeochemistry* 7(3):1043–1064.
- Crick F (1970) Central dogma of molecular biology. *Nature* 227(5258):561–563.
- Reed DC, Algar CK, Huber JA, Dick GJ (2014) Gene-centric approach to integrating environmental genomics and biogeochemical models. *Proc Natl Acad Sci USA* 111(5):1879–1884.
- Reed DC, et al. (2015) Predicting the response of the deep-ocean microbiome to geochemical perturbations by hydrothermal vents. *ISME J* 9(8):1857–1869.
- Anderson JJ, Devol AH (1973) Deep water renewal in Saanich Inlet, an intermittently anoxic basin. *Estuar Coast Mar Sci* 1(1):1–10.
- Ulloa O, Canfield DE, DeLong EF, Letelier RM, Stewart FJ (2012) Microbial oceanography of anoxic oxygen minimum zones. *Proc Natl Acad Sci USA* 109(40):15996–16003.
- Wright JJ, Konwar KM, Hallam SJ (2012) Microbial ecology of expanding oxygen minimum zones. *Nat Rev Microbiol* 10(6):381–394.
- Ward BB, et al. (2009) Denitrification as the dominant nitrogen loss process in the Arabian Sea. *Nature* 461(7260):78–81.
- Lam P, Kuypers MM (2011) Microbial nitrogen cycling processes in oxygen minimum zones. *Annu Rev Mar Sci* 3:317–345.
- Zaikova E, et al. (2010) Microbial community dynamics in a seasonally anoxic fjord: Saanich Inlet, British Columbia. *Environ Microbiol* 12(1):172–191.
- Walsh DA, Hallam SJ (2011) *Handbook of Molecular Microbial Ecology II: Metagenomics in Different Habitats*, ed de Bruijn FJ (Wiley, Hoboken, NJ), pp 253–267.
- Hawley AK, Brewer HM, Norbeck AD, Paša-Tolić L, Hallam SJ (2014) Metaproteomics reveals differential modes of metabolic coupling among ubiquitous oxygen minimum zone microbes. *Proc Natl Acad Sci USA* 111(31):11395–11400.
- Murray JW, Grundmanis V, Smethie WM, Jr (1978) Interstitial water chemistry in the sediments of Saanich Inlet. *Geochim Cosmochim Acta* 42(7):1011–1026.
- Juniper S, Brinkhurst R (1986) Water-column dark CO₂ fixation and bacterial-mat growth in intermittently anoxic Saanich Inlet, British-Columbia. *Mar Ecol Prog Ser* 33(1):41–50.
- Walsh DA, et al. (2009) Metagenome of a versatile chemolithoautotroph from expanding oceanic dead zones. *Science* 326(5952):578–582.
- Anderson J (1984) The oxic/anoxic interface in Saanich Inlet. *Canadian Technical Report of Hydrography and Ocean Sciences*, eds Juniper S, Brinkhurst R (Department of Fisheries and Oceans Canada, Sidney, BC, Canada), No. 83, pp. 17–23.
- Schunck H, et al. (2013) Giant hydrogen sulfide plume in the oxygen minimum zone off Peru supports chemolithoautotrophy. *PLoS One* 8(8):e68661.
- Lam P, et al. (2009) Revising the nitrogen cycle in the Peruvian oxygen minimum zone. *Proc Natl Acad Sci USA* 106(12):4752–4757.
- Jin Q, Roden EE, Giska JR (2013) Geomicrobial kinetics: Extrapolating laboratory studies to natural environments. *Geomicrobiol J* 30(2):173–185.
- Roden EE, Jin Q (2011) Thermodynamics of microbial growth coupled to metabolism of glucose, ethanol, short-chain organic acids, and hydrogen. *Appl Environ Microbiol* 77(5):1907–1909.
- Shah V, Morris RM (2015) Genome sequence of “Candidatus Thioglobus autotrophica” strain EF1, a chemoautotroph from the SUP05 clade of marine gammaproteobacteria. *Genome Announc* 3(5):e01156-15.
- Shah V, Chang BX, Morris RM (July 19, 2016) Cultivation of a chemoautotroph from the SUP05 clade of marine bacteria that produces nitrite and consumes ammonium. *ISME J*, 10.1038/ismej.2016.87.
- Brettar I, Rheinheimer G (1991) Denitrification in the Central Baltic: Evidence for H₂S-oxidation as motor of denitrification at the oxic-anoxic interface. *Mar Ecol Prog Ser* 77(2–3):157–169.
- Zubkov M, Sazhin A, Flint M (1992) The microplankton organisms at the oxic-anoxic interface in the pelagial of the Black Sea. *FEMS Microbiol Lett* 101(4):245–250.
- Grote J, Jost G, Labrenz M, Herndl GJ, Jürgens K (2008) Epsilonproteobacteria represent the major portion of chemoautotrophic bacteria in sulfidic waters of pelagic redoxclines of the Baltic and Black Seas. *Appl Environ Microbiol* 74(24):7546–7551.
- Wenk CB, et al. (2013) Anaerobic ammonium oxidation (anammox) bacteria and sulfide-dependent denitrifiers coexist in the water column of a meromictic south-alpine lake. *Limnol Oceanogr* 58(1):1–12.
- Voss M, Montoya JP (2009) Nitrogen cycle: Oceans apart. *Nature* 461(7260):49–50.
- Hannig M, et al. (2007) Shift from denitrification to anammox after inflow events in the central Baltic Sea. *Limnol Oceanogr* 52(4):1336–1345.
- Babbitt AR, Keil RG, Devol AH, Ward BB (2014) Organic matter stoichiometry, flux, and oxygen control nitrogen loss in the ocean. *Science* 344(6182):406–408.
- Fagerbakke K, Heldal M, Norland S (1996) Content of carbon, nitrogen, oxygen, sulfur and phosphorus in native aquatic and cultured bacteria. *Aquat Microb Ecol* 10(1):15–27.
- DeLorenzo S, et al. (2012) Ubiquitous dissolved inorganic carbon assimilation by marine bacteria in the Pacific Northwest coastal ocean as determined by stable isotope probing. *PLoS One* 7(10):e46695.
- Manning CC, Hamme RC, Bourbonnais A (2010) Impact of deep-water renewal events on fixed nitrogen loss from seasonally-anoxic Saanich Inlet. *Mar Chem* 122(1–4):1–10.
- Murray JW, Jorgensen BB, Fossing H, Wirsen CO, Jannasch HW (1991) Sulfide oxidation in the anoxic black sea chemocline. *Deep Sea Res A* 38:51083–51103.
- Zopfi J, Ferdelman TG, Jorgensen BB, Teske A, Thamdrup B (2001) Influence of water column dynamics on sulfide oxidation and other major biogeochemical processes in the chemocline of Mariager Fjord (Denmark). *Mar Chem* 74(1):29–51.
- Jost G, Zubkov MV, Yakushev E, Labrenz M, Jurgens K (2008) High abundance and dark CO₂ fixation of chemolithoautotrophic prokaryotes in anoxic waters of the Baltic Sea. *Limnol Oceanogr* 53(1):14–22.
- Li XN, Taylor GT, Astor Y, Varela R, Scranton MI (2012) The conundrum between chemoautotrophic production and reductant and oxidant supply: A case study from the Cariaco basin. *Deep Sea Res Part 1 Oceanogr Res Pap* 61:1–10.
- Prokopenko MG, et al. (2013) Nitrogen losses in anoxic marine sediments driven by Thioploca-anammox bacterial consortia. *Nature* 500(7461):194–198.
- Sorokin DY, Foti M, Pinkart HC, Muyzer G (2007) Sulfur-oxidizing bacteria in Soap Lake (Washington State), a meromictic, haloalkaline lake with an unprecedented high sulfide content. *Appl Environ Microbiol* 73(2):451–455.
- Moran MA, et al. (2013) Sizing up metatranscriptomics. *ISME J* 7(2):237–243.
- Wallenstein MD, Weintraub MN (2008) Emerging tools for measuring and modeling the in situ activity of soil extracellular enzymes. *Soil Biol Biochem* 40(9):2098–2106.
- Lodish HF (1976) Translational control of protein synthesis. *Annu Rev Biochem* 45:39–72.
- Amy PS, Pauling C, Morita RY (1983) Recovery from nutrient starvation by a marine *Vibrio* sp. *Appl Environ Microbiol* 45(5):1685–1690.
- Kemp PF, Lee S, Laroche J (1993) Estimating the growth rate of slowly growing marine bacteria from RNA content. *Appl Environ Microbiol* 59(8):2594–2601.
- Golding I, Paulsson J, Zawilski SM, Cox EC (2005) Real-time kinetics of gene activity in individual bacteria. *Cell* 123(6):1025–1036.
- Goldberg AL, St John AC (1976) Intracellular protein degradation in mammalian and bacterial cells: Part 2. *Annu Rev Biochem* 45:747–803.
- Takayama K, Kjelleberg S (2000) The role of RNA stability during bacterial stress responses and starvation. *Environ Microbiol* 2(4):355–365.
- Marouga R, Kjelleberg S (1996) Synthesis of immediate upshift (Iup) proteins during recovery of marine *Vibrio* sp. strain S14 subjected to long-term carbon starvation. *J Bacteriol* 178(3):817–822.
- Kolter R, Siegel DA, Tormo A (1993) The stationary phase of the bacterial life cycle. *Annu Rev Microbiol* 47:855–874.

51. Whitman WB, Coleman DC, Wiebe WJ (1998) Prokaryotes: The unseen majority. *Proc Natl Acad Sci USA* 95(12):6578–6583.
52. Rosenfeld N, Alon U (2003) Response delays and the structure of transcription networks. *J Mol Biol* 329(4):645–654.
53. Mitchell A, et al. (2009) Adaptive prediction of environmental changes by microorganisms. *Nature* 460(7252):220–224.
54. Lawson CE, et al. (2015) Rare taxa have potential to make metabolic contributions in enhanced biological phosphorus removal ecosystems. *Environ Microbiol* 17(12):4979–4993.
55. Rosenfeld N, Elowitz MB, Alon U (2002) Negative autoregulation speeds the response times of transcription networks. *J Mol Biol* 323(5):785–793.
56. Zeisel A, et al. (2011) Coupled pre-mRNA and mRNA dynamics unveil operational strategies underlying transcriptional responses to stimuli. *Mol Syst Biol* 7:529.
57. Inskeep WP, et al. (2010) Metagenomes from high-temperature chemotrophic systems reveal geochemical controls on microbial community structure and function. *PLoS One* 5(3):e9773.
58. Dick JM, Shock EL (2013) A metastable equilibrium model for the relative abundances of microbial phyla in a hot spring. *PLoS One* 8(9):e72395.
59. Larsen PE, et al. (2011) Predicted Relative Metabolomic Turnover (PRMT): Determining metabolic turnover from a coastal marine metagenomic dataset. *Microb Inform Exp* 1(1):4.
60. Scranton MI, et al. (2014) Interannual and subdecadal variability in the nutrient geochemistry of the Cariaco Basin. *Oceanography (Wash D C)* 27(1):148–159.
61. Leibold MA, et al. (2004) The metacommunity concept: A framework for multi-scale community ecology. *Ecol Lett* 7(7):601–613.
62. Crump BC, Hopkinson CS, Sogin ML, Hobbie JE (2004) Microbial biogeography along an estuarine salinity gradient: Combined influences of bacterial growth and residence time. *Appl Environ Microbiol* 70(3):1494–1505.
63. Smith MB, et al. (2015) Natural bacterial communities serve as quantitative geochemical biosensors. *MBio* 6(3):e00326-15.
64. Prosser JI (2015) Dispersing misconceptions and identifying opportunities for the use of 'omics' in soil microbial ecology. *Nat Rev Microbiol* 13(7):439–446.
65. Munro SA, et al. (2014) Assessing technical performance in differential gene expression experiments with external spike-in RNA control ratio mixtures. *Nat Commun* 5:5125.
66. Smets W, et al. (2016) A method for simultaneous measurement of soil bacterial abundances and community composition via 16S rRNA gene sequencing. *Soil Biol Biochem* 96:145–151.
67. Ramos JL (2003) Lessons from the genome of a lithoautotroph: Making biomass from almost nothing. *J Bacteriol* 185(9):2690–2691.
68. Cohen Y (1978) Consumption of dissolved nitrous oxide in an anoxic basin, Saanich Inlet, British Columbia. *Nature* 272(5650):235–237.
69. Lilley MD, Baross JA, Gordon LI (1982) Dissolved hydrogen and methane in Saanich Inlet, British Columbia. *Deep Sea Res A* 29(12):1471–1484.
70. Eliason SR (1993) *Maximum Likelihood Estimation: Logic and Practice* (SAGE Publications, Newbury Park, CA).
71. The MathWorks Inc. (2010) MATLAB (R2010a) (The MathWorks Inc., Natick, MA), Version 7.10.0.
72. Frey HC, Patil SR (2002) Identification and review of sensitivity analysis methods. *Risk Anal* 22(3):553–578.
73. Hawley AK, et al. (2013) *Methods in Enzymology: Microbial Metagenomics, Metatranscriptomics, and Metaproteomics*, ed DeLong EF (Academic, San Diego), Vol 531, pp 305–329.
74. Aßhauer KP, Wemheuer B, Daniel R, Meinicke P (2015) Tax4Fun: Predicting functional profiles from metagenomic 16S rRNA data. *Bioinformatics* 31(17):2882–2884.