

# Addressing Challenges in Simulating Inter-annual Variability of Gross Primary Production

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**Key Points:**

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- We investigated the limitations of biogeochemical models in simulating inter-annual variability (IAV) of gross primary production (GPP).

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- Capturing year-to-year variability of model parameters and diurnal GPP peaks can be key to understanding IAV.

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- Variability in model performance is majorly influenced by model types, parameterization strategies, and site characteristics.

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**Abstract**

A long-standing challenge in studying the global carbon cycle has been understanding the factors controlling inter-annual variation (IAV) of carbon fluxes related to vegetation photosynthesis and respiration, and improving their representations in existing biogeochemical models. Here, we compared an optimality-based mechanistic model and a semi-empirical light use efficiency model to understand how current models can be improved to simulate IAV of gross primary production (GPP). Both models simulated hourly GPP and were parameterized for (1) each site-year, (2) each site with an additional constraint on IAV ( $Cost^{IAV}$ ), (3) each site, (4) each plant-functional type, and (5) globally. This was followed by forward runs using calibrated parameters, and model evaluations at different temporal scales across 198 eddy covariance sites. Both models performed better on hourly scale than annual scale for most sites. Specifically, the mechanistic model substantially improved when drought stress was explicitly included. Most of the variability in model performances was due to model types and parameterization strategies. The semi-empirical model produced statistically better hourly simulations than the mechanistic model, and site-year parameterization yielded better annual performance for both models. Annual model performance did not improve even when parameterized using  $Cost^{IAV}$ . Furthermore, both models underestimated the peaks of diurnal GPP in each site-year, suggesting that improving predictions of peaks could produce a comparatively better annual model performance. GPP of forests were better simulated than grassland or savanna sites by both models. Our findings reveal current model deficiencies in representing IAV of carbon fluxes and guide improvements in further model development.

**Plain Language Summary**

Terrestrial vegetation assimilates and releases carbon dioxide through photosynthesis and respiration, respectively, and their net magnitude determines if vegetation can be a sink or source of carbon. We are interested in understanding what controls the inter-annual variability (IAV) of gross primary production (GPP) which represents photosynthesis, in a given location and how their representations can be improved in models simulating GPP. Here, we considered a mechanistic model that can be applied equally well globally, and a data-driven semi-empirical model. We found both models better simulated diurnal and seasonal cycles of GPP than IAV. Such differences probably stem from model parameters, as critical ecosystem functions they represent may not be well-constrained or model structures may lack critical representations via inaccurate simulation of peak diurnal GPP and drought stress. The IAV of GPP was comparatively better simulated if model parameters were fine-tuned with data from specific years. Another challenge is that IAV of GPP can also be observed due to disturbances, such as forest fire, and human management besides natural causes, which were also not represented in models. Our results suggest that learning the variability of model parameters over the years can be key to better simulation of the IAV of GPP.

**1 Introduction**

The global carbon cycle is an important biogeochemical cycle, which affects the climate on Earth (Schimel, 2001). Terrestrial vegetation, which covers a large part of the land area, assimilates atmospheric carbon dioxide ( $CO_2$ ) through photosynthesis. Simultaneously,  $CO_2$  of similar magnitude is released into the atmosphere during terrestrial ecosystem respiration (TER). The net balance of these two fluxes determines if terrestrial vegetation acts as a sink or source of carbon (Ruehr et al., 2023). Terrestrial gross primary production (GPP) can be defined as ‘apparent’ photosynthesis, i.e., the rate at which the vegetation assimilates carbon through photosynthesis minus the loss of carbon only through photorespiration (Plummer, 2006; Wohlfahrt & Gu, 2015). GPP can be estimated directly using gas exchange measurements at the leaf and canopy scales (Jez

120 et al., 2021), and indirectly through measurements of net ecosystem exchange (NEE) us-  
121 ing the eddy covariance (EC) method at the ecosystem or landscape scale (D. D. Bal-  
122 docchi, 2003). Though the GPP estimated using the EC method represents ‘apparent’  
123 photosynthesis, its magnitude can be closer to ‘true’ photosynthesis which is the actual  
124 amount of carbon assimilated due to overestimation of daytime mitochondrial respira-  
125 tion in flux-partitioning algorithm (Reichstein et al., 2005; Wohlfahrt & Gu, 2015). Fur-  
126 thermore, a large variety of biogeochemical models have been developed to simulate and  
127 upscale carbon fluxes from local to regional or global scales to better describe the global  
128 carbon cycle (Xiao et al., 2014; Burton et al., 2023; Dannenberg et al., 2023; Nelson et  
129 al., 2024).

130 Biogeochemical models that simulate GPP can be of different types and complex-  
131 ities. On the one hand, process-based models, such as the models used in the Trends in  
132 Net Land-Atmosphere Carbon Exchange (TRENDY) project, mechanistically describe  
133 the physiological processes involved in photosynthesis or plant respiration (Sitch et al.,  
134 2015). The ability of these process-based models to capture a certain process largely de-  
135 pends on the underlying model structure and calibration of model parameters (Anav et  
136 al., 2015). Similar, but simpler than fully mechanistic approaches are the models con-  
137 structed on the concept of light use efficiency (LUE), which treat a canopy as one big  
138 leaf, but where the GPP is calculated as the product of the absorbed photosynthetically  
139 active radiation (aPAR) and LUE (Monteith, 1972). These models are semi-empirical  
140 as they combine both the simplicity of empirical models and the theoretical mechanisms  
141 that underpin process-based models (Running et al., 2000; Yuan et al., 2007; J. Chen,  
142 2021). On the other hand, empirical models are largely based on learning regression func-  
143 tions to establish a general relation between input data, such as meteorology and ecosys-  
144 tem properties, and the desired output, such as GPP. At the site level, the ability of such  
145 data-driven models (Jung et al., 2011, 2020) to accurately simulate the GPP fluxes gen-  
146 erally outperforms mechanistic approaches, but they are largely reliant on good qual-  
147 ity training data and generally lack comprehensive representations of long-term forcing  
148 functions, such as CO<sub>2</sub> fertilization effect, i.e., increased GPP with the increase in at-  
149 mospheric CO<sub>2</sub> concentration (Schimel et al., 2015).

150 Considering the methodological diversity and differences in GPP estimates, var-  
151 ious model benchmarking and model-data integration experiments have been designed  
152 to compare approaches, but also to unveil drivers of ecosystem functioning for various  
153 bioclimatic and vegetation types, across spatial and temporal scales. A long-standing  
154 challenge, and still a key area of interest, lies in understanding the factors controlling  
155 inter-annual variability (IAV) of the various carbon fluxes (D. Baldocchi et al., 2018).  
156 The challenge presents itself from the mechanistic to the more data-driven approaches  
157 and contests the dominant role of meteorology in determining the IAV of ecosystem fluxes  
158 (Richardson et al., 2007). At the local ecosystem level, Wu et al. (2012) looked at the  
159 IAV of net ecosystem fluxes by fitting the parameters of a semi-empirical model at shorter  
160 timescales to capture the seasonality, but also annual variability of model parameters.  
161 The approach allows testing the role of changes in ecosystem functioning in the IAV of  
162 carbon fluxes (Richardson et al., 2007). They concluded that climate and parametric vari-  
163 ability control IAV of ecosystem fluxes at shorter and longer timescales, respectively. Si-  
164 multaneously, Fatichi and Ivanov (2014) highlighted the role of climate when using 200  
165 years of hourly synthetic meteorological data to force an ecohydrological model to find  
166 that the random occurrence of favourable weather conditions at certain hours of the day  
167 can be a major predictor of IAV of net primary production (NPP). This statistical re-  
168 lationship was corroborated by Zscheischler et al. (2016) using actual flux data from EC  
169 sites from forested areas in North America, where the 91<sup>st</sup> percentile values of hourly  
170 GPP flux, i.e., peak GPP values, substantially contributed to the IAV of GPP flux. These  
171 studies highlight the correlation between the distribution tails and the IAV in EC fluxes.  
172 However, there is no robust pattern across sites nor do they challenge there is no vari-  
173 ability in ecosystem function.

174 More recently, a model selection study compared an ensemble of 5600 possible semi-  
175 empirical LUE model structures to find a global best model structure (Bao et al., 2022).  
176 The best LUE model is calibrated at a daily timescale per site and explains the variabil-  
177 ity of GPP fluxes across the FLUXNET EC network (Pastorello et al., 2020), consid-  
178 ering the effect of various environmental conditions on maximum LUE through partial  
179 sensitivity functions. Though the best global model performed similarly to the best model  
180 selected for each site at the daily resolution, it failed to represent the variability of an-  
181 nually aggregated GPP fluxes for 74% of sites, i.e., the Nash-Sutcliffe efficiency (NSE)  
182 of model performance (Nash & Sutcliffe, 1970) was below or equal to 0.5. This finding  
183 may be attributed to (1) the use of daily data in the study, as the model had no infor-  
184 mation on the favourable conditions that occurred in a diurnal cycle and failed to sim-  
185 ulate the diurnal GPP peaks which had a major influence on IAV (Fatichi & Ivanov, 2014;  
186 Zscheischler et al., 2016; Bao et al., 2022), (2) the assumption of invariance in ecosys-  
187 tem function, i.e., values of model parameters remain constant for all site-years in a site,  
188 and (3) the need to explicitly consider different timescales in the cost function (Desai,  
189 2010).

190 In contrast, Mengoli et al. (2022) proposed an optimality-based framework (Wang  
191 et al., 2017; Stocker et al., 2020), i.e., process-based P-model which simulates GPP mech-  
192 anistically and differentiates between instantaneous and acclimated photosynthetic re-  
193 sponses. This model demonstrated its capability in simulating half-hourly GPP dynam-  
194 ics at ten EC sites, covering four vegetation classes for limited time periods. Whereas,  
195 the performance of this modelling framework across sites representing diverse climate-  
196 vegetation features and various temporal resolutions were not evaluated. Though this  
197 modelling framework considers the effect of temperature, vapour pressure deficit (VPD),  
198 atmospheric CO<sub>2</sub> concentration, solar radiation, and the fraction of absorbed photosyn-  
199 thetically active radiation (fAPAR), it does not explicitly consider the effect of drought  
200 stress on GPP variability at sub-daily scale. Recently, Mengoli et al. (2023) proposed  
201 an improved version of this model by incorporating climatic aridity and calculating a scal-  
202 ing factor for GPP. However, in the improved model, the scaling factor could only be ap-  
203 plied to improve the simulation of daily GPP.

204 The challenge to correctly reproduce IAV is also apparent on a global scale. Anav  
205 et al. (2015) further drew attention to the disagreement in annual GPP, modelled by var-  
206 ious global GPP modelling frameworks, such as a data-driven model-tree-ensemble (Jung  
207 et al., 2011), process-based models in the TRENDY project (Sitch et al., 2015), and the  
208 CARBONES dataset (Kuppel et al., 2013) which was derived using a hybrid approach.  
209 These discrepancies highlight that site level limitations in simulating IAV propagate to  
210 larger scales where additional mechanisms play a role in the IAV of ecosystem fluxes, such  
211 as natural or anthropogenic disturbances and land-use landcover change (McGuire et  
212 al., 2001; Bultan et al., 2022).

213 As such, here we explore ecosystem-level estimations of GPP flux to systematically  
214 investigate how various factors can be linked to describing the IAV of GPP flux, such  
215 as peak values of diurnal GPP, climatic conditions, and variables represented by model  
216 parameters, which are usually hard to measure directly and can be difficult to interpret  
217 even when various modelling approaches are adopted. We tested the impact of the con-  
218 stant or time-varying parameterizations and evaluated their performances in capturing  
219 GPP variability at various temporal aggregation scales, especially at the annual scale.  
220 We also tested the hypothesis that observational constraints complement and enhance  
221 theoretically-grounded process formulations and that improving the model simulations  
222 at the sub-daily scale improves the prediction of IAV of GPP. Additional analysis on pa-  
223 rameter inversion approaches and cost functions, as well as on parametric variability are  
224 treated in a companion paper [companion paper citation, in prep]. In this study, we aim  
225 to answer

- 226 1. How well does a mechanistic model perform compared to a semi-empirical model  
 227 across various temporal scales with different model parameterization approaches?  
 228 2. Can the performance of a mechanistic model be improved if drought stress is in-  
 229 cluded?  
 230 3. What factors influence the variability of model performance at different tempo-  
 231 ral scales?  
 232 4. How much are the differences in model performance between a mechanistic and  
 233 a semi-empirical model as well as across plant-functional types (PFT) and climate-  
 234 vegetation types?  
 235 5. Does improved simulations of peak diurnal GPP lead to improved simulations of  
 236 IAV of GPP?

## 237 2 Methods and data

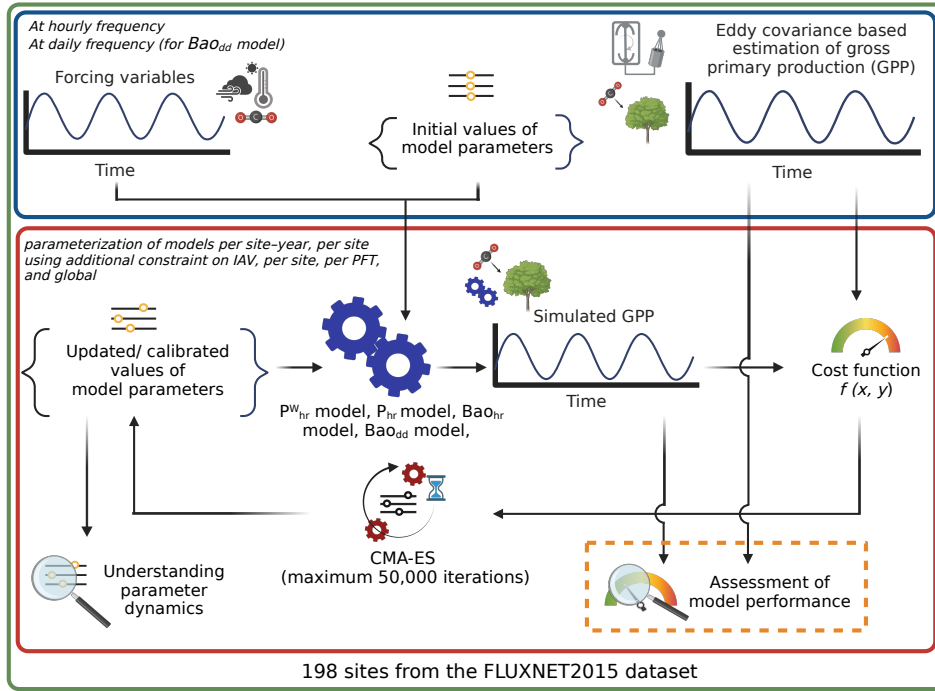
238 In this study, we focused on parameterization of both a semi-empirical model, at  
 239 daily and sub-daily scales, and a mechanistic model at a sub-daily scale using various  
 240 parameterization strategies consisting of different subsets of data and cost functions (Fig.  
 241 1). Thereafter, we performed forward runs of models with calibrated parameters at the  
 242 temporal resolution of model parameterization data and evaluated model performances  
 243 at different temporal aggregations (Fig. 1). The following sections describe each method-  
 244 ological step in a detailed manner.

### 245 2.1 Models

#### 246 2.1.1 Mechanistic model: *P-model of Mengoli*

247 Stocker et al. (2020) proposed the first version of the P-model based on theories  
 248 formulated by Wang et al. (2017), which unified the classic Farquhar–von Caemmerer–  
 249 Berry (FvCB) model (Farquhar et al., 1980) with the simplified formation of big leaf LUE  
 250 models (Monteith, 1972). The probable reasons behind using the ‘P’ in the P-model are  
 251 (1) ‘P’ stands for photosynthesis, (2) classically, GPP used to be denoted by ‘P’ (Monteith,  
 252 1972), and (3) the initial of the lead author (Prentice et al., 2014) who formulated the  
 253 theories behind the model starts with ‘P’ (B. D. Stocker, personal communication, May  
 254 06, 2024). The underlying equations of the P-model were formulated based on the op-  
 255 timality principle (Prentice et al., 2014) and the coordination principle (J.-L. Chen et  
 256 al., 1993; Maire et al., 2012). According to the optimality principle, plants aim to op-  
 257 timize the cost of transpiring water to assimilate CO<sub>2</sub> through the stomata. In the P-  
 258 model, the ratio of leaf internal and ambient CO<sub>2</sub> concentration ( $\chi = C_i/C_a$ ) is cal-  
 259 culated for which the above-described cost is minimal, and the sensitivity ( $\xi$ ) of  $\chi$  to VPD  
 260 is predicted. The coordination principle describes the achievement of equilibrium between  
 261 the maximum rate of carboxylation ( $V_{c_{max}}$ ) and electron transport ( $J_{max}$ ) by the plants.

262 Mengoli et al. (2022) adapted the first version of the P-model to simulate half-hourly  
 263 GPP dynamics. Here, we applied this same model at an hourly scale and called it the  
 264 P<sub>hr</sub> model. The major improvement in this version was defining an explicit differenti-  
 265 ation between instantaneous (such as RuBisCo and light-limited carbon assimilation),  
 266 and photosynthetic responses ( $V_{c_{max}}$ ,  $J_{max}$ , and  $\xi$ ) which acclimate over time in response  
 267 to environmental conditions. One of the important aspects of this P<sub>hr</sub> model is that the  
 268 parameters associated with cellular biochemistry acclimate to favourable conditions dur-  
 269 ing the day over a period of time or acclimation time ( $A_t$ ). In this study, we considered  
 270 the favourable condition as the average of three hourly input data points in the middle  
 271 of the day from 11:00 (hh:mm) LT, 12:00 (hh:mm) LT, and 13:00 (hh:mm) LT. A rolling  
 272 mean of the average condition from mid-day was taken over the  $A_t$ , which was used to  
 273 calculate optimal values of the model parameters, as described in Mengoli et al. (2022).  
 274 The value of  $A_t$  was calibrated as a parameter in our study (Table 1). We chose the mid-



**Figure 1.** Graphical representation of the model–data–integration workflow adopted in this study. The blue box indicates the preparation of forcing and observation data at hourly and daily scales for each site, as well as defines the initial value of parameters and their range by surveying literature. Then five different model parameterization tasks were performed for the light use efficiency (LUE) model from Bao et al. (2022) at hourly scale (Bao<sub>hr</sub> model) and at daily scale (Bao<sub>dd</sub> model), P-model from Mengoli et al. (2022) at hourly scale (P<sub>hr</sub> model), and P<sub>hr</sub> model with an explicit drought stress function (P<sub>hr</sub><sup>W</sup> model) using the Covariance Matrix Adaptation Evolution Strategy (CMA-ES) (Hansen & Kern, 2004), which is indicated by the red box. The cost function ( $f$ ) is a function of observed ( $y$ ) and simulated ( $x$ ) gross primary production. The green box denotes that the whole workflow was applied for the 198 sites from the FLUXNET2015 dataset (Pastorello et al., 2020). The dotted orange box highlights the focus of this study. The parameter dynamics is explored in detail in a companion paper [companion paper citation, in prep]. The figure was created in BioRender. De, R. (2024).

275 day and rolling mean approach from Mengoli et al. (2022) as it produced the best re-  
 276 sults in their evaluations of the P<sub>hr</sub> model at the half-hourly scale.

277 One of the known limitations of the P<sub>hr</sub> model is its tendency to overestimate GPP  
 278 fluxes in water-limited ecosystems, as no explicit representation of soil moisture condi-  
 279 tions was included (Mengoli et al., 2022, 2023). In order to relax such drawbacks here  
 280 we used the water availability index ( $WAI$ ) as a proxy of soil moisture (Tramontana et  
 281 al., 2016; Boese et al., 2019; Bao et al., 2022). The  $WAI$  represents the spatial and tem-  
 282 poral dynamics in plant available water based on a simple hydrological model where stor-  
 283 age is controlled by precipitation and evapotranspiration. We further introduced a drought  
 284 stress function that additionally scaled the GPP estimates of the P<sub>hr</sub> model, and we de-  
 285 noted this new version as the P<sub>hr</sub><sup>W</sup> model. We calibrated ten parameters in the P<sub>hr</sub><sup>W</sup> model  
 286 in which nine parameters were in the hydrological model and the drought stress func-  
 287 tion (Table 1). Further details on the implementation of the P<sub>hr</sub><sup>W</sup> model, along with the  
 288 drought stress function can be found in Sect. S1.1 and S1.2.

**Table 1.** Description, range, initial values, and units of calibrated model parameters

fX/ model name	Symbol	Definition	Initial value	Lower bound	Upper bound	Unit	Reference
$P_{hr}^W, P_{hr}$ models	$A_t$	Length of acclimation time	18	1	100	days	after Mengoli et al. (2022)
$Ba_{hr}, Ba_{dd}$ models	$\varepsilon_{max}$	Maximum light use efficiency	0.04	0	0.4	$\mu\text{molCO}_2 \cdot$ $\mu\text{mol photons}^{-1}$	Bao et al. (2022)
$fT$ ( $Ba_{hr}, Ba_{dd}$ models)	$T_{opt}$	Optimal temperature	10	5	35	$^{\circ}\text{C}$	Bao et al. (2022)
	$k_T$	Sensitivity to temperature changes	2	1	20	$^{\circ}\text{C}^{-1}$	Bao et al. (2022)
	$\alpha_{fT}$	Lag parameter for temperature effect	0.29	0	0.9	-	Bao et al. (2022)
$fVPD$ ( $Ba_{hr}, Ba_{dd}$ models)	$\kappa$	Sensitivity to VPD changes	$-5 \times 10^{-5}$	-0.01	$-1 \times 10^{-5}$	$\text{Pa}^{-1}$	Bao et al. (2022)
	$C_{\kappa}$	Sensitivity to atmospheric $\text{CO}_2$ concentration changes	0.4	0	10	-	Bao et al. (2022)
	$C_{a0}$	Minimum optimal atmospheric $\text{CO}_2$ concentration	380	340	390	ppm	Bao et al. (2022)
	$C_m$	$\text{CO}_2$ fertilization intensity indicator	2000	100	4000	ppm	Bao et al. (2022)
$fL$ ( $Ba_{hr}, Ba_{dd}$ models)	$\gamma$	Light saturation curvature indicator	$2 \times 10^{-3}$	0	0.05	$\mu\text{mol photons}^{-1}$ $\text{m}^2 \cdot \text{s}$	Bao et al. (2022)
$fCI$ ( $Ba_{hr}, Ba_{dd}$ models)	$\mu$	Sensitivity to cloudiness index changes	0.5	$10^{-3}$	1	-	Bao et al. (2022)
$fW$ ( $P_{hr}^W,$ $P_{hr}, Ba_{hr}, Ba_{dd}$ models)	$W_I$	Optimal soil moisture	0.26	0.01	0.99	$\text{mm} \cdot \text{mm}^{-1}$	Bao et al. (2022)
	$k_W$	Sensitivity to soil moisture changes	-11	-5	-30	-	Bao et al. (2022)
	$\alpha$	Lag parameter for soil moisture effect	0.98	0	1	-	Bao et al. (2022)
WAI ( $P_{hr}^W,$ $P_{hr}, Ba_{hr}, Ba_{dd}$ models)	$AWC$	Available water capacity	100	1	1000	mm	Bao et al. (2022)
	$\theta$	Rate of evapotranspiration	0.05	$10^{-4}$	0.1	$\text{mm} \cdot \text{h}^{-1}$	Bao et al. (2022)
	$PET_{scalar}$	Multiplicative scalar for potential evapotranspiration	1.2	0	5	-	Trautmann et al. (2018)
	$MR_{tair}$	Snow melt rate for temperature	0.125	0	0.5	$\text{mm} \cdot ^{\circ}\text{C}^{-1} \cdot$ $\text{h}^{-1}$	Trautmann et al. (2018)
	$MR_{netrad}$	Snow melt rate for net radiation	0.0375	0	0.125	$\text{mm} \cdot \text{MJ}^{-1} \cdot$ $\text{h}^{-1}$	Trautmann et al. (2018)
	$sn_a$	Sublimation resistance	0.44	0	3	-	Bao et al. (2022)



### 2.1.2 Semi-empirical model: Bao model

Vegetation stores energy from absorbed solar radiation in the form of biochemical energy through photosynthesis. The efficiency of the photosynthetic apparatus in performing this energy conversion is termed as light use efficiency ( $\varepsilon$ ). In a LUE model, GPP is calculated as the product of instantaneous  $\varepsilon$ , photosynthetic photon flux density ( $PPFD$ ), and the fraction of incident photosynthetically active radiation that is absorbed by vegetation ( $fAPAR$ ). Instantaneous  $\varepsilon$  reaches its maximum, i.e.,  $\varepsilon_{max}$ , when all environmental factors are optimal for photosynthesis. Instantaneous  $\varepsilon$  is determined as the product between  $\varepsilon_{max}$  and the partial sensitivity functions ( $fX$ ) for the different environmental factors controlling GPP, such as air temperature ( $T$ ),  $VPD$ , available soil water supply ( $W$ ), absorbed photosynthetic photon flux ( $L = PPFD \times fAPAR$ ), the cloudiness index ( $CI$ , Table A1), and atmospheric  $CO_2$  concentration (Mäkelä et al., 2008; Horn & Schulz, 2011; Bao et al., 2022).

$$GPP_{sim} = (\varepsilon_{max} \cdot fT \cdot fVPD \cdot fL \cdot fCI \cdot fW) \cdot PPFD \cdot fAPAR \quad (1)$$

$$fT = \frac{2 \cdot \exp\left(-\frac{T_f - T_{opt}}{k_T}\right)}{1 + \left(\exp\left(-\frac{T_f - T_{opt}}{k_T}\right)\right)^2} \quad (2)$$

$$T_f(t) = (1 - \alpha_{fT}) \cdot T(t) + \alpha_{fT} \cdot T_f(t-1) \quad (3)$$

$$fVPD = \exp\left(\kappa \cdot \left(\frac{C_{a0}}{CO_2}\right)^{C_\kappa} \cdot VPD\right) \cdot \left(1 + \frac{CO_2 - C_{a0}}{CO_2 - C_{a0} + C_m}\right) \quad (4)$$

$$fL = \frac{1}{\gamma(PPFD \cdot fAPAR) + 1} \quad (5)$$

$$fCI = CI^\mu \quad (6)$$

$$fW = \frac{1}{1 + \exp(k_W(W_{f_t} - W_I))} \quad (7)$$

$$W_{f_t} = (1 - \alpha) \cdot W_t + \alpha \cdot W_{f_{t-1}} \quad (8)$$

In this study, we used the LUE model of Bao et al. (2022, 2023) since it emerged as a robust representation from the systematic comparison across the large diversity of LUE formulations in the literature. The model selection followed a Bayesian approach that leveraged on the evaluation of modelling performance across FLUXNET EC sites (Pastorello et al., 2020) when forced and calibrated with daily data for each site. We denoted this model as the Bao<sub>hr</sub> model when we parameterized at hourly scale, and as the Bao<sub>dd</sub> model when we parameterized at daily scale. The model is described in Eqs. (1) to (8), where  $fT$ ,  $fVPD$ ,  $fW$ ,  $fL$ , and  $fCI$  are partial sensitivity functions for  $T$ ,  $VPD$ ,  $W$ ,  $L$ , and  $CI$ , respectively. In this case,  $W$  and  $fW$  were calculated similar to the implementation in P<sub>hr</sub><sup>W</sup> model, i.e., with a simple hydrological model (Sect. S1.1) and drought stress function (Eq. 7 and 8 are same as Eq. S1 and S2), respectively. Bold terms in the Eq. (1) to (8) are model parameters, and their initial values, units, and ranges are described in Table 1. The physical ranges for most of the parameters were based on Bao et al. (2022, 2023) and Trautmann et al. (2018). The  $fVPD$  term, viz. Eq. (4), also accounts for atmospheric  $CO_2$  concentration. The partial sensitivity functions range from zero to one (except the 2<sup>nd</sup> part of Eq. 4 which can be greater than one), where a value of zero completely diminishes, and of one completely favours GPP. In this study, we changed the denominator of Eq. (2) in comparison to the original exponential function  $\exp\left(-\frac{T_f - T_{opt}}{k_T}\right)^2$  of Bao et al. (2022, 2023), as the revised version produced a more realistic range of  $fT$  (Fig. S3). Sensitivity functions  $fT$  and  $fW$  also consider a lag effect of  $T$  and  $W$ . The lag effect of temperature was considered for Temperate, Boreal, and Polar regions where the first letter of the Köppen–Geiger (KG) climate class is ‘C’, ‘D’, ‘E’, and that of soil

332 water supply was considered for arid regions where the first letter of the KG climate class  
333 is ‘B’ (Rubel et al., 2017; Beck et al., 2018).

## 334 2.2 Data used

335 We selected 198 eddy covariance sites, for which the required forcing and obser-  
336 vation or derived data for model parameterization were available from the FLUXNET2015  
337 dataset (Pastorello et al., 2020; FLUXNET.org, 2024a). A list of these sites can be found  
338 in Table S2 (Sect. S3) and their spatial distributions is plotted in Fig. S4. The variables  
339 which were used to force, and parameterize models as well as data processing steps such  
340 as gap-filling, and quality control are described in detail in Table A1, Appendix B and  
341 Appendix C. We prepared these data in both hourly and daily resolutions.

342 In our study, a total of 13 different PFTs (as defined in FLUXNET.org, 2024b) were  
343 represented: croplands (CRO; 19 sites), deciduous broadleaf forests (DBF; 25 sites), de-  
344 ciduous needle leaf forest (DNF; one site), evergreen broadleaf forests (EBF; 13 sites),  
345 evergreen needle leaf forests (ENF; 47 sites), grasslands (GRA; 35 sites), mixed forests  
346 (MF; nine sites), closed shrublands (CSH; three sites), open shrublands (OSH; 13 sites),  
347 savannas (SAV; six sites), permanent wetlands (WET; 20 sites), woody savannas (WSA;  
348 six sites), and land cover under snow for most of the year (SNO; one site). The major  
349 KG climate classes (Rubel et al., 2017; Beck et al., 2018; FLUXNET.org, 2024c) are rep-  
350 resented by 12 tropical sites, 18 arid sites, 87 temperate sites, 71 boreal sites, and 10 po-  
351 lar sites. We also classified sites into 9 climate-vegetation types, similar to Bao et al. (2022),  
352 in which seven sites are tropical forests (TropicalF), five sites are tropical grassland (Trop-  
353 icalG), six sites are arid forest (AridF), 12 sites are arid grassland (AridG), 51 sites are  
354 temperate forest (TemperateF), 36 sites are temperate grassland (TemperateG), 52 sites  
355 are boreal forest (BorealF), 19 sites are boreal grassland (BorealG), and 10 sites have  
356 polar vegetation.

## 357 2.3 Model parameterization

358 We primarily defined four different parameterization strategies consisting of var-  
359 ious subsets of data to calibrate the model parameters controlling hourly GPP dynam-  
360 ics. These parameterization strategies were used to determine a vector of calibrated pa-  
361 rameter values (1) for each site-year, (2) for each site, (3) for each PFT, and (4) for all  
362 sites at once (global parameterization). We also performed another parameterization per  
363 site using a modified cost function which used an additional constraint on the IAV of GPP  
364 ( $Cost^{IAV}$ ). We parameterized and forced the  $Bao_{hr}$  model, and the  $Bao_{dd}$  model using  
365 hourly and daily data, respectively to perform a comparative analysis (Table 2). The  $P_{hr}^W$   
366 model and the  $P_{hr}$  model were only parameterized and forced using hourly data (Table  
367 2).

368 We used Python (Python Core Team, 2021) implementation (pyma v3.3.0.1) of  
369 the Covariance Matrix Adaptation Evolution Strategy (CMA-ES) (Hansen & Kern, 2004;  
370 Hansen et al., 2019) as our global search algorithm to find the values of model param-  
371 eters for which cost function reached its minimum. This is a derivative-free, evolution-  
372 ary algorithm, which is designed to find global minima in a rugged parameter space.

$$373 \quad Cost_{i_P} = (1 - GPP_{NNSE_i}) + (1 - ET_{NNSE_i}) \quad (9)$$

$$374 \quad Cost_{i_{Bao}} = (1 - GPP_{NNSE_i}) + (1 - ET_{NNSE_i}) + Cost_{ideal} + Cost_{non\_ideal} \quad (10)$$

375 A robust cost function is a necessity for the numerical optimizer to find the global  
376 minimum. The cost functions for  $P_{hr}^W$ ,  $P_{hr}$  models ( $Cost_{i_P}$ ) and the  $Bao_{hr}$ ,  $Bao_{dd}$  mod-  
377 els ( $Cost_{i_{Bao}}$ ) were calculated as Eq. (9) and (10), respectively, in case of per site-year

**Table 2.** Description of models and tasks accomplished with each specific model. The tasks are described in the footnote of the table.

Models	Description	Parameterization strategies				
		per site–year	per site using Cost <sup>IAV</sup>	per site	per PFT	global
P <sub>hr</sub>	P-model of Mengoli et al. (2022) parameterized using hourly data	a, d	a, d	a, d	a, d	a, d
P <sup>W</sup> <sub>hr</sub>	P-model of Mengoli et al. (2022) with an additional constraint on drought stress and parameterized using hourly data	a, b, d, e	a, b, d, e	a, b, d, e	a, b, d, e	a, b, d, e
Bao <sub>hr</sub>	LUE model of Bao et al. (2022) parameterized using hourly data	a, c, d, e	a, c, d, e	a, c, d, e	a, c, d, e	a, c, d, e
Bao <sub>dd</sub>	LUE model of Bao et al. (2022) parameterized using daily data	a, c, d	a, c, d	a, c, d	a, c, d	a, c, d

a: evaluation of model performance across timescale with different model types, parameterization strategies, and cost functions.  
b: evaluation of a mechanistic model with an explicit drought stress function.  
c: evaluation of a semi-empirical model with different temporal resolutions of data used for model parameterization.  
d: factors behind variability of model performance across timescales.  
e: variability of annual model performance with model performance in simulating diurnal gross primary productivity (GPP) peaks.

378 and per-site parameterization. Here,  $i$  is either a site or site-year based on parameter-  
379 ization type. For PFT-specific model parameterization, the cost functions were  $\sum_{i=1}^{N_{PFT}} Cost_{i_P}$   
380 and  $\sum_{i=1}^{N_{PFT}} Cost_{i_{Bao}}$  for P<sup>W</sup><sub>hr</sub>, P<sub>hr</sub> models and Bao<sub>hr</sub>, Bao<sub>dd</sub> models, respectively.  $i$  denotes  
381 a site and  $N_{PFT}$  denotes the total number of sites in a specific PFT. In the case of global  
382 model parameterization, the cost functions were  $\sum_{i=1}^N Cost_{i_P}$  and  $\sum_{i=1}^N Cost_{i_{Bao}}$  for the  
383 P<sup>W</sup><sub>hr</sub> model and Bao<sub>hr</sub>, Bao<sub>dd</sub> models, respectively.  $i$  denotes a site and  $N$  denotes the  
384 total number of sites used in this study.

$$385 \quad NNSE_i = \frac{1}{2 - NSE_i} \quad (11)$$

$$386 \quad NSE_i = 1 - \frac{\sum_{t=1}^{N_{t,i}} (\sigma_{weight_{t,i}} \cdot (EC_{t,i} - sim_{t,i}))^2}{\sum_{t=1}^{N_{t,i}} (\sigma_{weight_{t,i}} \cdot (EC_{t,i} - \overline{EC_{t,i}}))^2} \quad (12)$$

$$387 \quad \sigma_{weight_{t,i}} = 1 - \frac{\sigma_{t,i} - \min(\sigma_i)}{\max(\sigma_i) - \min(\sigma_i)} \quad (13)$$

388  $GPP_{NNSE_i}$  and  $ET_{NNSE_i}$  were calculated (Eq. 11) as a weighted normalized NSE,  
 389 viz. NNSE (Hundecha & Merz, 2012) between the time series of good quality data points  
 390 (see Appendix B for the selection criteria) of EC derived and simulated GPP and ET,  
 391 respectively. The GPP and ET derived from EC measurements are denoted as  $GPP_{EC}$   
 392 and  $ET_{LE}$ , respectively. The simulated GPP and ET are denoted as  $GPP_{sim}$  and  $ET_{sim}$   
 393 (see Fig. S1 for calculation of  $ET_{sim}$ ), respectively. We considered  $ET$  as well in our cost  
 394 function to better constrain the parameters of the simple hydrological model used in this  
 395 study. The NNSE values (Nash & Sutcliffe, 1970) are between zero and one, where one  
 396 is the best, and zero is the worst agreement between observed and simulated data. Here,  
 397 we used these normalized values so that minimizing  $(1 - NNSE)$  always results in bet-  
 398 ter model performance in comparison to using  $(1 - NSE)$ , where  $NSE$  can have values  
 399 between  $-\infty$  (worst agreement) and one (best agreement). In Eq. (12),  $N_{t,i}$  is the to-  
 400 tal number of good quality data points from each timestep  $t$  for a site-year or site  $i$ .  $\sigma$   
 401 in Eq. (13) is random uncertainty (which is the standard deviation of fluxes in a slid-  
 402 ing window of  $\pm 5$  days and  $\pm 1$  hour of the time-of-day of the current timestamp) of NEE  
 403 or ET (Table A1).

$$404 \quad Cost_{ideal} = ((1 - \max(fT_r)) + (1 - \max(fVPD_r)) + (1 - \max(fW_r)) \quad (14)$$

$$405 \quad + (1 - \max(fL_r))) \cdot 10^3$$

$$406 \quad Cost_{non\_ideal} = \sum_r ((fT_r - \theta_{fT})(T < 0^\circ C \ \& \ fT_r > \theta_{fT})) \quad (15)$$

$$407 \quad + \sum_r ((fVPD_r - \theta_{fVPD})(VPD > 2000 Pa \ \& \ fVPD_r > \theta_{fVPD}))$$

$$408 \quad + \sum_r ((fW_r - \theta_{fW})(W < 0.01 \ \& \ fW_r > \theta_{fW}))$$

409 The  $Cost_{ideal}$  and  $Cost_{non\_ideal}$  were introduced as a regularizers in  $Cost_{i_{Bao}}$  to avoid  
 410 over-fitting of the sensitivity functions (Bao et al., 2022, 2023). These cost function com-  
 411 ponents ensure that values of partial sensitivity functions were not penalized and favoured  
 412 under ideal and non-ideal conditions, respectively. The ideal and non-ideal conditions  
 413 were determined by certain constant thresholds for all sites. Equation (14) ensured that  
 414 the partial sensitivity functions,  $fT$  (Eq. 2), only left part of the  $fVPD$  (Eq. 4),  $fW$  (Eq.  
 415 7) and  $fL$  (Eq. 5) approaches one, when certain ideal environmental conditions ( $PPFD$   
 416  $\in [0 \text{ to } 600 \mu\text{mol photons} \cdot \text{m}^{-2} \cdot \text{s}^{-1}]$ ,  $fAPAR \in [0 \text{ to } 1]$ ,  $T \in [-5 \text{ to } 40 \text{ }^\circ\text{C}]$ ,  $VPD \in [0$   
 417  $\text{ to } 4500 \text{ Pa}]$ ,  $W \in [0 \text{ to } 1]$ ) occur (these ranges are denoted by subscript  $r$ ), so that the  
 418  $\varepsilon_{max}$  in Eq. (1) reaches its maximum potential. The factor  $10^3$  in Eq. (14) was included  
 419 to match the ranges of all other components in the cost function for the  $Bao_{hr}$ ,  $Bao_{dd}$   
 420 models ( $Cost_{i_{Bao}}$ ) so that all the components had equal weight. Equation (15) penalized  
 421 the cases when the values of  $fT$  (Eq. 2), only left part of  $fVPD$  (Eq. 4), and  $fW$  (Eq.  
 422 7), were greater than a certain threshold ( $\theta_{fT} = 0.2$ ,  $\theta_{fVPD} = 0.9$ ,  $\theta_{fW} = 0.2$ ) under non-  
 423 ideal conditions ( $T < 0 \text{ }^\circ\text{C}$ ,  $VPD > 2000 \text{ Pa}$ ,  $W < 0.01$ ) for photosynthesis.

$$424 \quad Cost_{iP}^{IAV} = (1 - GPP_{NNSE_i}) + (1 - GPP_{NNSE_i}^y) + (1 - ET_{NNSE_i}) \quad (16)$$

$$425 \quad Cost_{iBao}^{IAV} = (1 - GPP_{NNSE_i}) + (1 - GPP_{NNSE_i}^y) + (1 - ET_{NNSE_i}) + Cost_{ideal} \quad (17)$$

$$426 \quad + Cost_{non.ideal}$$

$$427 \quad GPP_{NNSE_i}^y = \frac{1}{2 - GPP_{NSE_i}^y} \quad (18)$$

$$428 \quad GPP_{NSE_i}^y = 1 - \frac{\sum_{t=1}^{N_{t,i}} (\sigma_{weight_{t,i}}^y \cdot (EC_{t,i}^y - sim_{t,i}^y))^2}{\sum_{t=1}^{N_{t,i}} (\sigma_{weight_{t,i}}^y \cdot (EC_{t,i}^y - \overline{EC}_{t,i}^y))^2} \quad (19)$$

$$429 \quad \sigma_{weight_{t,i}}^y = 1 - \frac{\sigma_{t,i}^y - \min(\sigma_i^y)}{\max(\sigma_i^y) - \min(\sigma_i^y)} \quad (20)$$

$$430 \quad EC_{t,i}^y = \sum_{t=1}^t EC_{t,y,i}; \quad sim_{t,i}^y = \sum_{t=1}^t sim_{t,y,i}; \quad \sigma_{t,i}^y = \sum_{t=1}^t \sigma_{t,y,i} \quad (21)$$

431 In the case of per-site-year parameterization using cost functions in Eq. (9) and  
 432 (10), we fitted the model so that the annual average of GPP can be captured well for  
 433 each site-year. Whereas, in the case of per-site parameterization using cost functions  
 434 in Eq. (9) and (10), the model was parameterized for each site. We performed another  
 435 experiment as a balance between these two experiments using the  $Cost^{IAV}$ , which is sim-  
 436 ilar to Desai (2010) to put an additional constraint on IAV, and parameterized  $P_{hr}^W$ ,  $P_{hr}$ ,  
 437  $Bao_{hr}$ , and  $Bao_{dd}$  models for each of the EC sites. The cost functions,  $Cost_{iP}^{IAV}$  for  $P_{hr}^W$ ,  
 438  $P_{hr}$  models (Eq. 16) and  $Cost_{iBao}^{IAV}$  for  $Bao_{hr}$ , and  $Bao_{dd}$  models (Eq. 17) now include  
 439 an additional term  $(1 - GPP_{NNSE_i}^y)$  to constrain the annual cumulative sum of GPP  
 440 flux from each site  $i$ .  $EC_{t,i}^y$ ,  $sim_{t,i}^y$ , and  $\sigma_{t,i}^y$  (Eq. 21) are cumulative sums of  $GPP_{EC}$ ,  
 441  $GPP_{sim}$ , and  $\sigma_{NEE}$  from start of each year  $y$  to timestep  $t$  for each site  $i$ , respectively.

## 442 2.4 Simulating and evaluating GPP estimates

### 443 2.4.1 Forward runs

444 In the case of the site-year parameterization, we performed a forward run for each  
 445 site-year using the respective set of calibrated parameter values and forcing data for that  
 446 year. Afterwards, we concatenated  $GPP_{sim}$  from all the years for a given site to assess  
 447 model performance. For per-site parameterization using  $Cost^{IAV}$ , and per-site param-  
 448 eterization, we used site-specific values of calibrated parameters to perform site-level model  
 449 evaluation. We also applied calibrated model parameters for a certain PFT to simulate  
 450 GPP at all the sites which belong to a certain PFT. Similarly, for the global param-  
 451 eterization, a single set of calibrated parameter values was used to simulate GPP for each  
 452 site.

### 453 2.4.2 Model performance metrics

454 We performed forward runs at an hourly scale and averaged the hourly simulations  
 455 to daily, weekly, monthly, and annual temporal frequencies to calculate model perfor-  
 456 mance measures at different temporal aggregations. Model performance was only eval-  
 457 uated for temporal aggregations from daily to annual for the  $Bao_{dd}$  model. We applied  
 458 a data screening procedure (Appendix C) before calculating model performance mea-  
 459 sures. We evaluated how well a model can simulate the IAV of GPP based on how well  
 460 a model simulated the annual average GPP for a site. In this study, we performed most  
 461 of our analysis using NSE (Nash & Sutcliffe, 1970) and normalized NSE, viz. NNSE (which  
 462 is  $\frac{1}{2 - NSE}$ ) as NSE indicates the degree to which scatter between observed and simulated  
 463 data fits to the 1:1 line. In addition, we calculated the square of the Pearson correlation

464 coefficient ( $R^2$ ) (PCC, 2008) which explains whether the dispersion of observed and sim-  
 465 ulated data matches and in the case of an unbiased model, values of NSE will be closer  
 466 to values of  $R^2$ . Whereas, if a model is systematically biased, it will result in higher  $R^2$   
 467 values, but bad NSE values (Krause et al., 2005). We also calculated Root Mean Squared  
 468 Error (RMSE) (Chai & Draxler, 2014) to quantify how closely the mean of simulated  
 469 data matches with the mean of the observed data.

$$NSE = 2 \cdot \alpha_{NSE} \cdot r - \alpha_{NSE}^2 - \beta_n^2 \quad (22)$$

$$\alpha_{NSE} = \frac{\sigma_{sim}}{\sigma_{EC}} \quad (23)$$

$$\beta_n = \frac{\mu_{sim} - \mu_{EC}}{\sigma_{EC}} \quad (24)$$

470 Moreover, using Eq. (22) to (24), we decomposed NSE values to linear correlation  
 471 ( $r$ ), relative variability ( $\alpha_{NSE}$ ), and bias ( $\beta_n$ ) in some cases to investigate which of these  
 472 were improved or diminished between different model parameterization strategies (Gupta  
 473 et al., 2009). In Eq. (23), and (24),  $\sigma_{sim}$  and  $\sigma_{EC}$  are standard deviations of  $GPP_{sim}$   
 474 and  $GPP_{EC}$ , respectively,  $\mu_{sim}$  and  $\mu_{EC}$  are mean  $GPP_{sim}$  and  $GPP_{EC}$ , respectively.  
 475 We calculated these metrics using the Python (Python Core Team, 2021) package Per-  
 476 metrics v1.5.0 (Van Thieu, 2023; Van Thieu & Mirjalili, 2023), and the definition of each  
 477 of the model performance metrics can be found in the package documentation.

### 478 **2.4.3 Factors associated with simulating GPP flux**

479 We selected potential factors that can affect model performance at different tem-  
 480 poral resolutions. These factors can be of two types. There were factors which we de-  
 481 termined based on our experiment design, which included model types ( $P_{hr}^W$  model,  $P_{hr}$   
 482 model,  $Bao_{hr}$  model, and  $Bao_{dd}$  model), parameterization strategies (per site-year, per  
 483 site using  $Cost^{IAV}$ , per site, per PFT, and global parameterization), number of years  
 484 with good quality data (Appendix C) in a site. Whereas, other factors represent site-  
 485 specific characteristics, including PFT, KG climate class, and climate-vegetation types.

486 First, we conducted Levene’s test (Levene, 1960) to find out if the assumption of  
 487 homoscedasticity is fulfilled across groups in the controlling factors. Then, we performed  
 488 an N-way Analysis of Variance (ANOVA) (Kaufmann & Schering, 2014) with the po-  
 489 tential controlling factors to determine which of them played a major role in deter-  
 490 mining model performance at hourly and annual temporal scales. For analysis at an hourly  
 491 scale, the  $Bao_{dd}$  model was not included as this model produced simulations at a daily  
 492 scale. We performed two N-way ANOVA analyses once including the performance of the  
 493  $P_{hr}$  model, and then excluding the performance of the  $P_{hr}$  model. The Levene’s test and  
 494 N-way ANOVA analyses were implemented using SciPy v1.11.3 (Virtanen et al., 2020)  
 495 and statsmodels v0.14.0 (Seabold & Perktold, 2010), respectively.

### 496 **2.4.4 Evaluating GPP estimates in water-limited ecosystems**

497 We investigated to determine whether explicit accounting of the drought stress func-  
 498 tion in the  $P_{hr}^W$  model had improved its performance at arid sites. For this purpose, we  
 499 chose the aridity index (AI) to determine which sites were arid or semi-arid, as this in-  
 500 dex provided a numerical representation of moisture availability (Zomer et al., 2022) at  
 501 a location. The AI values were calculated by dividing the average precipitation ( $P$ ) per  
 502 hour by the average potential evapotranspiration ( $PET$ ) per hour for the whole obser-  
 503 vation period at a site.

504 We drew examples from a few site-specific results to highlight different aspects of  
 505 the behaviour of  $P_{hr}^W$  and  $P_{hr}$  models for ecosystems with contrasting soil moisture con-

506 trols on GPP and with a larger availability of good-quality measurements. For this purpose,  
 507 we chose a water-limited semi-arid site (annual average precipitation of 318 mm)  
 508 in central Australia (Alice Springs, AU-ASM). This site also features a complex mixture  
 509 of Mulga woodland and savanna (Cleverly et al., 2013; Pastorello et al., 2020). In contrast,  
 510 we also highlighted the behaviours of  $P_{hr}^W$  and  $P_{hr}$  models in an irrigated cropland  
 511 (Mead - irrigated continuous maize site, US-Ne1) in the mid-western U.S.A (Amos et  
 512 al., 2005; Pastorello et al., 2020).

#### 513 **2.4.5 Effect of temporal resolution of data used for model parameter-** 514 **ization on model performance**

515 We parameterized the LUE model of Bao et al. (2022) with hourly and daily data  
 516 for  $Bao_{hr}$  model and  $Bao_{dd}$  model, respectively. We performed a comparison between  
 517 these two versions of the model to highlight whether the resolution of data used for model  
 518 parameterization can substantially affect the prediction of the annual average or IAV of  
 519 GPP fluxes. Here, we also drew a site-specific example from an energy-limited decid-  
 520 uous forest in central Germany (Hainich, DE-Hai) as this site had a very long observa-  
 521 tion period (Knohl et al., 2003).

#### 522 **2.4.6 Evaluation between modelling experiments of various complexi-** 523 **ties**

524 We formulated our experiments using models and parameterization strategies con-  
 525 sisting of varying numbers of model parameters to be calibrated. The number of param-  
 526 eters calibrated for a detailed parameterization strategy, such as per site-year param-  
 527 eterization was substantially higher than a generic parameterization strategy, such as global  
 528 parameterization. We used Akaike’s Information Criterion (AIC) to investigate whether  
 529 a complex modelling experiment with a higher number of parameters can better simu-  
 530 late GPP (Burnham & Anderson, 2004).

$$AIC = n \log \left( \frac{\sum (EC_i - sim_i)^2}{n} \right) + 2K \quad (25)$$

$$AIC_c = n \log \left( \frac{\sum (EC_i - sim_i)^2}{n} \right) + 2K + \frac{2K(K+1)}{n-K-1} \quad (26)$$

531 Following recommendations of Burnham and Anderson (2004), we used Eq. (25)  
 532 to calculate AIC when  $n/K > 40$ , where  $n$  is the total number of observations and  $K$   
 533 is the total number of parameters. Otherwise, we used a corrected version of AIC ( $AIC_c$ ,  
 534 Eq. 26). Though the values of AIC or  $AIC_c$  can be in any range, the lowest value of AIC  
 535 or  $AIC_c$  determines the preferred modelling experiments.  $EC_i$  and  $sim_i$  are  $i^{\text{th}}$  obser-  
 536 vations of EC-derived GPP and simulated GPP, respectively in Eq. (25) and (26). We  
 537 considered  $GPP_{sim}$  from all the four variations of models, i.e.,  $P_{hr}^W$  model,  $P_{hr}$  model,  
 538  $Bao_{hr}$  model, and  $Bao_{dd}$  model for calculation of AIC or  $AIC_c$ . We calculated AIC at  
 539 hourly and daily aggregations by concatenating good quality (Appendix C) hourly or  
 540 daily data, and daily averages  $GPP_{EC}$  and  $GPP_{sim}$  from all the days from all sites. Sim-  
 541 ilarly, we used monthly and annual aggregations for calculating  $AIC_c$  at monthly and  
 542 annual scales, respectively.  $AIC_c$  was calculated at monthly and annual aggregation, as  
 543  $n$  was usually smaller than  $K$  in these cases. The value of  $K$  was the total number of  
 544 model parameters calibrated for all the site-years, for all the sites, for all the PFT, and  
 545 for a specific model in case of per site-year parameterization, per site parameterization  
 546 using  $Cost^{IAV}$ , per site parameterization, per PFT parameterization, and global param-  
 547 eterization, respectively.

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### 2.4.7 Simulating GPP peaks

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We assessed model performance in predicting peak  $GPP_{EC}$ . We defined peak  $GPP_{EC}$  and peak  $GPP_{sim}$  as the 90<sup>th</sup> percentiles of hourly  $GPP_{EC}$  ( $P90_{GPP_{EC}}$ ) and  $GPP_{sim}$  ( $P90_{GPP_{sim}}$ ), respectively, following the concept of good hours by Zscheischler et al. (2016) and Fatichi and Ivanov (2014). We calculated  $P90_{GPP_{EC}}$  and  $P90_{GPP_{sim}}$  for each site-year considering only good quality hourly data (Appendix B). We compared the ratios of peak  $GPP_{sim}$  from  $P_{hr}^W$  model and  $Bao_{hr}$  model to  $GPP_{EC}$  for each parameterization strategy in order to identify possible biases.

$$\Delta NNSE_{P90} = NNSE_{P90}^{j1} - NNSE_{P90}^{j2} \quad (27)$$

$$\Delta NNSE_y = NNSE_y^{j1} - NNSE_y^{j2} \quad (28)$$

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We furthermore investigated whether improving the simulation of peaks of  $GPP_{EC}$  improved the simulation of IAV of GPP. We calculated NNSE between  $P90_{GPP_{EC}}$  and  $P90_{GPP_{sim}}$  ( $NNSE_{P90}^j$ ) from all the site-years in a site considering only good site-years and only for sites with more than 3 years of good quality data (Appendix C) for a parameterization strategy  $j$ . Similarly, we calculated NNSE between the annual average of  $GPP_{EC}$  and  $GPP_{sim}$  ( $NNSE_y^j$ ) for sites with more than 3 years of good quality data (Appendix C) for a parameterization strategy  $j$ . Then, differences between  $NNSE_{P90}^{j1}$  and  $NNSE_y^{j2}$  were calculated for a pair of parameterization strategies where  $j1$  and  $j2$  are two different parameterization experiments, for both  $P_{hr}^W$  model and  $Bao_{hr}$  model (Eq. 27 and 28). Correlation between  $\Delta NNSE_{P90}$  and  $\Delta NNSE_y$  were then investigated to study whether a certain parameterization strategy for a given model better captured the  $GPP_{EC}$  peaks, and thus contributed to higher annual model performance.

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## 3 Results

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### 3.1 Overall model performance

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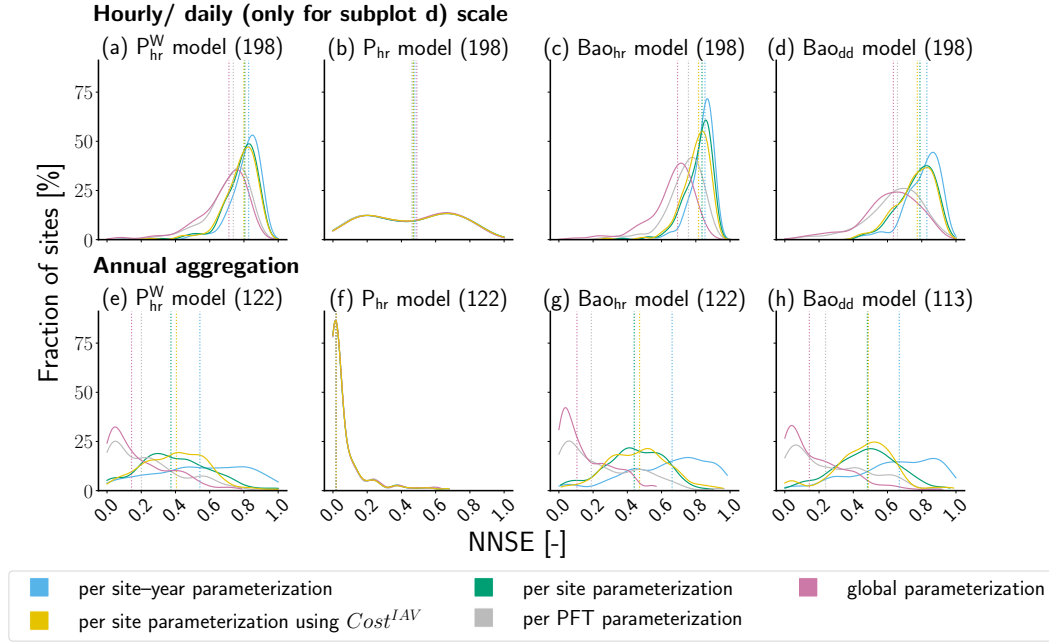
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All four models, i.e.,  $P_{hr}^W$ ,  $P_{hr}$ ,  $Bao_{hr}$ , and  $Bao_{dd}$  models performed significantly better at the hourly scale than the annual scale (Fig. 2). The use of an additional constraint on IAV, i.e.,  $Cost^{IAV}$  did not contribute to better model performance across sites at an annual scale and performed closer to parameterization per site and poorer than site-year parameterization (Fig. 2). The median model performance was highest for the model parameterization per site-year among all model parameterization strategies (per site-year, per site using  $Cost^{IAV}$ , per site, per PFT, and global parameterization) for all four models (Table D1). Model parameterization per site-year also produced the best model performance at all temporal aggregation levels including annual aggregation (Fig. 2 and Sect. S2.1).  $P_{hr}^W$  model performed substantially better for the majority of the sites compared to  $P_{hr}$  model at all temporal aggregation levels as it explicitly considered site-specific water availability (Fig. 2 and Sect. S2.1). Comparison of model performances at different temporal aggregations also revealed that  $Bao_{hr}$  and  $Bao_{dd}$  models performed slightly better than the  $P_{hr}^W$  model across all timescales (hourly, daily, weekly, monthly, and annual), as the  $Bao_{hr}$  and  $Bao_{dd}$  models were more flexible than the  $P_{hr}^W$  model and captured ecosystem response with a broad range of parameters (Fig. 2, Table D1 and Sect. S2.1). For example, the median NNSE(s) at the hourly resolution were 0.827 and 0.853 for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively. Conversely, at the annual resolution, the median NNSE(s) were 0.543 and 0.661 for the  $P_{hr}^W$  model and  $Bao_{hr}$  model, respectively.





**Figure 2.** Distributions of model performance measure (normalized Nash-Sutcliffe efficiency, viz. NNSE) at hourly/daily scale (first row) and at annual timescale (second row) from P-model of Mengoli et al. (2022) with drought stress, parameterized at hourly scale ( $P_{hr}^W$ ), P-model of Mengoli et al. (2022) without drought stress, parameterized at hourly scale ( $P_{hr}$ ), global best model of Bao et al. (2022) parameterized at hourly scale ( $Bao_{hr}$ ), and global best model of Bao et al. (2022) parameterized at daily scale ( $Bao_{dd}$ ). For the  $Bao_{dd}$  model, subplot (d) shows model performance at daily scale as this model was parameterized at daily scale.  $Cost^{IAV}$  denotes the usage of an additional constraint on annual gross primary production flux during per-site parameterization. The dotted vertical lines represent the median model performances, which are summarized in Table D1. The numbers in parentheses beside the model name on top of each of the sub-figures represent the total number of sites. The model performance at an annual scale was calculated for fewer sites as some sites have a very low measurement period (Appendix C).

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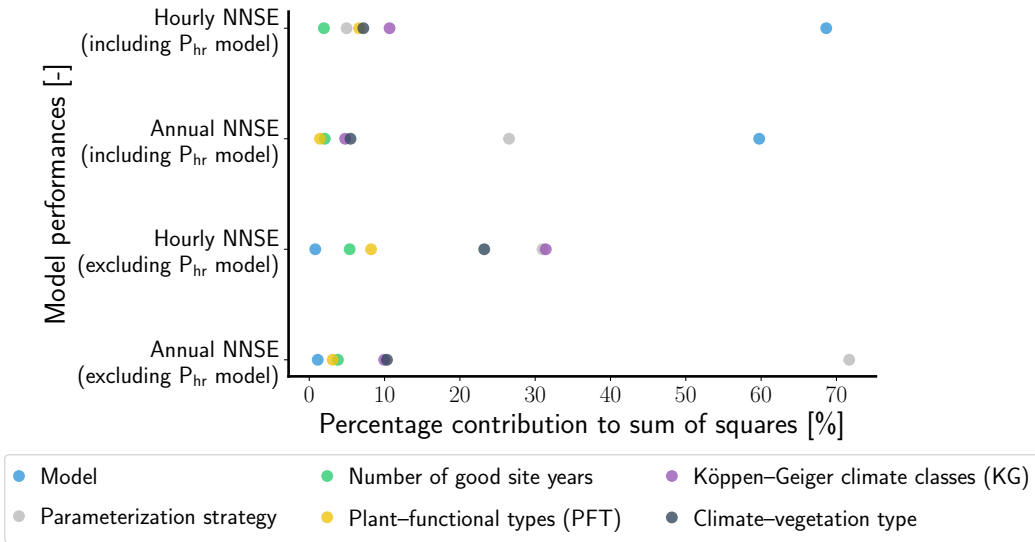
### 3.2 Factors behind variability in model performance

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We summarized the percentage contributions of factors which influenced model performance at hourly and annual scales and found that most of the variability in model performance came from how we designed our modelling experiments (Fig. 3). Model types was a crucial factor when the  $P_{hr}$  model was included in N-way ANOVA analysis, as this model had comparatively poor performance at both hourly and annual scales and resulted in greater variability in the NNSE values (69% and 60% contribution to the sum of squares of the regression, viz. SSR in hourly and annual scale, respectively). We then excluded the  $P_{hr}$  model from further analysis to uncover the other factors behind the model performance and found that for the hourly scale, the model performance varied the most across the groups of KG classes (31.4% contribution to the SSR), followed by parameterization type (31.0% to the SSR) and climate-vegetation type (23.2% contribution to the SSR) (Fig. 3). However, at an annual scale, the parameterization strategy strongly affected (71.7% contribution to the SSR) the model performance, as per-site-year parameterization usually better simulated the annual  $GPP_{obs}$  compared to other parameterization strategies. The number of good years (Appendix C) used for calculating an-

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606 nual NNSE also exerted a small influence (3.8% contribution to the SSR) on the annual  
 607 model performance. In general, there were only slight performance differences between  
 608 models when the  $P_{hr}$  model was not considered, and model parameterization played a  
 609 bigger role in the variability of model performance.



**Figure 3.** Percentage contributions of factors influencing variability in model performance (normalized Nash-Sutcliffe efficiency, viz. NNSE) in the sum of squares in N-way Analysis of Variance (ANOVA). The percentage contributions show the influence of various factors on hourly and annual model performance when the  $P$ -model (Mengoli et al., 2022) without any explicit drought stress function, parameterized at hourly scale ( $P_{hr}$  model) was considered in the analysis, as well as on hourly and annual model performance excluding the  $P_{hr}$  model. The sum of squares of residual was removed before plotting the percentage contributions of the factors and only the explained variance is shown.

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### 3.3 Effect of drought stress on model performance

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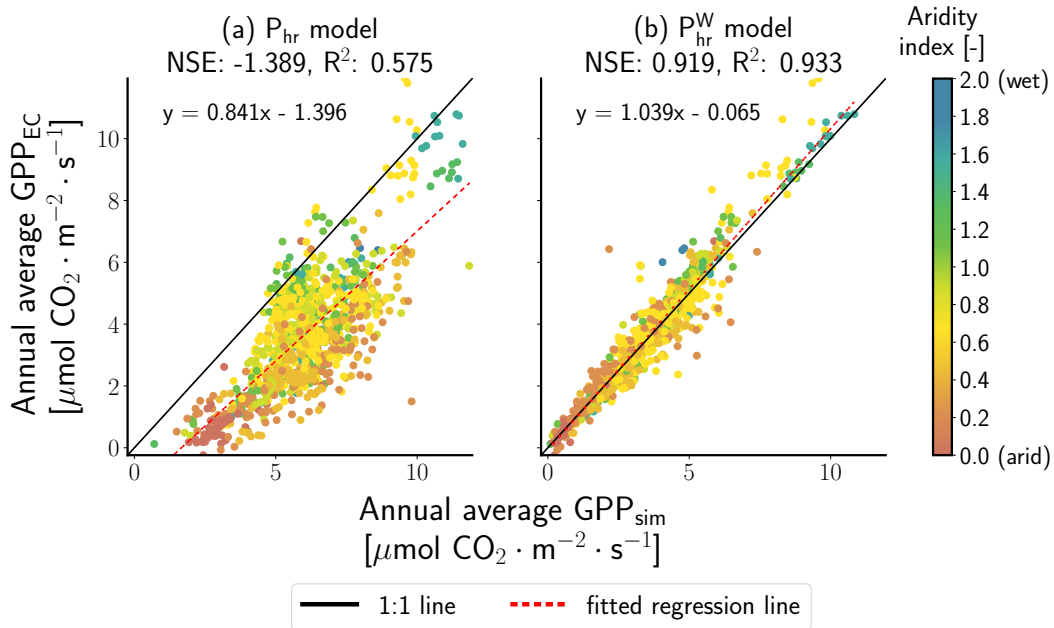
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The performance of the  $P_{hr}^W$  model to predict the annual average of  $GPP_{EC}$  from each of the site-years substantially improved in comparison to  $P_{hr}$  model (from an NSE of -0.94 to 0.93) after explicit consideration of soil water supply in the model (Fig. 4). Most of this improvement came from the better prediction of the annual average of  $GPP_{EC}$  at the arid and semi-arid sites (with AI values lower than 0.5). For the semi-arid site (AU-ASM) the predicting performance of the  $P_{hr}^W$  model for all the parameterization strategies largely benefited from the explicit inclusion of soil water supply constraints (Fig. 5). The systematic bias in model simulations was also improved after the inclusion of a drought stress constraint, as well as the modelling bias also improved from a generalized to a detailed parameterization strategy. Although the coupling of a simple hydrological model which calculated water-availability based on precipitation and evapotranspiration and inclusion of drought stress function generally improved the  $P_{hr}^W$  model for most of the site-years at an arid site, the model failed to capture the  $GPP_{EC}$  (Sect. S2.2, Fig. S6) at an irrigated cropland site (US-Ne1), as the simple hydrological model which we used to calculate water-availability lacked representation of human management.



**Figure 4.** Scatter plot of the annual average (from good quality site–years, see Appendix C) of eddy covariance measurements derived gross primary production ( $GPP_{EC}$ ) versus simulated gross primary production ( $GPP_{sim}$ ) from P-model of Mengoli et al. (2022) parameterized at hourly scale (a) without drought stress ( $P_{hr}$  model) and (b) with drought stress ( $P_{hr}^W$  model). The results in this plot are from parameterization for each site–year. We only used good–quality site–years in this figure (Appendix C). The dots in the scatter represent a site–year and are coloured by the aridity index (AI) of the site. The model performance metrics (Nash–Sutcliffe efficiency, viz. NSE) are shown at the top of each subplot. The equations of fitted regression lines are shown in respective subplots.

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### 3.4 Effect of temporal resolution of the data used for model parameterization on model performance

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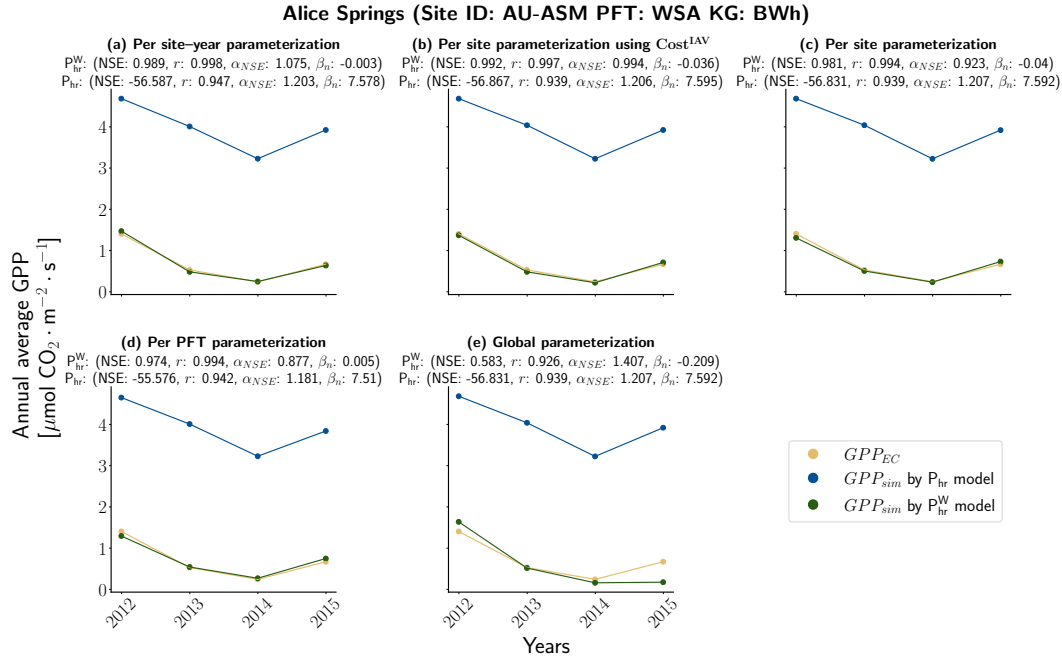
The use of hourly data to constrain  $Bao_{hr}$  model parameters and aggregating hourly values of  $GPP_{sim}$  to annual scale did not have a significant effect on the  $Bao_{hr}$  model performance in comparison to parameterization of the same model with daily data, i.e.,  $Bao_{dd}$  model, in simulating the annual average of  $GPP_{EC}$  (Fig. 6) for each site–year. The value of NSE slightly decreased from 0.964 to 0.956 for the  $Bao_{hr}$  model compared to the  $Bao_{dd}$  model, and both models performed almost equally well. Here, for the  $Bao_{hr}$  model, we also focus on a site-specific example at a site (DE-Hai) in central Germany with a deciduous broadleaf forest where the  $Bao_{hr}$  model proved to be capable of simulating annual average of  $GPP_{EC}$  flux relatively well when the model was parameterized for each site–year and each site (Fig. 7). However,  $GPP_{EC}$  was underestimated in cases of PFT-specific and global parameterization. For this specific site, the  $Bao_{hr}$  model performed relatively better in comparison to  $Bao_{dd}$  model (Fig. 7 and S7).

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### 3.5 Role of parameterization strategies on model performance

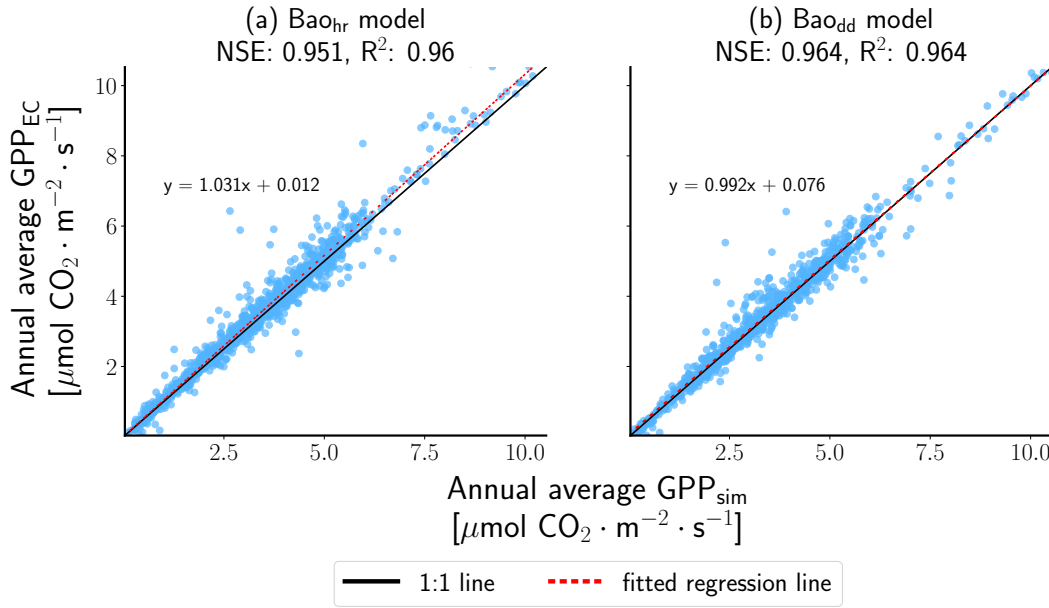
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Model performances at an annual scale generally increased with a more detailed parameterization strategy (Fig. 8). For the  $P_{hr}^W$  model, the median differences in annual model performance between the most detailed parameterization strategy, i.e., site–year



**Figure 5.** Comparison of annual average of gross primary production (GPP) derived by eddy covariance measurements ( $GPP_{EC}$ ), and GPP simulated ( $GPP_{sim}$ ) by the P-model of Mengoli et al. (2022) parameterized at hourly scale without drought stress ( $P_{hr}$  model) and with drought stress ( $P_{hr}^W$  model). The five subplots show simulated GPP from (a) site-year specific parameterization, (b) site-specific parameterization using an additional constraint on inter-annual variability in the cost function ( $Cost^{IAV}$ ), (c) site-specific parameterization, (d) plant-functional types (PFT) specific parameterization, and (e) global parameterization. The values of model performance measures (Nash-Sutcliffe efficiency, viz. NSE, correlation coefficient, viz.  $r$ , relative variability, viz.  $\alpha_{NSE}$ , and bias, viz.  $\beta_n$ ) are shown on top of respective subplots. This site is dominated by Mulga (*Acacia aneura*), and had an annual average temperature of  $\approx 22^\circ\text{C}$ , and an annual average precipitation of  $\approx 318$  mm during the observation period (Cleverly et al., 2013; Pastorello et al., 2020). The site ID, PFT, and Köppen-Geiger climate class (KG) of the site are provided on top of the figure in bold.

644 parameterization, and other detailed parameterization strategies, i.e., per site param-  
 645 eterization using  $Cost^{IAV}$  and per site parameterization were small, which were 0.12,  
 646 and 0.11, respectively. In contrast, the median differences in annual model performance  
 647 between the most detailed parameterization strategy, i.e., site-year parameterization, and  
 648 other generalized parameterization strategies, i.e., PFT-specific parameterization, and  
 649 global parameterization were quite large, which were 0.28, and 0.37, respectively. Sim-  
 650 ilarly, for the  $Bao_{hr}$  model, the median differences in annual model performance between  
 651 the most detailed parameterization strategy, i.e., site-year parameterization, and other  
 652 detailed parameterization strategies, i.e., per site parameterization using  $Cost^{IAV}$  and  
 653 per site parameterization were 0.20, and 0.21, respectively. In contrast, the median dif-  
 654 ferences in annual model performance between the most detailed parameterization strat-  
 655 egy, i.e., site-year parameterization, and other generalized parameterization strategies,  
 656 i.e., PFT-specific parameterization, and global parameterization were 0.36, and 0.50,  
 657 respectively. The positive values of median annual model performance confirm the high-

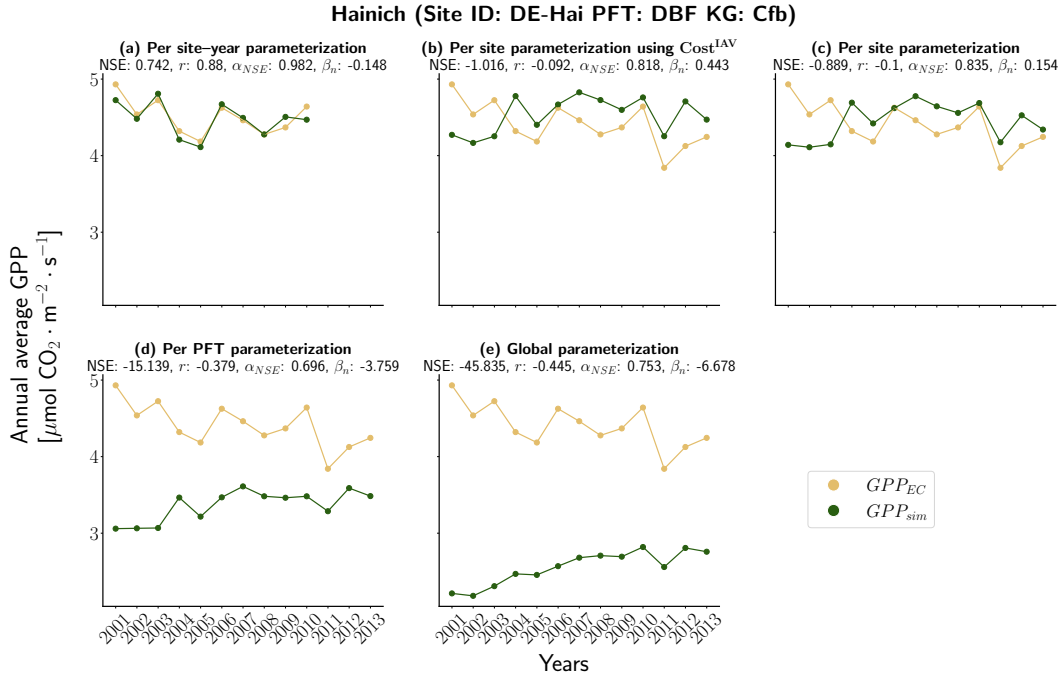


**Figure 6.** Scatter plot of annual average (from good quality site-years, see Appendix C) eddy covariance derived gross primary production ( $GPP_{EC}$ ) versus simulated gross primary production ( $GPP_{sim}$ ) by the light use efficiency model of Bao et al. (2022) parameterized at hourly (Bao<sub>hr</sub> model) and daily scale (Bao<sub>dd</sub> model) for each site-year. The plots show the performance of the (a) Bao<sub>hr</sub> model, and (b) Bao<sub>dd</sub> model. The dots in the scatter represent a site-year. The model performance metrics (Nash-Sutcliffe efficiency, viz. NSE) are shown on the top of each subplot. The equations of fitted regression lines are shown in respective subplots.

658 est median performance of site-year parameterization compared to the other four pa-  
 659 rameterization strategies.

660 At an hourly scale, differences in model performance between a pair of similar pa-  
 661 rameterization strategies, such as a pair of detailed parameterization (i.e., between site-  
 662 year-specific and site-specific) or a pair of generalized parameterization (i.e., between per  
 663 PFT and global) approaches for both models were small (Fig. S8). However, this dif-  
 664 ference can be higher between a detailed and a generalized model parameterization strat-  
 665 egy. The median differences in hourly NNSE between site-year-specific and site-specific  
 666 model parameterization were 0.02 and 0.01 for the  $P_{hr}^W$  model and Bao<sub>hr</sub> model, respec-  
 667 tively. In contrast, the median differences in hourly NNSE between site-year-specific and  
 668 global model parameterization were 0.11 and 0.16 for the  $P_{hr}^W$  model and Bao<sub>hr</sub> model,  
 669 respectively.

670 The median annual model performance between per-site parameterization using  
 671  $Cost^{IAV}$  and per-site parameterization were relatively small, which were 0.01 and 0.02  
 672 for  $P_{hr}^W$  model and Bao<sub>hr</sub> model, respectively, and it shows the additional constraint on  
 673 IAV of GPP flux in the cost function did not substantially improve annual model per-  
 674 formance. At hourly scale, the median differences in model performance between per-  
 675 site parameterization using  $Cost^{IAV}$  and per-site parameterization were also negligible,  
 676 which were 0.00 and -0.01 for  $P_{hr}^W$  model and Bao<sub>hr</sub> model, respectively. Though the per-  
 677 site parameterization using  $Cost^{IAV}$  did not improve the annual model performance, it  
 678 also did not degrade the hourly model performance.



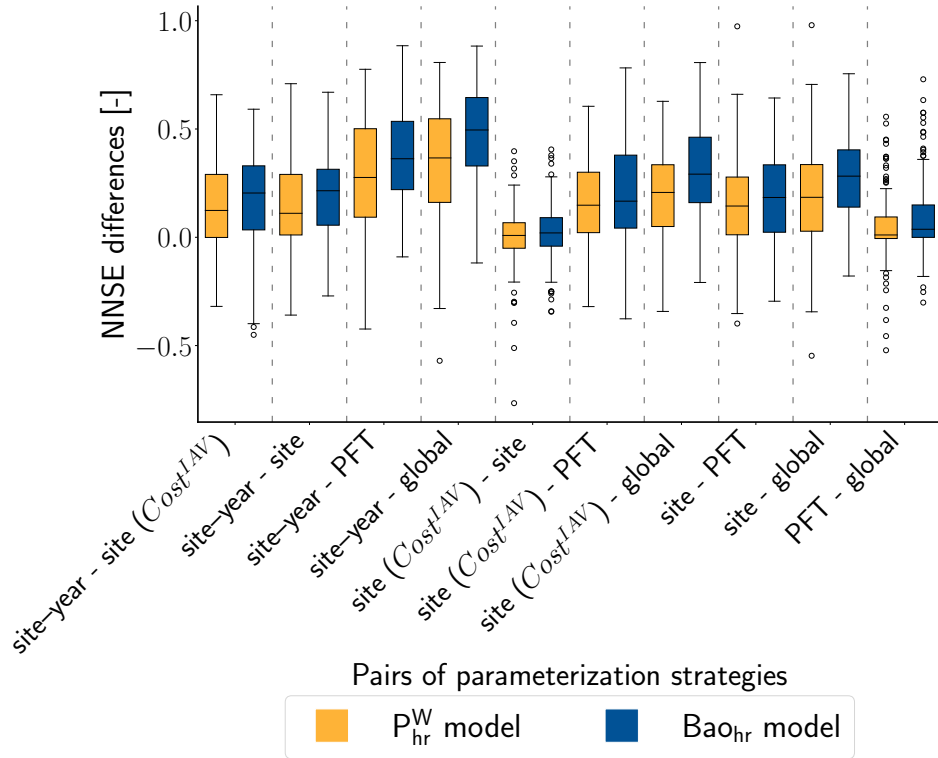
**Figure 7.** Comparison of annual average of gross primary production (GPP) derived by eddy covariance measurements ( $GPP_{EC}$ ), and GPP simulated ( $GPP_{sim}$ ) by the light use efficiency model of Bao et al. (2022), which was parameterized with hourly data ( $Bao_{hr}$  model). The five subplots show simulated GPP from (a) site-year specific parameterization, (b) site-specific parameterization using an additional constraint on inter-annual variability in the cost function ( $Cost^{IAV}$ ), (c) site-specific parameterization, (d) plant-functional types (PFT) specific parameterization, and (e) global parameterization. The years 2010 to 2012 could not be parameterized in the case of site-year parameterization, as there were no good quality evapotranspiration estimates from latent heat flux measurements for those years. The values of model performance measures (Nash-Sutcliffe efficiency, viz. NSE, correlation coefficient, viz.  $r$ , relative variability, viz.  $\alpha_{NSE}$ , and bias, viz.  $\beta_n$ ) are shown on top of respective subplots. This site represents an average 140-year-old deciduous forest (Tamrakar et al., 2018) with a distinct seasonal cycle and an annual average temperature of  $\approx 8.3$  °C, and an annual average precipitation of 750–800 mm during the observation period (Knohl et al., 2003; Pastorello et al., 2020). The site ID, PFT, and Köppen-Geiger climate class (KG) of the site are provided on top of the figure in bold.

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### 3.6 Differences between modelling experiments involving a mechanistic and a semi-empirical model structure

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The lowest AIC values were obtained for per site-year parameterization for all the models at hourly and daily scales or aggregations, suggesting the sum of squares errors ( $SSE$ ) was substantially reduced even when a comparatively complex parameterization strategy with a large number of model parameters was chosen (Table 3). The AIC values gradually increased from per site-year, per site, per PFT to global parameterization at hourly and daily scales or aggregations for all three models which are  $P_{hr}^W$  model,  $Bao_{hr}$  model, and  $Bao_{dd}$  model (Table 3). Semi-empirical models, i.e.,  $Bao_{hr}$  model and  $Bao_{dd}$  model also had mostly lower values of AIC compared to mechanistic  $P_{hr}^W$  model even though more parameters were parameterized for these models (Table 3). At the daily scale, the  $Bao_{dd}$  model had the lowest AIC for all the parameterization experiments due to the pa-



**Figure 8.** Distributions of the differences between model performance measures (normalized Nash-Sutcliffe efficiency, viz. NNSE) calculated at annual scale, from various pairs of model parameterization experiments conducted for the P-model of Mengoli et al. (2022) with drought stress, parameterized with hourly data ( $P_{hr}^W$  model) and the light use efficiency model of Bao et al. (2022), parameterized with hourly data ( $Bao_{hr}$  model).  $Cost^{IAV}$  in parentheses denotes the usage of an additional constraint on annual gross primary production flux during per-site parameterization. The boxes are spanned between the first and third quartiles of the differences, and the line in the middle represents the median. The whiskers show the farthest data point from the box within  $1.5\times$  of the interquartile range. The circles represent the outliers that go beyond the limits of the whiskers. The vertical dotted grey line separates each pair of model parameterization strategies.

691 parameterization at daily scale. Whereas, for the other two models, parameterization and  
 692 forward runs were performed at an hourly scale and then simulations were aggregated  
 693 to daily resolution. The  $P_{hr}$  model was not included in AIC or  $AIC_c$  analysis as previ-  
 694 ous results proved this model significantly underperformed compared to the other mod-  
 695 els, and this will always result in higher AIC or  $AIC_c$  values.

696 At monthly and annual scales, we show the differences in  $AIC_c$  values between  $P_{hr}^W$ ,  
 697  $Bao_{hr}$ , and  $Bao_{dd}$  models for the same parameterization strategy, and not between pa-  
 698 rameterization strategies in a same model. The reason behind this is  $AIC_c$  values largely  
 699 depend on the relationship between sample size, i.e.,  $n$ , and the total number of param-  
 700 eters which were parameterized, i.e.,  $K$ . The values of  $AIC_c$  became very large even when  
 701 a significantly smaller  $SSE$  was obtained, and they became unreliable when the value  
 702 of  $n$  was closer to  $K$ . For example, at monthly aggregation, per site-year parameteri-  
 703 zation of the  $P_{hr}^W$  model had a very high  $AIC_c$  value of  $1.30\times 10^6$  even when it had the  
 704 lowest  $SSE$  among all the five parameterization strategies (Table S3, S4, and S5). The

705 Bao<sub>dd</sub> model proved to be better able to capture the seasonal cycle, i.e., monthly GPP  
 706 estimates compared to the other two models for most of the parameterization experiments  
 707 considering the number of parameters parameterized (Table 3). However, the P<sub>hr</sub><sup>W</sup> model  
 708 had the lowest AIC<sub>c</sub> value in the case of per site parameterization using *Cost*<sup>IAV</sup> and  
 709 per site parameterization at monthly aggregation. In contrast, at an annual scale, the  
 710 P<sub>hr</sub><sup>W</sup> model had mostly the lowest AIC<sub>c</sub> values, and some of the experiments also suffered  
 711 from the above-described unreliable AIC<sub>c</sub> estimates, where *n* and *K* had similar values  
 712 (Table 3, S3, S4, and S5).

**Table 3.** Akaike’s Information Criterion (AIC) or corrected AIC (AIC<sub>c</sub>) values for modelling experiments of various complexities

Temporal scale/ aggregation	Models	parameterization strategies				
		per site–year	per site using <i>Cost</i> <sup>IAV</sup>	per site	per PFT	global
Hourly scale (AIC)	P <sub>hr</sub> <sup>W</sup>	$1.72 \times 10^7$	$1.84 \times 10^7$	$1.86 \times 10^7$	$2.16 \times 10^7$	$2.25 \times 10^7$
	Bao <sub>hr</sub>	$1.57 \times 10^7$	$1.71 \times 10^7$	$1.78 \times 10^7$	$2.11 \times 10^7$	$2.41 \times 10^7$
Daily scale/ aggregation (AIC)	P <sub>hr</sub> <sup>W</sup>	$4.58 \times 10^5$	$5.05 \times 10^5$	$5.11 \times 10^5$	$6.88 \times 10^5$	$7.42 \times 10^5$
	Bao <sub>hr</sub>	$3.86 \times 10^5$	$4.35 \times 10^5$	$4.50 \times 10^5$	$6.74 \times 10^5$	$8.00 \times 10^5$
	Bao <sub>dd</sub>	$2.78 \times 10^5$	$3.25 \times 10^5$	$3.44 \times 10^5$	$5.10 \times 10^5$	$5.65 \times 10^5$
Monthly aggregation (AIC <sub>c</sub> )	P <sub>hr</sub> <sup>W</sup>	$1.30 \times 10^6$	$1.63 \times 10^4$	$1.63 \times 10^4$	$2.04 \times 10^4$	$2.25 \times 10^4$
	Bao <sub>hr</sub>	$-4.71 \times 10^4$	$1.81 \times 10^4$	$1.77 \times 10^4$	$2.01 \times 10^4$	$2.42 \times 10^4$
	Bao <sub>dd</sub>	$-3.39 \times 10^4$	$1.70 \times 10^4$	$1.73 \times 10^4$	$1.55 \times 10^4$	$1.79 \times 10^4$
Annual aggregation (AIC <sub>c</sub> )	P <sub>hr</sub> <sup>W</sup>	$-3.56 \times 10^3$	$-5.22 \times 10^3$	$-5.56 \times 10^3$	$9.09 \times 10^2$	$9.28 \times 10^2$
	Bao <sub>hr</sub>	$-3.93 \times 10^3$	$-3.87 \times 10^3$	$-4.07 \times 10^3$	$1.45 \times 10^3$	$1.21 \times 10^3$
	Bao <sub>dd</sub>	$-3.70 \times 10^3$	$-3.51 \times 10^3$	$-3.53 \times 10^3$	$1.10 \times 10^3$	$8.31 \times 10^2$

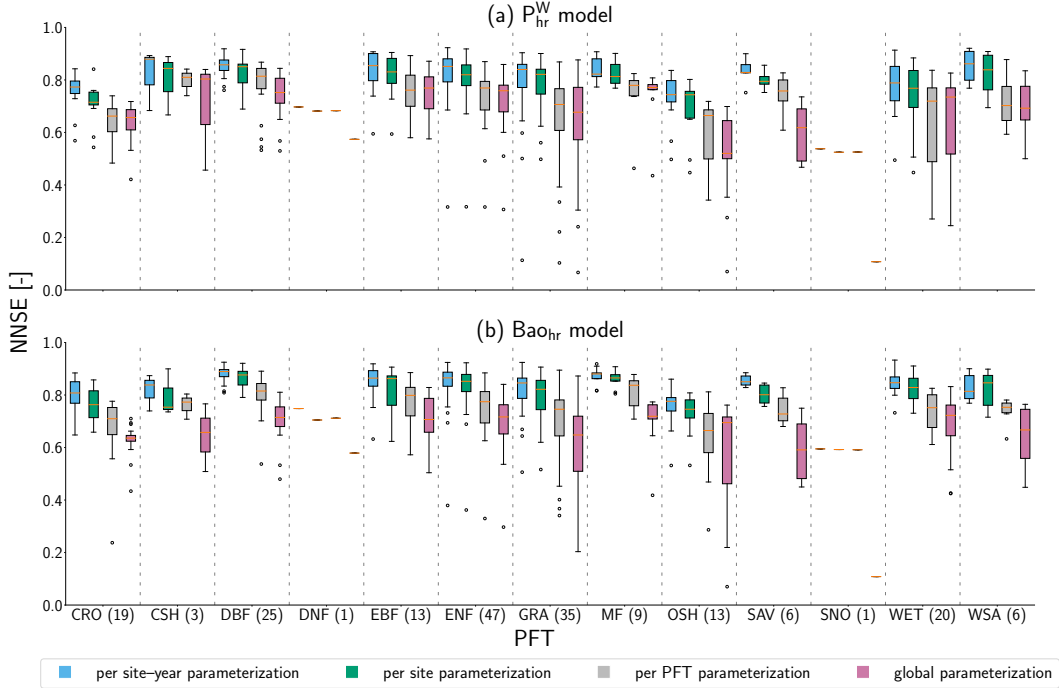
### 713 3.7 Model performances across different plant–functional types

714 Generally better model performances were achieved with both the P<sub>hr</sub><sup>W</sup> model and  
 715 the Bao<sub>hr</sub> model when parameterized detailed model parameterization strategies were  
 716 used (Fig. 9). In this analysis, we removed the per–site parameterization experiment us-  
 717 ing *Cost*<sup>IAV</sup> as it performed very similar to per–site parameterization, and also we did  
 718 not consider the P<sub>hr</sub> model as it produced poor performance across all the PFTs. The  
 719 highest median NNSEs were obtained with per–site–year parameterization for all the PFTs.  
 720 For the P<sub>hr</sub><sup>W</sup> model parameterization experiments, the highest median value of NNSE was  
 721 found for CSH for per–site–year parameterization (median NNSE: 0.88), DBF for per–  
 722 site parameterization (median NNSE: 0.85), CSH and DBF for per-PFT parameteriza-  
 723 tion (median NNSE: 0.81), and CSH for global parameterization (median NNSE: 0.80).  
 724 However, CSH had only three sites and highest median model performance for CSH should  
 725 be interpreted with caution. For the Bao<sub>hr</sub> model parameterization experiments, the high-  
 726 est median value of NNSE was found for DBF for per–site–year parameterization (me-  
 727 dian NNSE: 0.89), DBF for per–site parameterization (median NNSE: 0.88), MF for per-  
 728 PFT parameterization (median NNSE: 0.84), and DBF and MF for global parameter-  
 729 ization (median NNSE: 0.77). We found similar results also for climate–vegetation types



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(Sect. 2.2), where a more detailed parameterization strategy achieved higher model performance than a generalized parameterization strategy (Sect. 2.6 and Fig. S9).



**Figure 9.** Box-plots showing the range of the hourly model performance metric (normalized Nash-Sutcliffe efficiency, viz. NNSE), for the sites in different plant-functional types (PFT), and different parameterization experiments. The subplots show the model performance for (a) P-model of Mengoli et al. (2022) with drought stress function, parameterized with hourly data ( $P_{hr}^W$  model), and (b) the light use efficiency model of Bao et al. (2022) parameterized with hourly data ( $Bao_{hr}$  model). The numbers in parentheses beside the name of each PFT on the x-axis are the number of sites present in a specific PFT. The boxes are spanned between the first and third quartiles of NNSE values, and the line in the middle represents the median. The whiskers show the farthest data point from the box within  $1.5 \times$  of the interquartile range. The circles represent the outliers that go beyond the limits of the whiskers. For, deciduous needle-leaf forests (DNF), and areas covered by snow (SNO) only the median value could be shown as these PFTs have only one site. The vertical dotted grey line separates each PFT.

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### 3.8 Correlation between annual model performance and model performance in simulating diurnal GPP peaks

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One of the crucial reasons behind poor annual model performance (Fig. 2 and Table D1) can be the inability of both the  $P_{hr}^W$  model and the  $Bao_{hr}$  model to capture the peaks of  $GPP_{EC}$  (Fig. S10). Specifically,  $P90_{GPP_{EC}}$  was highly underestimated in the case of global parameterization. The median of the ratio of  $P90_{GPP_{sim}}$  to  $P90_{GPP_{EC}}$  were 0.77 and 0.53 for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively, during global parameterization. The underestimation generally decreased with more detailed parameterization strategies with little differences between both the models. The median values of the ratio of  $P90_{GPP_{sim}}$  to  $P90_{GPP_{EC}}$  were 0.95 and 0.93 for the site-year parameterization of the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively. Moreover, the lower values of the interquartile range (IQR) of these ratios signify the importance of site-year

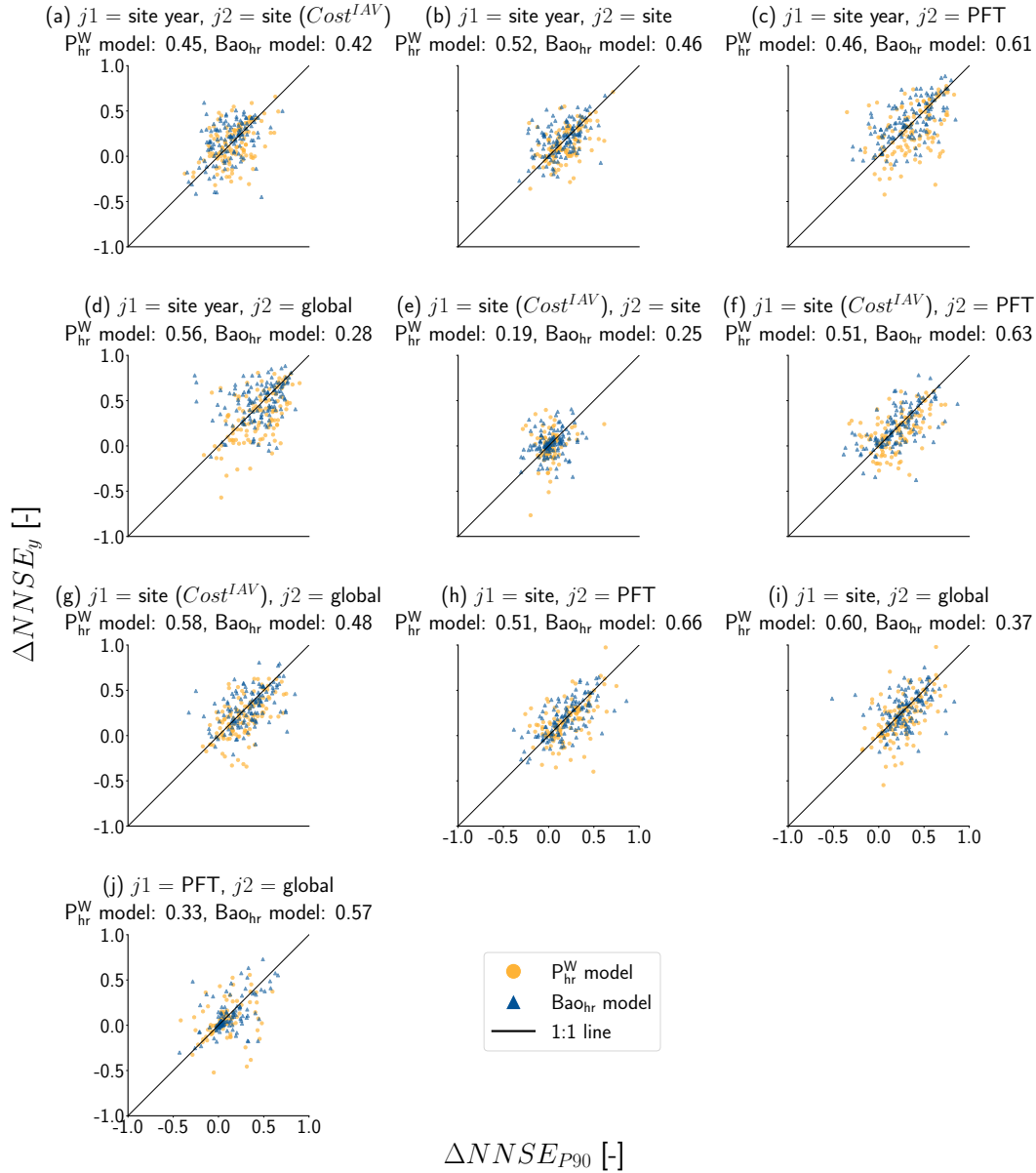
parameterization compared to per PFT or global parameterization to reliably capture the peak  $GPP_{EC}$  in diurnal cycles for most of the sites and to attain better model performance at the sub-daily scale. The values of IQR were 0.1 for both the  $P_{hr}^W$  model and the  $Bao_{hr}$  model in the case of site-year parameterization, 0.44 and 0.38 for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively in the case of PFT-specific parameterization, and 0.44 and 0.42 for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively in the case of global parameterization.

We further found that if a certain parameterization strategy better simulated the  $P90_{GPP_{EC}}$  for each site-year, it corresponded to a comparatively better annual model performance for a site which is demonstrated by the positive values of Pearson correlation coefficients (Fig. 10). Here also, a detailed parameterization strategy, such as site-year parameterization resulted in a better simulation of  $P90_{GPP_{EC}}$ , and thus better annual model performance for most of the sites compared to a generalized parameterization strategy, such as global parameterization. In this case when  $j1$  was site-year parameterization and  $j2$  was global parameterization, 91% and 93% sites had higher  $NNSE_{P90}^{j1}$  than  $NNSE_{P90}^{j2}$  and corresponding  $NNSE_y^{j1}$  than  $NNSE_y^{j2}$  for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively. When  $j1$  was parameterization per site using  $Cost^{IAV}$  and  $j2$  was parameterization per site, respectively, only 34.43% and 41.8% had positive values of  $\Delta NNSE_{P90}$  (i.e.,  $NNSE_{P90}^{j1} > NNSE_{P90}^{j2}$ ) and corresponding  $\Delta NNSE_y$  (i.e.,  $NNSE_y^{j1} > NNSE_y^{j2}$ ) for the  $P_{hr}^W$  model and the  $Bao_{hr}$  model, respectively. This signified that using an additional constraint related to the IAV of GPP in the cost function during model parameterization did not improve the prediction of peak GPP values for most of the sites.

## 4 Discussion

### 4.1 Uncertainties in modelling experiments

Any model-data-integration study is prone to uncertainties related to both data and the model. The EC dataset used in our study has a couple of well-known uncertainties. For example, the NEE measurements from eddy covariance can have uncertainties due to the accumulation of atmospheric  $CO_2$  under the canopy at night (storage) and a sudden turbulent mixing during the morning when the stable nighttime boundary layer breaks up, or because of advection of atmospheric  $CO_2$  out of the control volume sampled by the eddy covariance system (D. Baldocchi et al., 2000; Aubinet, 2008; Jocher et al., 2018). The GPP fluxes that we used were derived from NEE measurements by extrapolating the nighttime respiration of ecosystem (Reichstein et al., 2005) to daytime. Moreover, GPP can be estimated based on another well-known algorithm, the daytime partitioning method (Lasslop et al., 2010). We preferred nighttime partitioning as only respiration is modelled in this method. In daytime partitioning, both GPP and respiration are modelled, resulting in higher prediction errors. We believe the uncertainties in our modelling results due to the choice of partitioning algorithm should be small as quantified in a previous study by Desai et al. (2008). We also used ET in the cost function which is equivalent to latent heat flux. The mismatch between the summation of latent, sensible, and ground heat fluxes with net radiation calculated using incoming and outgoing radiation, the so-called lack of energy-balance closure, remains a long-standing challenge with EC measurements (Foken, 2008; Mauder et al., 2020; Zhang et al., 2024). Quality control of millions of data points at an hourly scale was also challenging, especially when we merged data from various sources, such as in-situ measurements, modelled re-analysis data, and remote sensing-based estimates. Another major uncertainty arises from the mismatch between the footprint of EC towers and the grid of remote sensing data which were used to calculate vegetation indices (Chu et al., 2021). The PFT classification of sites based on a simple PFT classification method may not accurately represent the vegetation of some of the sites. For example, a site at Alice Springs (AU-ASM) in central Australia was classified as a savanna in FLUXNET2015 (Pastorello et



**Figure 10.** Scatter between differences in model performance in simulating peak gross primary production, viz. GPP ( $\Delta NNSE_{P90}$ ) and annual average of GPP ( $\Delta NNSE_y$ ).  $j_1$  and  $j_2$  are a pair of parameterization strategies for which differences are calculated in each subplot from (a) to (j). Each dot in the scatter represents a site. The  $P_{hr}^W$  model and  $Bao_{hr}$  model are P-model of Mengoli et al. (2022) with drought stress function and the light use efficiency model of Bao et al. (2022), both were parameterized with hourly data. The values on top of each subplot indicate the Pearson correlation coefficient (PCC, 2008) for respective models.

796 al., 2020). In fact, this site is dominated by a discontinuous canopy of Mulga (*Acacia*  
 797 *aneura*) that has needle leaves and a seasonal understory grassy layer (Cleverly et al.,  
 798 2013). This site can be classified as a woody savanna as well. An arctic site in Bayelva  
 799 (SJ-Blv) has a combination of snow, wet grounds, and specific tundra vegetation (Boike  
 800 et al., 2018) which were not well represented by the snow classification of FLUXNET2015

(Pastorello et al., 2020). Another limitation of the dataset is that sites are mostly clustered in European and North American countries and, hence do not necessarily represent global ecosystem functioning particularly well due to sampling bias (Papale et al., 2015). Similarly, some PFTs are represented by very few sites, which makes PFT-specific parameterization challenging.

## 4.2 General performance of models in simulating GPP

In this study, we evaluated the performance of the optimality-based  $P_{hr}^W$  and  $P_{hr}$  models across a wide range of sites, representing various vegetation and climate types. We uncovered poor model performance of the  $P_{hr}$  model at many sites, especially at arid sites. Calculating *WAI* using a simple hydrological model and inducing a moisture stress function, i.e., the introduction of the  $P_{hr}^W$  model substantially improved model performance to simulate the annual average of GPP fluxes across many sites, including both water-limited and energy-limited sites. The inclusion of the moisture stress function to the  $P_{hr}$  model not only improved the annual model performance but also improved the model performance across all the temporal scales or aggregation levels. This highlights the importance of the representation of soil moisture conditions for modelling approaches, which are aimed at accurately representing ecosystem functioning and vegetation response. However, the coupling of the hydrological model raised the need to calibrate nine more parameters, which counters the vision of developing a parameter-sparse approach using theories that demand a lower site or site-year specific fine-tuning of model parameters (Prentice et al., 2015). Further experimentation is needed to find a fine balance between the number of key model parameters, which require calibration, and an accurate representation of ecosystem functioning.

Coming to the differences in model structure, we found that semi-empirical models ( $Bao_{hr}$  and  $Bao_{dd}$  models) performed statistically better, i.e., had a lower value of AIC compared to mechanistic model ( $P_{hr}^W$  model) at hourly and daily scale or aggregations for most of the parameterization experiments even though the semi-empirical modelling experiments needed more parameters to be parameterized. At the monthly aggregation level, the seasonal cycles were also significantly better captured by the parameter-heavy semi-empirical model parameterized with daily data ( $Bao_{dd}$  model) for most of the parameterization experiments. However, at the annual aggregation level, the mechanistic model, i.e., the  $P_{hr}^W$  model was comparatively better in most cases and a more flexible semi-empirical model with a higher number of parameters did not have a substantial improvement in annual model performance.

Though the partial sensitivity functions of environmental variables used in the  $Bao_{hr}$  model and the  $Bao_{dd}$  model were found to be applicable for most of the sites, they can be of many different types and may vary across site conditions (Bao et al., 2022). The EC sites were also affected by human management, such as irrigation, harvesting, and mowing as well as natural disturbances, such as fire, and pest attacks. These factors can affect the IAV of GPP flux which was estimated from EC measurements. Models used in this study may not be able to account for all of these factors due to structural limitations. For example, in the hydrological model, we only used precipitation and ET to calculate the mass balance of water. However, human management (such as irrigation and drainage) can play an important role, and the *WAI* estimates in managed sites, such as at an irrigated maize site (US-Ne1) may not be accurate.

## 4.3 The importance of the parameterization approach to estimate IAV of GPP

We also emphasize the importance of determining the parameterization approach by the inter-comparison of parameterization strategies. parameterization of model parameters largely determines model performance and parameterized parameters capture

851 the individual characteristics of sites or climatic events of site–years (Wu et al., 2012).  
 852 Detailed model parameterization strategies, such as parameterization specific to site–years  
 853 also comparatively better predicted the annual average of GPP fluxes and year-specific  
 854 parameters explained some parts of the IAV of GPP flux. As the fast rate of change in  
 855 climatic characteristics has become more frequent in recent years, developing a gener-  
 856 alized model structure to simulate carbon fluxes between years and/or between sites of  
 857 similar vegetation types has become even more challenging (Knauer et al., 2023). More-  
 858 over, the generalized model parameterization strategy, i.e., global parameterization was  
 859 also dominated by PFTs, such as ENF and GRA which were represented by many sites,  
 860 and certain PFT, such as DNF was represented by only one site in the FLUXNET2015  
 861 dataset (Pastorello et al., 2020). This may imply that global parameterization or param-  
 862 eter up-scaling experiments using the FLUXNET2015 dataset (Pastorello et al., 2020)  
 863 may result in biased parameter sets that cannot be generalized to the global scale or a  
 864 weighted site representation may be necessary in this case. Besides model parameter-  
 865 ization, a recent study by (Zou et al., 2024) also found that the importance of each in-  
 866 dependent driver, and their relative contributions varies over time and ecosystem type.  
 867 The relative importance of forcing variables may also be another factor besides model  
 868 parameterization.

869 Though we have demonstrated the capability of the  $P_{hr}^W$  model that included drought  
 870 stress and the  $Ba_{o_{hr}}$  model to simulate the hourly fluxes of GPP, accurate estimation  
 871 of IAV of GPP fluxes at the site level with these models requires further developments.  
 872 Particularly, both models failed to capture the peak GPP in diurnal cycles at many sites  
 873 even after model parameterization at a sub-daily scale and using an additional constraint  
 874 on the IAV of GPP in the cost function. These underestimations at an hourly scale may  
 875 have accumulated to a larger error when the fluxes were aggregated at an annual scale  
 876 to study the IAV. We also showed that comparatively better model performances were  
 877 achieved when the GPP peaks per site–year were better simulated. These results are sim-  
 878 ilar to another study by Lin et al. (2023), directed at evaluating terrestrial ecosystem  
 879 models’ capability in explaining the IAV of GPP which also found an underestimation  
 880 of GPP. However, it is also true that some peak values of diurnal GPP can also be an  
 881 outlier produced by data processing algorithms, such as the gap-filling algorithm, and  
 882 it is hard to distinguish these outliers from true GPP values. Moreover, it was found both  
 883 the  $P_{hr}^W$  model and the  $Ba_{o_{hr}}$  model showed the highest model performance at the sub-  
 884 daily scale mostly for forest sites compared to savannas or grasslands, this in turn led  
 885 to the poor simulation of IAV at many sites which are not forests.

886 The poor representation of IAV of GPP can be attributed to either limitations of  
 887 models or model parameterization strategies. It is important to discover which seasonal  
 888 phases of the GPP dynamics for particular vegetation types or climatic zones are not  
 889 well represented in models simulating the IAV of GPP. It is particularly important to  
 890 focus on the meteorological sensitivity of GPP during periods of high productivity where  
 891 improvements in the prediction of high fluxes would tend to improve the description of  
 892 IAV. Another aspect could also be to decompose the metric (Gupta et al., 2009) used  
 893 in the cost function or develop a more detailed model evaluation to understand which  
 894 other parts of the time series were not well constrained during model parameterization.

## 895 5 Conclusions

896 We have demonstrated the capability of an improved version of an optimality-based  
 897 mechanistic model ( $P_{hr}^W$  model) and a semi-empirical LUE model ( $Ba_{o_{hr}}$  and  $Ba_{o_{dd}}$  model)  
 898 to simulate sub-daily GPP fluxes across 198 EC sites, representing 13 different vegeta-  
 899 tion types including forests, grasslands, savannas, croplands, and tundra. We conclude  
 900 that explicit accounting of drought stress in the optimality-based ecosystem model is a  
 901 necessity as it proved to be an important factor in controlling GPP fluxes at all tempo-  
 902 ral scales including at annual aggregation. We found that the semi-empirical model mostly

903 produced better results at hourly, daily, and monthly scales compared to the mechanis-  
 904 tic model. However, at an annual scale, the improvement in the performance of the semi-  
 905 empirical model was not significant even though more parameters were parameterized  
 906 to flexibly capture the ecosystem dynamics. While these models generally performed well  
 907 in simulating hourly GPP dynamics, the small errors at the sub-daily scale, particularly  
 908 related to the estimation of GPP peaks, accumulated to bigger errors at the annual scale  
 909 and led to poor performance of models in explaining the IAV of GPP. We found that com-  
 910 paratively better annual model performance could be achieved when the peaks of GPP  
 911 were better simulated. Moreover, both models performed better mostly at forest sites  
 912 compared to grasslands or savannas which may also lead to poor estimation of IAV of  
 913 GPP at many sites globally. Our results further suggest the need to focus on sub-daily  
 914 GPP dynamics during the various seasonal phases, especially highly productive ones, to-  
 915 wards an improved constraint on GPP sensitivities. Hence, better annual model perfor-  
 916 mance with a detailed parameterization strategy, such as site-year parameterization, sig-  
 917 nifies that temporally varying model parameters are necessary to better capture the vari-  
 918 ations of annual average GPP and indicate that ecosystem functioning is not stable be-  
 919 tween years. We believe these new understandings can guide us towards developing mod-  
 920 els and parameterization strategies for simulating the inter-annual variations in ecosys-  
 921 tem GPP more successfully, and improve our understanding of the global carbon cycle  
 922 response to changing climatic conditions.

923 **Appendix A Data description**

Table A1: Description of forcing and model parameterization data

Abbreviation	Definition	Unit	Variable name in dataset/ remarks	Reference
$GPP_{EC}^a$	GPP derived from EC based net ecosystem exchange (NEE) using night-time partitioning* method	$\mu\text{molCO}_2 \cdot \text{m}^{-2} \cdot \text{s}^{-1}$	GPP_NT_VUT_USTAR50	Pastorello et al. (2020); Reichstein et al. (2005)
$\sigma_{NEE}$	Random uncertainty for NEE	$\mu\text{molCO}_2 \cdot \text{m}^{-2} \cdot \text{s}^{-1}$	NEE_VUT_USTAR50_ RANDUNC	Pastorello et al. (2020)
$LE$	Latent heat flux	$\text{W} \cdot \text{m}^{-2}$	LE_F_MDS	Pastorello et al. (2020)
$\sigma_{LE}$	Random uncertainty for latent heat flux	$\text{W} \cdot \text{m}^{-2}$	LE_RANDUNC	Pastorello et al. (2020)
$SW\_IN^b$	Incoming shortwave radiation	$\text{W} \cdot \text{m}^{-2}$	SW_IN_F	Pastorello et al. (2020)
$NETRAD^{b, c}$	Net radiation	$\text{W} \cdot \text{m}^{-2}$	NETRAD	Pastorello et al. (2020)
$SW\_IN\_POT$	Potential incoming shortwave radiation	$\text{W} \cdot \text{m}^{-2}$	SW_IN_POT	Pastorello et al. (2020)
$PPFD\_IN^a$	Incoming photosynthetic photon flux density	$\mu\text{mol photons} \cdot \text{m}^{-2} \cdot \text{s}^{-1}$	$PPFD\_IN$ gap-filled with $2.04 \times SW\_IN$	Pastorello et al. (2020); see Sect. 3.4.2 of Stocker et al. (2020) for the gap-filling equation

*Continued on next page*

Table A1 – *Continued from previous page*

Abbreviation	Definition	Unit	Variable name in dataset/ remarks	Reference
$T^b$	Air temperature	°C	TA_F_MDS	Pastorello et al. (2020)
$VPD^b$	Vapor pressure deficit	Pa	VPD_F_MDS	Pastorello et al. (2020)
$P^{b,d}$	Precipitation	mm · h <sup>-1</sup> or mm · d <sup>-1</sup>	P	Pastorello et al. (2020)
$CO_2$	Atmospheric CO <sub>2</sub> concentration dry air mole fractions from quasi-continuous measurements at Mauna Loa	ppm	co2_mlo_surface-insitu_1_ccgg-DailyData (interpolated linearly to hourly scale). The measurements from Mauna Loa were used for all sites as the CO <sub>2</sub> concentration measurements at EC sites are often noisy and discontinuous.	Thoning et al. (2021)
$elev$	Site elevation	m a.s.l.	Collected from literature	Bao et al. (2022)
$ET_{LE}^a$	Evapotranspiration derived from $LE$ flux	mm · h <sup>-1</sup> or mm · d <sup>-1</sup>	Calculated from $LE$ with a dependency on $T$	Henderson-Sellers (1984)
$\sigma_{ET}$	Random uncertainty for $ET_{LE}$	mm · h <sup>-1</sup> or mm · d <sup>-1</sup>	Calculated from $LE_{RANDUNC}$ with a dependency on $T$	Henderson-Sellers (1984)
$PET$	Potential evapotranspiration	mm · h <sup>-1</sup> or mm · d <sup>-1</sup>	Calculated from $T$ , $NETRAD$ and $elev$ using the method of Priestley and Taylor	Priestley and Taylor (1972)
$CI$	Cloudiness index	-	Calculated as $1 - \left(\frac{SW\_IN}{SW\_IN\_POT}\right)$	Bao et al. (2022); Fu and Rich (1999); Turner et al. (2006)
$WAI$	Water availability indicator	mm	Described in Sect. S1 of the supplement	Bao et al. (2022); Tramontana et al. (2016); Trautmann et al. (2018)
$W$	Soil water supply	mm · mm <sup>-1</sup>	calculated as $\frac{WAI}{AWC}$ ( $AWC$ is defined in Table 1)	Bao et al. (2022)
$NDVI$	Normalized difference vegetation index	-	Daily $NDVI$ from FluxnetEO v2 (MODIS) was linearly interpolated to hourly	Walther et al. (2022, 2023)

*Continued on next page*

Table A1 – *Continued from previous page*

Abbreviation	Definition	Unit	Variable name in dataset/ remarks	Reference
$fAPAR$	Fraction of incident photosynthetic photon flux that is absorbed by vegetation	-	Linear relationship between $NDVI$ and $fAPAR$ was assumed. $\begin{cases} NDVI, & \text{if } NDVI > 0 \\ 0, & \text{if } NDVI \leq 0 \end{cases}$	Bao et al. (2022); Myneni et al. (1997)
$QC^a$	Data quality flags	-	1.0 (good quality), 0.5 (medium quality), and 0.0 (bad quality) in the case of hourly data, which is the fraction of good quality measured or gap-filled data from two half-hours. In the case of daily, $QC$ can have any values between 0.0 and 1.0, which is a fraction representing the percentage of good quality measured or gap-filled data in a day. The daily data with $QC > 0.8$ was considered good.	Pastorello et al. (2020); Nelson et al. (2024)

a: For  $GPP_{EC}$ , the  $QC$  flags of  $NEE$ , and for  $ET_{LE}$  the  $QC$  flags of  $LE$  were used, as they were derived from the respective variables.  $QC$  flags of  $SW_{IN}$  were used to determine bad and medium quality data of  $PPFD_{IN}$ , which were replaced with a gap-filling procedure.

b: Bad, medium quality (value of  $QC$  is 0 and 0.5) data and gaps were filled with downscaled (Besnard et al., 2019) ERA5 (Hersbach et al., 2023) or ERA-Interim v2.0 data (Berrisford et al., 2011).

c: We have collected good quality  $SW_{IN}$  and  $NETRAD$  values from all the sites and fitted a linear regression model using the RANdom SAMple Consensus (RANSAC) algorithm (Fischler & Bolles, 1981) to determine the relation between them. The fitted equation ( $NETRAD = 0.7066 \times SW_{IN} - 0.1345$ ) was used to fill gaps in  $NETRAD$  using  $SW_{IN}$ . The gap-filling with regression was only applied for a few sites at hourly scale.

d: At hourly scale, the data gaps or bad quality data in  $P$  were filled by distributing the daily downscaled  $P$  (Besnard et al., 2019) from ERA-Interim v2.0 (Berrisford et al., 2011) for a certain day to the hourly timesteps, based on hourly  $P$  from gridded ERA5 data (Hersbach et al., 2023).

\*We preferred the night-time partitioning (Reichstein et al., 2005) over daytime partitioning (Lasslop et al., 2010) as only respiration is modelled in this case and GPP is derived as the difference between measured  $NEE$  and respiration. Whereas, in the daytime partitioning method, GPP is modelled as well and can have prediction errors due to uncertain model parameters.

## Appendix B Data screening for model parameterization

We used only good-quality data to calibrate model parameters. At hourly scale, we selected  $GPP_{EC}$  and  $ET_{LE}$  as good quality data when the values of their respective  $QC$  flag were 1 (Table A1). At the daily scale, we considered a  $GPP_{EC}$  and  $ET_{LE}$  data point as good when the value of the  $QC$  flag was greater than 0.8. We also removed any



929 data gaps from observed and simulated data,  $\sigma_{NEE}$ , and  $\sigma_{LE}$  (Table A1). There were  
 930 certain negative values in our  $GPP_{EC}$  data, as it was calculated using night-time based  
 931 partitioning method (Reichstein et al., 2005). In this case, if a negative  $GPP_{EC}$  value  
 932 occurred, when the  $SW_{IN}$  (Table A1) is zero i.e., during night hours, we replaced those  
 933 data points with 0 and used them in the cost function. If the negative  $GPP_{EC}$  occurred  
 934 during day hours, we excluded them.

### 935 Appendix C Data screening for evaluation of model performance

936 The good quality data at an hourly scale were selected using the same criteria de-  
 937 scribed in Appendix B. The data screening at a daily scale was also similar to Appendix  
 938 B, when the LUE model was parameterized using daily data. For all other cases, we as-  
 939 signed a flag (0 = not considered, 1 = considered) to identify which data points were con-  
 940 sidered during model parameterization. We aggregated this flag to daily, weekly, monthly,  
 941 and annual scales by taking averages. Then this flag indicated the fraction of good qual-  
 942 ity data used to calculate a data point in a certain temporal resolution. We only used  
 943 data points at certain temporal resolutions which were calculated using more than 50%  
 944 (flag value  $> 0.5$ ) good quality data points from hourly/daily resolution. We calculated  
 945 monthly, and annual model performance metrics for a certain site if at least three data  
 946 points were present. We couldn't calculate annual metrics for 76 and 85 sites due to low  
 947 numbers of good quality site-years when the annual data was aggregated from hourly,  
 948 and daily data, respectively. The monthly metrics were not calculated for the three sites  
 949 due to the same reason when they were aggregated from daily data.

### 950 Appendix D Median values of model performance

951 The median values of the model performance metric, i.e., NNSE which are plot-  
 952 ted in Fig. 2 are summarized in Table D1.

**Table D1.** Median NNSE obtained at each modelling experiment at hourly/daily scale and annual aggregations

Temporal scale/ aggregation	Models	parameterization strategies				
		per site-year	per site using Cost <sup>IAV</sup>	per site	per PFT	global
<b>Hourly/ daily scale</b>	P <sub>hr</sub> <sup>W</sup>	0.827	0.799	0.805	0.738	0.712
	P <sub>hr</sub>	0.478	0.470	0.469	0.461	0.490
	BaO <sub>hr</sub>	0.853	0.816	0.836	0.757	0.693
	BaO <sub>dd</sub>	0.830	0.774	0.790	0.658	0.634
<b>annual aggregation</b>	P <sub>hr</sub> <sup>W</sup>	0.543	0.405	0.373	0.201	0.143
	P <sub>hr</sub>	0.018	0.019	0.018	0.018	0.019
	BaO <sub>hr</sub>	0.661	0.471	0.440	0.190	0.105
	BaO <sub>dd</sub>	0.669	0.489	0.482	0.238	0.143

### 953 Open Research Section

954 The codes that were used to perform all the necessary analyses and plot all the fig-  
 955 ures in this study are available at De (2024). The data from eddy covariance sites are

956 available from FLUXNET (Pastorello et al., 2020; FLUXNET.org, 2024a). The Fluxnet-  
 957 tEO MODIS version 2 dataset is available at Walther et al. (2022, 2023). ERA5 dataset  
 958 and ERA-Interim v2.0 data are available from Hersbach et al. (2023) and Berrisford et  
 959 al. (2011), respectively. Atmospheric CO<sub>2</sub> measurements at Mauna Loa observatory are  
 960 available at Thoning et al. (2021).

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