

1 **Valorization of deinking sludge as a substrate for lignocellulolytic**  
2 **enzymes production by *Pleurotus ostreatus***

3 Maša Vodovnik<sup>a\*</sup>, Katja Vrabec<sup>a</sup>, Patrick Hellwig<sup>b</sup>, Dirk Benndorf<sup>b,g</sup>, Mija Sežun<sup>c</sup>, Andrej  
4 Gregori<sup>d</sup>, Lalitha D. Gottomukkala<sup>e</sup>, Robin C. Anderson<sup>f</sup>, Udo Reichl<sup>b,g</sup>

5<sup>a</sup> Chair of Microbiology and Microbial Biotechnology, Biotechnical Faculty, University of Ljubljana,  
6Domžale, Slovenia

7<sup>b</sup> Institute of Process Engineering, Otto von Guericke University Magdeburg, Magdeburg, Germany

8<sup>c</sup> Pulp and Paper Institute, Ljubljana, Slovenia

9<sup>d</sup> MycoMedica d.o.o, Kranjska Gora, Slovenia

10<sup>e</sup> Celignis Limited, University of Limerick, Limerick, Ireland

11<sup>f</sup> United States Department of Agriculture, Southern Plains Agricultural Research Center, Food and  
12Feed Safety Research Unit, College Station, Texas

13<sup>g</sup> Max Planck Institute for Dynamics of Complex Technical Systems, Germany

14

15 \*Corresponding author:

16 [masa.vodovnik@bf.uni-lj.si](mailto:masa.vodovnik@bf.uni-lj.si)

17

18

19

20

21

22

23

## 24Abstract

25Disposal of waste sludges produced in large amounts in the pulp and paper industry imposes  
26significant environmental and economical problems. One strategy to address these issues  
27involves revalorization of paper mill sludges by their application as substrates for microbial  
28production of biotechnologically relevant enzymes. The application of lignocellulolytic  
29enzymes in paper, textile and bioenergy industries is encouraged in order to decrease  
30chemicals and energy consumptions. In the following work, deinking sludge was assessed as a  
31substrate for production of lignocellulases. Based on the results of growth and activity  
32screenings, *Pleurotus ostreatus* PLAB was chosen as the most promising candidate among 30  
33tested strains and its secretome was further studied by quantitative enzyme assays and mass  
34spectrometry. While endoglucanase and xylanase activities detected in *P. ostreatus* secretome  
35produced on deinking sludge were similar to activities of cultures grown on other  
36lignocellulosic substrates, average laccase activity was significantly higher (46 000 U / kg<sub>DIS</sub>).  
37Mass spectrometry identification of the most prominent proteins in the secretome of the target  
38strain confirmed that significant amounts of different lignin-modifying oxidases were  
39produced on this substrate despite its low lignin content, indicating the presence of other  
40inducible compounds. The findings of this study suggest deinking sludge may represent a  
41good substrate for fungal production of the aforementioned enzymes with broad  
42biotechnological applications, including bioremediation, paper and bioenergy industries.

43**Key words:** paper mill waste / deinking sludge / fungi / lignocellulolytic enzymes / solid-state  
44fermentation

45

46

**47Abbreviations:** DIS – deinking sludge, SF – submerged fermentation, SSF – solid state  
48fermentation, TOC – total organic carbon, TON – total organic nitrogen, WRF – white rot  
49fungi, PDA – potato dextrose agar, SEEs – secreted enzyme extracts, CMC – carboxymethyl  
50cellulose, BX – birchwood xylan, EG – endoglucanase, EX – endoxylanase, ABTS - 2,2'-  
51azino-bis (3-ethylbenzothiazoline-6-sulphonic acid), PDA – potato dextrose agar, SDS PAGE  
52– sodium dodecyl sulphate polyacrylamide gel electrophoresis, TCA – trichloroacetic acid,  
53LC MS/MS – liquid chromatography coupled to tandem mass spectrometry

54

## 551. Introduction

56 Eleven million tonnes of waste are produced yearly by the European pulp and paper industry  
57 (Monte et al., 2009). Approximately 70% of these originate from manufacturing tissue paper  
58 from recovered fibre, leading to the generation of considerable amounts of deinking sludge  
59 (150 kg dry solids/t paper manufactured), which must be properly managed to avoid negative  
60 effects on the environment (Deviatkin et al., 2016). Deinking sludge (DIS) exists as a mixture  
61 of short cellulosic fibers and inorganic fillers, such as calcium carbonate and china clay, and  
62 residual chemicals dissolved in water (Likon and Saarela, 2012). DIS originating from printed  
63 recycle mills is high in ash content compared to sludge originating from corrugated recycle  
64 mills and virgin pulp mills (Boshoff et al., 2016).

65 Traditional methods of DIS management include landfilling, landspreading, composting,  
66 incineration and pyrolysis, utilisation as construction material and landfill capping material  
67 (Likon and Trebše, 2012). However, due to high moisture content some of these recovery  
68 methods, such as incineration and pyrolysis, are expensive for large amounts of sludge while  
69 the environmental impact of others is questionable due to the possibility of hazardous  
70 substances leaking into the environment. As a result, numerous possibilities for biological  
71 valorisation of paper sludge waste, including its fermentation and anaerobic digestion are  
72 currently being explored (Gottomukkala et al., 2016). These aim for efficient microbial  
73 transformation of cellulose waste into bioethanol (Boshoff et al., 2016), biomethane (Mohan  
74 et al., 2016a), biohydrogen or other value added chemicals (Liguori and Faraco, 2016).

75 The main advantage of exploiting paper sludges as sources of cellulose-derived energy and  
76 chemicals in comparison to other lignocellulose substrates is their amenability, which is  
77 associated with an extensive pulping process that removes the majority of the lignin and  
78 exposes cellulose fibers to enzymes. This results in substantial cost savings on energy for  
79 substrate pretreatment in comparison to other lignocellulose fuel production technologies

80(Gottumukkala et al., 2016). Studies on direct production of bioethanol from paper sludge  
81have shown that tissue printed recycle sludge (a type of DIS) resulted in significantly lower  
82ethanol yields when compared to corrugated recycle and virgin pulp mill sludges (Williams  
832017), so other possible ways for valorization of this type of substrate need to be explored.  
84 One of the major limiting factors for bioconversion of cellulosic waste such as paper sludge  
85to valuable products is the cost of the enzymes, as they are commercially produced using high  
86cost feedstocks. Reducing the enzyme dosage per gram cellulose/feedstock has been a major  
87research area for the past few years (Robus et al., 2016). Efficient transformation of cellulosic  
88feedstock to value added products requires a mix of synergistically acting enzymes  
89(CAZYmes) that are able to work at low dosages. White-rot fungi (WRF) are known to  
90produce significant amounts of powerful extracellular oxidative and hydrolytic enzymes that  
91degrade lignin and cellulose biopolymers (Manavalan et al., 2015). Major functional groups  
92of glycoside hydrolases (cellulases and hemicellulases) produced by WRF involve  
93endoglucanases (EG; EC 3.2.1.4), cellobiohydrolases (EC 3.2.1.91, EC 3.2.1.176), beta-  
94glucosidases (3.2.1.21), endoxylanases (EX; EC 3.2.1.8), beta-xylosidases (EC 3.2.1.37) and  
95alpha-glucuronidases (E.C. 3.2.1.131). On the other hand, lignin degradation enzyme systems  
96are based on oxidative enzymes, such as lignin peroxidases (EC 1.11.1.14), manganese  
97peroxidases (EC 1.11.1.13), versatile peroxidases (EC1.11.1.16), and laccases (EC 1.10.3.2)  
98(Manavalan et al., 2015). Recently, oxidative enzymes, namely lytic polysaccharide  
99monooxygenases, have also been shown to play an important role in the degradation of  
100cellulose (Garajova et al., 2016). The composition of the enzyme mixtures produced on  
101different substrates (ratios between different types of enzymes) reflects substrate composition  
102(Elisashvili et al., 2008).

103The applications of fungal enzymes in paper industry involve biobleaching of pulp, pulp de-  
104inking, degradation of dissolved and suspended organic compounds in concentrated effluents

105of mills and enhanced fibrillation. Enzymes usage is encouraged in paper industry in order to  
106reduce the use of chemicals and energy consumption as well as to improve quality of the  
107products (Bajpai, 2014). A desirable side-effect of processing paper mill waste by WRF  
108enzymes is non-specific degradation of organic environmental pollutants, commonly present  
109in these substrates (Kúes, 2015).

110Most industrial enzymes currently produced by fungi are products of submerged fermentation  
111(SF) which uses significant amounts of water and generates large quantities of liquid waste  
112stream during the filtration process. Major cost contributor in commercial cellulase  
113production is cost of the feedstock which is mostly glucose and accounts for 50% of the total  
114process cost (Humbird et al., 2011). However, solid state fermentation (SSF) has emerged as  
115an economically attractive alternative for the *in situ* production of lignocellulolytic enzymes  
116due to lower energy consumption, direct use of low-cost lignocellulosic wastes as substrates,  
117reduction in cost of dewatering in downstream processing, higher concentrations of enzymes  
118and lower demand for the sterility of the equipments (Yoon et al., 2014). At present, the  
119production of lignocellulolytic enzymes by SSF on several agricultural wastes has been  
120reported, such as coffee pulp (Velazquez-Cedeno et al., 2002), spend brewery grains (Gregori  
121et al., 2008), straw and fruit peels (Kurt and Buyukalaca, 2010), rice straw (Khalil et al.,  
1222011) and tomato pomace (Iandolo and Piscitelli, 2011). Conversely, there are few reports on  
123enzyme production from paper sludge and no report specifically on fungal enzymes  
124production from deinking paper sludge.

125The aim of the following work was to untap the potential of DIS as a low-cost substrate for  
126fungal production of lignocellulolytic enzymes via SSF. Fungal strains with the ability to  
127transform the untreated DIS to a mixture of industrially relevant lignocellulases were  
128identified. Furthermore, the composition, production dynamics and performance of the  
129extracted enzymes were studied in detail in the most promising of the tested strains (*P.*

130 *ostreatus PLAB*). The advantages of lignocellulolytic enzymes production on DIS and their  
131 applications in different industries is also discussed from a perspective of cleaner production.

132

## 1332. **Materials and methods**

### 1342.1. **Substrate characterization**

135 DIS was obtained from a paper mill in Vipav, Krško, Slovenia. After the substrate was sieved  
136 through a mesh with 0.9 cm openings, its pH and moisture content were measured (Kern  
137 MLS-D). For the determination of cellulose, hemicellulose and lignin content of the substrate,  
138 10 g of air-dried DIS was first mixed with 200 ml of 6 M HCl and then incubated for 1h to  
139 dissolve the inorganic fraction. Following filtration and drying, the cellulose content of DIS  
140 was determined according to the method described in Kurschner and Hoffer (1993),  
141 hemicellulose content as described in Wise and Karl (1962) and lignin content according to  
142 the method as described in Fukushima et al., 2015. Dry content was measured according to  
143 European standard SIST EN 14346:2007, total organic carbon (TOC) according to standard  
144 SIST EN 13137:2002, total organic nitrogen (TON) according to persulfate digestion method  
145 10071 (Test'N Tube, Hach Lange). Heavy metals were determined according to the following  
146 standards: DIN 38406-E7-2:1991 (Cu), SIST ISO 8288:1996 (Zn), DIN 38406-E11-2:1991  
147 (Ni), SIST EN ISO 5961:1996 (Cd), SIST ISO 5666:2000 (Hg), SIST ISO 9174:1999 (Cr),  
148 DIN 38406-E6-2:1998 (Pb).

149

150

151

152

### 1532.2. **Screening of fungal strains based on growth efficiency on DIS**

154Thirty fungal strains of different origin obtained from culture collection Mycomedica d.o.o.  
155(Suppl. 1) were first screened for their growth on agar plates incorporating the target substrate  
156(DIS agar). The strains were stored as slant test tubes on Potato Dextrose Agar (PDA, Difco,  
157USA) at 3 °C. Before each experiment, mycelium from each fungal strain was transferred to  
158PDA agar and incubated in the darkness at 23 °C until overgrowing the entire surface. A 9 mm  
159diameter disc was cut from the resultant mycelium and used to inoculate a DIS agar plate. DIS  
160agar, prepared by mixing bacteriological agar (Sigma), deinking sludge and distilled water in  
161ratio 1:10:100, was autoclaved for 20 min at 121 °C, 1.1 bar. Additional sets of media were  
162prepared similarly except without autoclaving in order to test the growth on non-pretreated  
163substrate (energetically favourable option). All strains were inoculated in triplicates and their  
164growth was observed for 50 days and classified in different categories according to its  
165efficiency. Ten strains with the highest average growth rate on DIS agar were chosen for  
166further enzyme activity screenings.

167

### 1682.3. Enzyme activity screening

169Based on the highest growth rates achieved on DIS agar, 10 strains were selected for  
170additional semi-quantitative screening for endoglucanase, xylanase and laccase activities.  
171Secreted enzyme extracts (SEEs) were prepared by first homogenizing the overgrown DIS  
172agar by Ultraturax disperser (Ika) at 1000 rpm for 5 min. Aliquotes (20 ml) of homogenate  
173were mixed with equal amounts of 0.5 M Na-phosphate buffer pH 7 and shaken on ice at 150  
174rpm for 3 h. Following centrifugation at 18500 x g, the clear supernatants (representing SEEs)  
175were transferred to sterile Falcon tube and stored at -20 °C until further use. Aliquots of 10 µL  
176of SEEs from different fungal strains were applied in the hollow areas of equal size (0.5 mm  
177diameter) wells cut into 0.6% (w/v) agarose plates incorporating target substrates. The  
178following substrates were used: 0.2% (w/v) carboxymethyl cellulose (CMC) when screening

179for endoglucanase activity, 0.4% (w/v) birchwood xylan (BX), when screening for xylanase  
180activity and 0.4% (v/v) guaiacol when screening for laccase activity. SEEs denatured by  
181heating at 100 °C for 15 min were used as negative controls. The plates were incubated at  
18230 °C for 4 h. Guaiacol plates were assessed directly by measuring radius of brown areas that  
183developed around wells with applied SEEs due to laccase activity. Plates incorporating CMC  
184and BX were stained by Kongo red for 30 min and subsequently washed by 1 M NaCl  
185(Teather and Wood, 1982). The radii of clear zones indicating substrate degradation were  
186measured and compared in order to assay xylanase and endoglucanase activities in SEEs from  
187different fungal strains. Based on growth and enzyme screening, *P. ostraetus* PLAB was  
188selected for further investigation.

189

#### 1902.4. Enzymes production and extraction from deinking paper sludge (DIS)

191*P. ostraetus* PLAB mycelia were first inoculated to PDA media in petri dishes as described in  
192the section 2.2. After suspending in distilled water, 5 ml of mycelium was used to inoculate  
193200 g of each sterilized DIS medium (prepared by sieving DIS through the mesh and adjusting  
194its moisture content to 65% (w/w) in 1 L glass jars. Uninoculated DIS medium was used as a  
195negative control. Jars were incubated at 23 °C until mycelial growth was observed. For  
196enzyme activity quantification, cultures were incubated for 51 days and collected in 7 day  
197intervals (first sampling took place after mycelium started to grow in all of the samples). Four  
198cultures were collected at each time point and frozen for further analysis. Following  
199homogenization, selected buffers were added to each homogenate at the ratio 1:1. Based on  
200literature and preliminary experiments, 0.5 M acetate buffer pH 4 was used for laccase, 0.5 M  
201acetate buffer pH 5 for endoglucanase (EG) and 0.5 M citrate buffer pH 6 for xylanase  
202extraction. Extraction was performed by incubating the samples on ice for 3 h with  
203intermittent mixing by Vortex (every 30 min for 30 s).

204 Following centrifugation at 18500 x g, clear supernatant (representing SEEs) was separated,  
205 aliquoted and stored at -20 °C until further processing.

206

### 207 **2.5. Spectrophotometric assays for quantitative determination of total protein,** 208 **endoglucanase and xylanase activities in SEEs**

209 Cellulase (endoglucanase) and xylanase activities in SEEs were measured via quantitative  
210 determination of released reducing sugars by para-hydroxybenzoic acid (PAHBAH) assay  
211 (Lever, 1977). Suspensions of 1% beechwood xylan (BX) in 0.5 M citrate buffer pH 6 (for  
212 xylanase activity) or 1% carboxymethyl cellulose (CMC) in 0.5 M acetate buffer pH 5  
213 (endoglucanase activity) were used as substrates. Twenty µl of enzyme extracts were added to  
214 230 µl of substrate suspensions and incubated in microcentrifuge tubes at 30 °C for 150 min.  
215 Reactions were stopped by adding 30 µl of 15% trichloroacetic acid (TCA). In control  
216 samples, enzyme extracts were added after the incubation, just before enzyme inactivation.  
217 100 µl of each incubated sample was added to 5 ml of reaction mixture in test tubes and  
218 heated in boiling water for 10 min. Absorbance of the mixture was measured at 420 nm. Total  
219 protein concentrations in SEEs were quantified by Lowry assay, using bovine serum albumin  
220 (BSA) as a standard (Lowry et al., 1951). Two technical replicates were performed for each  
221 sample in order to assess method variability.

222

### 223 **2.6. Spectrophotometric assay for quantitative determination of laccase activity in SEEs**

224 A modified assay described in Childs and Bardsley (1975) was used to measure laccase  
225 activities in SEEs. Reaction mixtures (prepared in duplicate) consisted of 100 µl 5 mM 2,2'-  
226 azino-bis (3-ethylbenzothiazoline-6-sulphonic acid) (ABTS, Calbiochem), 895 µl 0.1 M  
227 acetate buffer pH 4.5 and 5 µl enzyme extract. Cooled ABTS was added to the sample after  
228 heating reaction mixture to 30 °C and absorbance was measured at 420 nm continuously.

229 Enzyme extracts previously inactivated by heating at 100 °C for 20 min were used as negative  
230 controls.

231

### 232 2.7. SDS-PAGE and mass spectrometry

233 Four 30-days old cultures were homogenized and SEEs were prepared by extraction of  
234 proteins from 20 g of each homogenate in equal amounts of phosphate buffer (pH 8) with the  
235 addition of protease inhibitors (cOmplete ULTRA Tablets, Sigma-Aldrich) and freeze-dried  
236 for transport at ambient temperature. Before the analysis, the samples were dissolved in 750  
237  $\mu$ l deionized water and precipitated with the same volume of 20% TCA at 4 °C for 1 h. The  
238 precipitate was collected by centrifugation (16400 x g, 10 min, 4 °C), subsequently washed  
239 twice with 2.25 ml of acetone (15 min, -20 °C) and 2.25 ml of ethanol (15 min, -20 °C). After  
240 air-drying the final precipitates were dissolved in 200  $\mu$ l of urea buffer. The protein  
241 concentration determination was carried out with Amidoblack (Schweickl et al., 1989) using  
242 BSA as standard. SDS-PAGE was performed according to Laemmli (1970). After separation,  
243 the bands were cut from the gel and digested with trypsin as described previously by Heyer et  
244 al. (2013). The peptides were extracted, dried and resuspended in 11  $\mu$ L of chromatographic  
245 mobile phase A (98% LC-MS water, 2% acetonitrile, 0.05% trifluoroacetic acid). Five  $\mu$ L of  
246 sample were injected and separated by UltiMate® 3000 nano splitted reversed phase nano  
247 HPLC (Thermo Fisher Scientific, Dreieich, Germany) equipped with a reversed phase trap  
248 column (nano trap cartridge, 300  $\mu$ m i.d. x 5 mm, packed with Acclaim PepMap100 C18, 5  
249  $\mu$ m, 100 Å, nanoViper, Bremen, Germany) and a reversed phase separation column (Acclaim  
250 PepMap RSLC, C18, 2  $\mu$ m, 75  $\mu$ m, 15 cm, Bremen, Germany).

251 The gradient was 4% to 55% mobile phase B (80% acetonitrile, 20% LC-MS water, 0.1%  
252 formic acid) over 120 min at a flow rate of 0.3  $\mu$ L min<sup>-1</sup>. The amaZon™ ETD Ion Trap mass  
253 spectrometer (Bruker Daltonik GmbH, Bremen, Germany) was equipped with a Captive

254Spray (using acetonitrile) nanoBooster™ Module and operated in positive ion mode. MS/MS  
255data acquisition was performed using collision induced dissociation (CID) with 5 precursor  
256ions per measurement and an active exclusion of the same precursor after 2 spectra for 9  
257seconds. DataAnalysis 2.0 (Bruker Daltonik) and ProteinScape 3.1.3 461 software (Bruker  
258Daltonik) were used for automatic deconvolution and analysis as well as database search  
259using Mascot™ 2.5 (Matrix Science, London, UK) in NCBI database (version 2014/01/22).  
260One missed cleavage was allowed. Other parameters were chosen as follows: monoisotopic  
261mass, variable carbamidomethyl (C) and oxidation (M) modification, 0.3 Da precursor and  
262MS/MS tolerance, one <sup>13</sup>C and peptide charge of +2/+3.

263

#### 2642.8. Zymograms

265Separated active glycoside hydrolases in SEEs were detected by modified SDS-PAGE  
266(Laemmli, 1970). Separation gels (10%) were modified by adding 0.2% (w/v) CMC (for  
267detection of endoglucanases) or 0.4% (w/v) water soluble fraction of 0.4% (w/v) BX (for  
268detection of xylanases). Enzyme extracts were mixed with equal amounts of loading buffer,  
269heated for 6 min at 85 °C and applied on 4% (w/v) stacking gel. For detection of enzymes  
270with laccase activity, native samples were applied on the gels. Glycoside hydrolases were  
271renaturated and stained as described previously (Vodovnik and Marinšek Logar, 2012).  
272Separated enzymes exhibiting laccase activity were detected by applying ABTS solution  
273(Calbiochem, 194430) directly over the gel after its equilibration in 0.1 acetate buffer pH 4 as  
274described by Castanera et al. (2012). The properties of the detected enzymes were compared  
275to the information on putative *P. ostreatus* enzymes available in the NCBI protein (Wheeler et  
276al., 2002) and BRENDA databases (Schomburg et al., 2004). Where absent, putative  
277conserved domains were detected by Interproscan (Quevillon et al., 2005) while other

278 properties of interest (Mw and pI) were calculated using ProtParam tool (Gasteiger et al.,  
279 2005).

280

## 281 2.9. Data processing and statistical analysis

282 Results are presented as the mean  $\pm$  standard deviation (SD) of at least 3 replicates. The data  
283 presentation and analysis were carried out using R version 3.1.3 and Microsoft Excell 2013.  
284 Statistically significant differences in enzyme activities between different fungi were  
285 evaluated by one-way ANOVA, while growth-point-related activities analysis was done by  
286 repeated measures ANOVA (paired samples Student's t-test). A significant difference was  
287 considered if  $p < 0.05$  applying Tukey multiple-comparisons HSD.

288

## 289 3. Results and discussion

### 290 3.1. Substrate characterization

291 Commercial cellulases are produced using recombinant fungal strains to yield enzyme titres  
292 of close to 100 g/L with superior activity. Inducible carbon sources such as purified cellulose,  
293 lactose and sophorose are used for high-level cellulase production. Feedstock cost is an  
294 important consideration for commercial enzyme production as this contributes more than  
295 50% of the total cost of the final enzyme (Ellila et al, 2017). Other than cost factor, there are  
296 also environmental concerns that impact commercial enzyme production as bioprocesses  
297 using genetically modified organisms, excessive amounts of nutrients used for process  
298 intensification and use of large quantities of fresh water can adversely affect regulatory costs  
299 and carbon emissions (Junker, 2010). Utilisation of waste streams rich in cellulose such as  
300 paper sludge as feedstocks for enzyme production will not only reduce the feedstock cost for  
301 enzyme production but also support cleaner production by reducing the amounts of waste

302 disposal in comparison to high amounts of nutrient rich liquid and solid waste generated by  
303 the bioprocess industry that mainly uses pure substrate and nutrient source.

304 Deinking paper mill sludge, a major waste stream from the pulp and paper industry, was  
305 selected for this study, as it is mainly composed of short cellulose fibers. It also includes  
306 fines, inks and other additives that can influence the growth of the microorganisms and  
307 enzyme production. It has been reported that even among mills using similar processes and  
308 feedstocks, the sludge composition may vary due to different operating conditions (Williams,  
309 2017). So DIS used in our study was characterized for its chemical composition.

310 DIS used in our study has a moisture content of 39% and cellulose content of 34.2%.  
311 Hemicellulose and lignin contents are 6.2% and 3.7% respectively, which were lower than  
312 stated in previous reports (Gottomukkala et al., 2016). The remaining is mostly ash and heavy  
313 metals (Table 1). Cellulose is the major carbon source contributor for fungal growth and also  
314 an inducer of cellulases. Hemicellulose and lignin fractions on the other hand play an  
315 important role in production of plant biomass-degrading enzyme cocktail including xylanases  
316 and ligninases (Fernández-Fueyo et al., 2016).

317 The high moisture content of DIS is due to the water holding capacity of cellulose fibres in  
318 the sludge which makes it unsuitable feedstock for thermochemical conversion process.

319 Breaking down of cellulose fibres in DIS by enzymes has been shown to not only add value to  
320 the feedstock, but also significantly reduce the water content and bulk of the sludge that has to  
321 be discarded (Steffen et al. 2017a). Williams (2017) has studied three different types of paper  
322 sludge for ethanol production and has reported that usage of enzymes and break down of  
323 cellulose has significantly reduced the water holding capacity and increased the bulk density  
324 of all the three types of paper sludge. This clearly indicates the advantage of using DIS as  
325 feedstock for enzyme production as cellulase producing fungi are able to use cellulose present  
326 in the DIS as a substrate and reduce the bulk of sludge to be disposed.

327 Other important parameters anticipated to influence the growth and enzyme production  
328 (activity) of microorganisms on DIS include pH, C/N ratio as well as concentrations of  
329 bioavailable nutrients, ash and heavy metals (Table 1). The DIS used in this study has a dry  
330 weight of 61.35% with 34.3% total organic carbon (TOC) and 1.48% nitrogen, indicating C/N  
331 ratio of 23.3, which falls within the optimal range for maintaining optimal fungal growth and  
332 cellulose utilization (Wan and Li, 2012). This indicates that there is no additional requirement  
333 for a nitrogen source for fungal enzyme production using DIS as a substrate. This is  
334 advantageous in comparison to other cellulose-rich substrates, where nitrogen content and  
335 availability are among the main limitations affecting microbial degradation and enzyme  
336 production (Huang et al. 2017). Excess amounts of salts and nutrients used for process  
337 intensification and to maintain high cell mass not only increase the cost of the process, but can  
338 also be left unutilised or accumulated in the cells and are of environmental concern (Junker  
339 2010).

340 Low concentrations of certain heavy metals are essential for transcription activation and  
341 activity of certain oxidative enzymes, but excessive amounts may have inhibitory or toxic  
342 effects on some production strains (Lorenzo et al., 2006; Bhattacharya et al., 2013). The  
343 concentrations of heavy metals in DIS (Table 1) were lower than reported for some other  
344 paper mill sludges (Beauchamp et al., 2002; Abdullah et al., 2015) and did not exceed the  
345 concentrations typically found in non-contaminated soils (Tóth et al., 2016) which indicates  
346 that the risk of heavy metal inhibition on enzymes production is negligible with DIS as a  
347 substrate.

348

349 Appreciable amounts of cellulose, low concentration of heavy metals, desirable C/N ratio  
350 (thus negating the need for supplementing additional nutrient requirements) indicate that DIS  
351 is a suitable feedstock for lignocellulolytic enzymes production. Enzyme production with DIS

352as a substrate can be further improved by reducing the ash concentration and increasing the  
353cellulose concentration in the substrate. Robus et al. (2016) has shown that washing the paper  
354sludge can reduce the ash content in the paper sludge significantly. On the other hand,  
355washing utilises large quantities of fresh water and hence is not considered suitable for  
356sustainable and cleaner production. So, the fungal strains capable of growing under high pH,  
357high ash conditions and minimum nutrients should be targeted for enzymes production.

358

359

### 3603.2. Screening growth of selected strains on DIS

361Results from our initial screening, conducted to select fungal strains exhibiting the best  
362growth on DIS, are presented in Fig. 2. Conceptually, these strains should be those producing  
363enzyme mixtures that are most effective at breaking down the non-pretreated DIS with no or  
364minimal supplementation of nutrients. Moreover, the strains exhibiting the best growth on  
365DIS should be resistant to high pH of the substrate, chlorinated organic compounds or inks  
366that are commonly found in this type of substrate (Monte et al., 2009). The growth of the 30  
367selected fungal strains was monitored for 50 days and classified in different categories  
368according to its ability / efficiency of growth on autoclaved (A-DIS) and non-pretreated (NA-  
369DIS) medium (Fig. 2). While the autoclaved substrate supported the growth of 15 fungal  
370strains, only 8 of these strains (all white rots), were able to grow on NA-DIS. However, all  
371except *P. eryngii* grew significantly better on the autoclaved substrate. This may be due to  
372partial breakdown of cellulose fibres in DIS during hot steam treatment that may increase the  
373bioavailability of nutrients, supporting larger spectrum of the strains (Zheng et al., 2014). In  
374addition, the indigenous microbial community colonizing the NA-DIS may have exerted a  
375competitive exclusion effect that precluded growth of some of the less competitive fungal  
376strains. Under such conditions, it is reasonable to conclude that the 8 strains able to grow on

377the NA-DIS were able to successfully compete against the indigenous community for essential  
378nutrients and resist the conditions established in the substrate.

379Pretreatment is considered as one of the major energy and cost intensive unit operation in  
380lignocellulose biomass conversion and pressure vessels for sterilisation are one of the major  
381capital investment cost in bioprocess industries (Brodeur et al., 2011). Steam sterilisation  
382requires fresh water to generate steam and to cool the waste steam. The ability of the fungal  
383strains to outgrow indigenous microbes colonizing DIS and to utilise untreated DIS is  
384therefore a significant advantage in terms of energy and water usage. The majority of the  
385strains that were able to grow on NA-DIS belonged to the genus *Pleurotus* which has  
386previously been reported to have high tolerance to different types of environmental pollutants,  
387such as heavy metals and polycyclic aromatic compounds (Kapahi and Sachdeva, 2017). This  
388has been associated with the ability to accumulate and immobilize (chelate) high levels of  
389heavy metals (Yang et al., 2017). In addition, *P. ostreatus* and some white-rots producing  
390ligninolytic enzymes have been shown to mineralize several persistent organic pollutants  
391(POPs) to CO<sub>2</sub> (Bezalel et al. 1996) and have therefore already been used for bioremediation  
392of contaminated solid organic wastes (such as olive mill wastes) and soils (Kües, 2015). The  
393potential and capacity of the strains selected in this work for bioremediation of DIS from  
394different paper mills and pollutant loadings should further be studied.

395

### 3963.3. Screening enzyme activities of selected strains on DIS

397Extracted secretomes of 10 strains exhibiting the most growth on A-DIS and ability to grow  
398on NA-DIS were screened for target enzyme activities by plate assays. The latter reflected the  
399inducibility of the genes encoding (ligno)cellulolytic enzymes by the elements in the  
400substrate, their translational efficiency as well as the stability and activity of the extract  
401mixtures in given conditions (Lee, 2012). The SEEs of all the strains exhibited similar

402xylanolytic activity, while endoglucanase activity was significantly higher in the four  
403*Pleurotus* strains (PLAB, PP3, PLO4, PLOZ), *C. purpureum* (CHP4), *H. ulmarius* (HYUL1),  
404*P. ciliates* (PC1) and *G. lucidum* (GAL5) than in the remaining two tested strains (Fig. 3).  
405Laccase activity was only detected in seven of the tested strains and was most pronounced in  
406two *Pleurotus* strains (PLAB, PP3) and *H. ulmarius* (HYUL1).  
407As extracellular enzymes were separated from mycelia before testing, the enzymes  
408performance must reflect their intrinsic stability (not associated with mycelial defense  
409mechanisms, such as accumulation of toxic compounds or buffering) in the applied  
410conditions, such as high pH and potentially inhibitory compounds.  
411Enzyme performance at high pH values (8.6 in case of our substrate), typically associated  
412with high amounts of mineral fillers present in DIS, is advantageous for many applications in  
413the paper industry, as most of the current commercial cellulases perform best in weakly acidic  
414conditions (pH 5-7)(Steffen et al, 2017a). Other factors that may affect stability and activity of  
415enzymes produced on DIS are remaining organic pollutants or heavy metals (originating from  
416inks). However, lignin-degrading oxidoreductases from different white rots have  
417demonstrated not only resistance, but also the ability to degrade or transform broad range of  
418organic pollutants (including pharmaceuticals) to less toxic species (Naghdi et al., 2018).  
419Consequently, they have already been used to purify contaminated water or solid materials  
420prior to release into an environment, as well as for bioremediation within different  
421environments (Kúes, 2015).  
422 On the other hand, variable effects of different heavy metals on lignocellulolytic enzyme  
423activities and stability were reported. While Fe<sup>2+</sup> and Cu<sup>2+</sup> mainly exerted inhibitory effects on  
424fungal endoglucanases, exoglucanases, and β-glucosidases, the effects of other divalent ions  
425seems to be inconsistent among the enzymes secreted by different microorganisms (de Cassia  
426Pereira et al., 2017). In contrary, most of the studies report positive effects of some heavy

427metals (in particular  $\text{Cu}^{2+}$  and  $\text{Cd}^{2+}$ ) on oxidative enzymes, especially laccases (Yin et al.,  
4282017).

429

430

#### 4313.4. Major proteins identified in *P. ostreatus* DIS-associated secretome

432During their evolution, microorganisms have optimized the expression of genes encoding  
433lignocellulolytic enzymes by sensing and responding to the composition of the available  
434substrate in most efficient ways (Fernández-Fueyo et al., 2016). Studying the composition of  
435enzyme consortia secreted on different target substrates is therefore useful for understanding  
436the mechanisms of their degradation and for the successful design and production of enzyme  
437cocktails optimized for specific industrial processes (Vodovnik and Marinšek Logar, 2010).

438Our study revealed six most abundant proteins in the secretome of *P. ostreatus* PLAB likely  
439involved in the degradation of DIS (Table 2, Suppl. 4, 5). Among these were two laccase  
440isoenzymes with the theoretical Mw of 57.4 and 56.7. Two enzymes with ABTS-oxidasing  
441activity were also confirmed by zymogram approach (Suppl. 6C) – however, their activity  
442was only detected following native separation conditions and was irreversibly lost if the  
443samples were pretreated by SDS. SDS-associated inhibition of isolated *Pleurotus sp.* laccase  
444has also been observed by More et al. (2011). Despite their sensitivity to certain compounds,  
445laccases have huge biotechnological potential in the pulp and textile industry as well as for  
446beverage processing, organic synthesis, the manufacture of biodevices and the detoxification  
447of pollutants (Cañas and Camarero, 2010).

448Feruloyl esterase, a hydrolase catalyzing the cleavage of bonds between plant cell wall  
449polysaccharide and phenolic acid, and bilirubin oxidase, an oxidoreductase from cupredoxin  
450family, were also identified among the secreted proteins. Both of these enzymes are

451 considered promising tools for different biotechnological applications. While feruloyl  
452 esterases may be used for bio-refining of lignocellulosic material for the recovery and  
453 purification of ferulic acid and related hydroxycinnamic acids ubiquitously found in the plant  
454 cell wall (Gopalan et al., 2015), bilirubin oxidases have promising applications in dye effluent  
455 decolorization (Liu et al., 2009) and in biofuel cells (Durand, 2012).

456 A cellobiohydrolase (enzyme releasing cellobiose from the non-reducing ends of cellulose  
457 chains) was also identified among the major *P. ostreatus* proteins produced on DIS (Table 2).  
458 In addition, six other cellulases were also detected in zymograms (Suppl. 6A). The Mw range  
459 of the three enzymes with the most pronounced activities correlated with four theoretically  
460 predicted Mw of *P. ostreatus* cellulases listed in the NCBI Protein database (Suppl. 3A). In  
461 addition, three bands revealing enzymes with xylanolytic activity correlating with *P. ostreatus*  
462 putative xylanases KDQ24406, KDQ24407 and KDQ33930 were visualized (Suppl. 3B).  
463 Another fainter degradation band with a Mw between 45 and 50 kDa did not correlate with  
464 any of the predicted gene products indicating it may be a post-translationally modified  
465 enzyme. Interestingly, glycoside hydrolases detected in the zymograms proved resistant to  
466 rigorous pre-separation treatment (heating in SDS-buffer), as their activities were recovered  
467 after the renaturation steps. Similar stability has previously been reported with some other  
468 extracellular glycoside hydrolases, i.e. from anaerobic cellulolytic bacteria (Vodovnik and  
469 Marinšek Logar, 2012; Hatefi et al., 2017).

470 In addition to proteins traditionally involved in lignocellulose degradation, peptidase 1 was  
471 also detected among proteins of *P. ostreatus* PLAB produced on DIS. Fungal proteases  
472 typically act as virulence factors (Monod et al., 2002), but the results of current research  
473 indicate that they may also play a role in debranching polysaccharide linkages within plant  
474 cell wall proteins and thus providing access to proteins as sources of carbon and nitrogen. The  
475 low content of available nitrogen that is typical for the majority of lignocellulosic substrates

476implies that the expression of specific proteases could be advantageous for releasing nitrogen  
477from organic matter such as that contained in plant material or associated microbial biomass  
478(Sato et al., 2007). High expression of proteases and peptidases was also detected in  
479*Aspergillus fumigatus* (Adav et al., 2015), *Trichoderma reesei* (Borin et al., 2015),  
480*Pycnoporus coccineus* (Couturier et al., 2015) and *Pleurotus ostreatus* PC9 (Fernández-Fueyo  
481et al., 2016) secretomes produced on different lignocellulosic biomasses. The enzyme  
482identified in our study belongs to family S8 peptidases (subtilase clan), which are typically  
483thermostable alkaline peptidases, that may be suitable for different biotechnological  
484applications (especially detergent industry) (Apolinar-Hernández et al., 2016).  
485Implementation of enzyme technologies in paper, textile and biofuel industry is a promising  
486means toward cleaner industrial production. The results of several studies show that  
487implementing enzymatic processes in place of conventional processes generally results in a  
488reduced contribution to global warming and also a reduced contribution to acidification,  
489eutrophication, photochemical ozone formation and energy use (Jeganathan and Nielsen,  
4902013).

491

### 4923.5. Enzymes production from DIS

493Due to its efficient growth on pretreated as well as non-pretreated DIS and its highest  
494screened enzyme activities, *P. ostreatus* PLAB was chosen as the most promising candidate  
495for further quantitative analysis of enzyme production on target substrate. Target enzyme  
496activities detected in enzyme extracts at different fermentation times are shown Fig. 4. The  
497endoglucanase activities recovered from overgrown substrate consistently remained between  
4981.5 – 3.0  $\mu\text{mol}/\text{min}/\text{g}_{\text{DIS}}$  throughout all the growth stages. This is higher than previously  
499reported for *P. ostreatus* during SSF on coffee pulp and similar as in cultures grown on a  
500mixture of supplemented wheat and rice straw as a substrate (Kurt and Buyukalaca, 2010). On

501the other hand, Khalil et al. (2011) reported higher endoglucanase activity after 12 days of  
502SSF on rice straw (Suppl. 2A).

503Xylanase activities in the extracts varied from 1.5 - 4.0  $\mu\text{mol}/\text{min}/\text{g}$ . These activities are lower  
504than detected in extracts from mycelia grown on corn and wheat bran, tomato waste or a  
505mixture of *Fagus sylvatica* leaves and wheat straw as substrates (Suppl. 2B). This may be  
506associated with the lack of inducing compounds originating from the substrate, since natural  
507(ligno)cellulosic substrates usually contain a substantial portion of hemicelluloses (20-30%),  
508while its amounts in paper sludge remain very low due to their removal during the pulping  
509process, with only 3.7% hemicellulose remaining in the DIS used in our study. The  
510mechanisms of regulation and substrate-associated expression patterns of genes encoding  
511hemicellulases in *P. ostreatus* are not yet well understood. However, research on related fungi  
512confirmed their coordinated regulation is associated with molecular signals reflecting  
513substrate composition (Cai et al., 2017).

514The changes in endoglucanase activity and xylanase activity did not correlate with total  
515protein concentration time-wise, since enzyme production apparently peaked in earlier  
516fermentation (growth) phases (maximal endoglucanase and xylanase activity was detected at  
51723 days of growth), which is in accordance with the fact that the production of these enzymes  
518is an essential predisposition for nutrient acquirement from the substrate (Fig. 4A).

519In contrast to endoglucanase and xylanase, laccase activity in *P. ostreatus* DIS-induced  
520secretome strongly correlated with total proteins (reflecting growth of fungal mycelia) and  
521peaked between 30 and 44 days of incubation, achieving its highest average activity of  $49 \pm 5$   
522 $\mu\text{mol ABTS}^{\cdot+}/\text{min}/\text{g}_{\text{DIS}}$  after 37 days and declining thereafter (Fig. 4B). Laccase activity  
523detected in enzyme extracts from *P. ostreatus* on DIS were significantly (10-30 times) higher  
524than reported for other lignocellulose substrates, such as lyophilized wheat straw, wheat bran

525and spent brewery grains (Suppl. 2C). Since DIS contains lower amounts of lignin than  
526reported for the aforementioned substrates (6.2% in our study), high laccase activity in  
527*P. ostreatus* PLAB secretome may be associated with other inducing compounds present in  
528this substrate. In addition to nitrogen source and concentration, heavy metal ions and certain  
529phenolic and (poly)aromatic compounds with structural similarities to lignin have previously  
530been shown to positively regulate expression of laccases (Zhuo et al., 2017) and other  
531ligninolytic enzymes (Zhang et al., 2016). The deinking sludge used in our study contained  
532significant amounts of copper (64.99 mg/kg dry content), which has been reported to act as a  
533strong transcriptional inducer and activity promotor of laccases (Lorenzo et al., 2006) and  
534lignin-modifying peroxidases with several white rots (Zheng et al., 2017).  
535Due to possible effects of substrate composition on enzyme production, it is important that the  
536substrates from different paper millily be characterized so as to support optimization of  
537amounts and types of enzymes produced. If the production of specific type of enzymes would  
538be an objective, the substrate composition should be adjusted by adding relevant inducers (i.e.  
539xylan for increasing xylanase production or specific heavy metals to increase ligninase  
540production).

541

### 5423.6. Advantages of enzyme production from DIS via SSF

543The advantage of SSF processes for enzyme production is lower water consumption and  
544hence relatively less energy is required to concentrate the target enzymes (Yoon et al., 2014).  
545DIS produced in current practice is a waste stream with zero value, substantial disposal costs  
546and no proper environmentally friendly disposal method, therefore repurposing it as a  
547feedstock not only reduces the waste to be disposed from tissue printed recycle mills but also  
548adds revenue to the industry (Gottomukkala, 2016). Laccases are enzymes with tremendous  
549application potential in paper and pulp industry for delignification and brightening of pulp, as

550 well as in bioremediation of toxic compounds released from the industry (Virk et al., 2012).  
551 Endoglucanases and small quantities of xylanases produced may be used for conversion of  
552 cellulosic rich feedstock to value added chemicals and hence can be used within the paper  
553 industry to convert paper sludge or undigested residues of enzyme fermentation to fuel and  
554 chemicals (Gottumukkala et al., 2016). This study has proven that DIS is a potential feedstock  
555 for production of lignocellulytic enzymes and induces laccase production at high  
556 concentrations. Moreover, this study has shown that DIS as a feedstock needs no additional  
557 supplementation of nutrients which thus reduces the burden of waste on the environment.  
558 Well known bioprocess like direct ethanol production from paper sludge requires high solids  
559 loading to achieve threshold ethanol concentrations of 40 g/L (Robus et al., 2016). Viscosity  
560 of paper sludge at such high solids loading is a major limitation and high ash content of the  
561 sludge significantly influences the final ethanol concentration (Gottumukkala et al., 2016).  
562 Hence, the suggested technology, to use *Pleurotus ostreatus* for enzymes production with DIS  
563 by SSF without any supplementation, pretreatment and sterilization requirement, may  
564 therefore be a suitable cost-effective alternative for paper sludge valorisation to ethanol.  
565 However, to increase the enzyme concentration and yields further experiments testing co-  
566 fermentation with additional nitrogen rich waste streams need to be explored in the future.  
567 This approach would help in valorisation and treatment of two different waste streams. In  
568 addition, the effects of the variations in the DIS composition on the ratios of the different  
569 types of the produced enzymes should be explored. The residues from the SSF of DIS should  
570 be investigated and potentially used for biogas production (Steffen et al., 2017b). Further  
571 studies are warranted, to optimize the production, economic feasibility and environmental  
572 benefits of this technology.

573

574

## 5754. **Conclusions**

576 Enzymes production for lignocellulose degradation is currently a cost intensive process and  
577 utilises fresh water and nutrients which in turn results in additional waste stream generation. If  
578 combined with efficient and robust microbial enzyme producers, lignocellulosic wastes have  
579 an enormous potential for developing a sustainable chemical and energy industry as the most  
580 inexpensive feedstock among all the renewable resources. The results of our study have  
581 shown that the deinking paper sludge represents a suitable feedstock for the production of  
582 lignocellulolytic enzymes by selected *Pleurotus* strains. The selected strains exhibited  
583 outstanding ability to transform untreated and unsterilized DIS to a mixture of alkali-stable  
584 enzymes (mainly with endoglucanase, xyylanase and particularly high laccase activities).  
585 Production of lignocellulases from unsterilized DIS by SSF should therefore be considered at  
586 an industrial level to add value to the major waste stream of paper industry by reducing waste  
587 disposal, costs associated with the disposal of wastes to landfill sites and by yielding high  
588 value enzymes suitable for application in biorefinery, paper and pulp, detergent and textile  
589 industries. Further research should be done to analyse (anticipated positive) effects of the  
590 process on the levels and bioavailability of environmental pollutants potentially present in  
591 DIS. In addition, the enzymes production should be further optimized and process conditions  
592 fine-tuned to increase the yield and productivity of individual enzymes and the cocktail.  
593 Major advantages of this process such as less water usage, less energy usage and reduced  
594 waste disposal should be studied in detail for its energy, economic and environment benefits.

595

## 596 **Acknowledgements**

597 This research was supported by Slovenian Research Agency (ARRS). We thank prof. R.  
598 Marinšek Logar for helpful advice, Dr. T. Kranjc for help with data management, N. Vrhovnik

599and G. Lavrič for technical help with preliminary growth experiments and dr. J. Burkeljca for  
600help with graphics.

601

## 602References

603Abdullah, R., Ishak, C.F., Kadir, W.R., Bakar, R.A., 2015. Characterization and feasibility

604assessment of recycled paper mill sludges for land application in relation to the environment.

605Int. J. Environ. Res. Public Health 12(8), 9314-9329. <https://doi.org/10.3390/ijerph120809314>

606Adav, S.S., Ravindran, A., Sze, S.K., 2015. Quantitative proteomic study of *Aspergillus*

607*Fumigatus* secretome revealed deamidation of secretory enzymes. J. Proteomics 119, 154–

608168. <https://doi.org/10.1016/j.jprot.2015.02.007>

609Apolinar-Hernandez, M.M., Pena-Ramirez, Y.J., Perez-Rueda, E., Canto-Canche, B.B., De

610Los Santos-Briones, C., O'Connor-Sanchez, A., 2016. Identification and in silico

611characterization of two novel genes encoding peptidases S8 found by functional screening in a

612metagenomic library of Yucatan underground water. Gene 593(1), 154-161.

613<https://doi.org/10.1016/j.gene.2016.08.009>

614Bajpai P., 2013. Recycling and Deinking of Recovered Paper. Recycling and Deinking of

615Recovered Paper. First ed. Elsevier.

616Beauchamp, C.J., Charest, M.H., Gosselin, A., 2002. Examination of environmental quality of

617raw and composting de-inking paper sludge. Chemosphere 46(6), 887-895.

618[https://doi.org/10.1016/S0045-6535\(01\)00134-5](https://doi.org/10.1016/S0045-6535(01)00134-5)

619Bhattacharya, S., Das, A., Prashanthi, K., Palaniswamy, M., Angayarkanni, J., 2013.

620Mycoremediation of Benzo[a]pyrene by *Pleurotus ostreatus* in the presence of heavy metals

621and mediators. 3 Biotech. 4(2), 205-211. <https://doi.org/10.1007/s13205-013-0148-y>

622 Boshoff, S., Gottumukkala, L.D., van Rensburg, E., Görgens, J., 2016. Paper sludge (PS) to  
623 bioethanol: Evaluation of virgin and recycle mill sludge for low enzyme, high-solids  
624 fermentation. *Bioresour. Technol.* 203, 103-111.  
625 <https://doi.org/10.1016/j.biortech.2015.12.028>

626 Brodeur, G., Yau, E., Badal, K., Collier, J., Ramachandran, K.B., Ramakrishnan, S., 2011.  
627 Chemical and physicochemical pretreatment of lignocellulosic biomass: A review. *Enzyme*  
628 *Res.* 2011, 1-17. <https://doi.org/10.4061/2011/787532>

629 Cai, Y., Gong, Y., Liu, W., Hu, Y., Chen, L., Yan, L., Zhou, Y., Bian, Y., 2017. Comparative  
630 secretomic analysis of lignocellulose degradation by *Lentinula edodes* grown on  
631 microcrystalline cellulose, liginosulfonate and glucose. *J. Proteomics* 163, 92–101.  
632 <https://doi.org/10.1016/j.jprot.2017.04.023>

633 Cañas, A.I., Camarero, S., 2010. Laccases and their natural mediators: Biotechnological tools  
634 for sustainable eco-friendly processes. *Biotechnol. Adv.* 28(6), 694-705.  
635 <https://doi.org/10.1016/j.biotechadv.2010.05.002>

636 Castanera, R., Pérez, G., Omarini, A., Alfaro, M., Pisabarro, A.G., Faraco, V., Amore, A.,  
637 Ramírez, L., 2012. Transcriptional and enzymatic profiling of *Pleurotus ostreatus* laccase  
638 genes in submerged and solid-state fermentation cultures. *Appl Env. Microbiol* 78, 4037-  
639 4045. <https://doi.org/10.1128/AEM.07880-11>

640 Childs, B.R.E., Bardsley, W.G., 1975. The Steady-State Kinetics of Peroxidase with 2,2'-  
641 Azino-di-(3-ethylbenzthiazoline- 6-sulphonic acid) as Chromogen. *Biochem. J.* 145, 93-103  
642 <https://doi.org/10.1042/bj1450093>

643 Couturier, M., Navarro, D., Chevret, D., Henrissat, B., Piumi, F., Ruiz-Dueñas, F.J., Martinez,  
644 A.T., Grigoriev, I. V., Riley, R., Lipzen, A., Berrin, J.-G., Master, E.R., Rosso, M.-N., 2015.

645Enhanced degradation of softwood versus hardwood by the white-rot fungus *Pycnoporus*  
646*coccineus*. *Biotechnol. Biofuels.* 8(216), 1-16. <https://doi.org/10.1186/s13068-015-0407-8>

647Dahl O. 2008. *Papermaking Science and Technology: Environmental Management and*  
648*Control*. Second Ed. Finnish Paper Engineers' Association, Helsinki.

649Deviatkin, I., Kapustina, V., Vasilieva, E., Isyanov, L., Horttanainen, M., 2016. Comparative  
650life cycle assessment of deinking sludge utilization alternatives. *J. Clean. Prod.* 112, 3232–  
6513243. <https://doi.org/10.1016/j.jclepro.2015.10.022>

652Durand, F., Kjaergaard, C.H., Suraniti, E., Gounel, S., Hadt, R.G., Solomon, E.I., Mano, N.,  
6532012. Bilirubin oxidase from *Bacillus pumilus*: A promising enzyme for the elaboration of  
654efficient cathodes in biofuel cells. *Biosens. Bioelectron.* 35, 140–146.  
655<https://doi.org/10.1016/j.bios.2012.02.033>

656Dyballa, N., Metzger, S., 2009. Fast and Sensitive Colloidal Coomassie G-250 Staining for  
657Proteins in Polyacrylamide Gels. *J. Vis. Exp.* 3, 2-5. <https://doi.org/10.3791/1431>

658Elisashvili, V., Penninckx, M., Kachlishvili, E., Tsiklauri, N., Metreveli, E., Kharziani, T.,  
659Kvesitadze, G., 2008. *Lentinus edodes* and *Pleurotus* species lignocellulolytic enzymes  
660activity in submerged and solid-state fermentation of lignocellulosic wastes of different  
661composition. *Bioresour. Technol.* 99, 457-462. <https://doi.org/10.1016/j.biortech.2007.01.011>

662Ellilä, S., Fonseca, L., Uchima, C., Cota, J., Goldman, G.H., Saloheimo, M., Sacon, V., Siika-  
663Aho, M., 2017. Development of a low-cost cellulase production process using *Trichoderma*  
664*reesei* for Brazilian biorefineries. *Biotechnol. Biofuels.* 10 (30), 1-17.  
665<https://doi.org/10.1186/s13068-017-0717-0>

666Fernández-Fueyo, E., Ruiz-Dueñas, F.J., López-Lucendo, M.F., Pérez-Boada, M., Rencoret,  
667J., Gutiérrez, A., Pisabarro, A.G., Ramírez, L., Martínez, A.T., 2016. A secretomic view of

668woody and nonwoody lignocellulose degradation by *Pleurotus ostreatus*. *Biotechnol.*  
669*Biofuels.* 9(49), 1-18. <https://doi.org/10.1186/s13068-016-0462-9>

670Fukushima, R.S., Kerley, M.S., Ramos, M.H., Porter, J.H., Kallenbach, R.L., 2015.  
671Comparison of acetyl bromide lignin with acid detergent lignin and Klason lignin and  
672correlation with in vitro forage degradability. *Anim. Feed Sci. Technol.* 201, 25-37.  
673<https://doi.org/10.1016/j.anifeedsci.2014.12.007>

674Furlan de Jesus, J.P., Sain, M., Jeng, R., Negrão, D.R., Leão, A.L., Nogueira de Andrade,  
675M.C., Almeida Minhoni, M.T. de, 2015. Potential application of *Ganoderma lucidum* in solid  
676state fermentation of primary sludge and wheat straw. *BioResources* 10, 3197-3209.  
677<https://doi.org/10.15376/biores.10.2.3197-3209>

678Garajova, S., Mathieu, Y., Beccia, M.R., Bennati-Granier, C., Biaso, F., Fanuel, M., Ropartz,  
679D., Guigliarelli, B., Record, E., Rogniaux, H., Henrissat, B., Berrin, J.-G., 2016. Single-  
680domain flavoenzymes trigger lytic polysaccharide monooxygenases for oxidative degradation  
681of cellulose. *Sci. Rep.* 6, 28276. <https://doi.org/10.1038/srep28276>

682Gasteiger E, Hoogland C, Gattiker A, Duvaud S, Wilkins MR, Appel RD, Bairoch A., 2009.  
683Protein Identification and Analysis Tools on the ExPASy Server. In: John M. Walker (Ed):  
684The Proteomics Protocols Handbook, Humana Press, New Jersey, pp. 571-607

685Gopalan, N., Rodríguez-Duran, L.V., Saucedo-Castaneda, G., Nampoothiri, K.M., 2015.  
686Review on technological and scientific aspects of feruloyl esterases: A versatile enzyme for  
687biorefining of biomass. *Bioresour. Technol.* 193, 534–544.  
688<https://doi.org/10.1016/j.biortech.2015.06.117>

689Gottumukkala, L.D., Haigh, K., Collard, F.X., van Rensburg, E., Görgens, J., 2016.  
690Opportunities and prospects of biorefinery-based valorisation of pulp and paper sludge.  
691*Bioresour. Technol.* 215, 37-49. <https://doi.org/10.1016/j.biortech.2016.04.015>

692 Gregori, A., Švagelj, M., Pahor, B., Berovič, M., Pohleven, F., 2008. The use of spent brewery  
693 grains for *Pleurotus ostreatus* cultivation and enzyme production. *N. Biotechnol.* 25, 157–161.  
694 <https://doi.org/10.1016/j.nbt.2008.08.003>

695 Hatefi, A., Makhdoumi, A., Asoodeh, A., Mirshamsi, O., 2017. Characterization of a bi-  
696 functional cellulase produced by a gut bacterial resident of Rosaceae branch borer beetle,  
697 *Osphranteria coerulescens* (Coleoptera: Cerambycidae). *Int. J. Biol. Macromol.* 103, 158–164.  
698 <https://doi.org/10.1016/j.ijbiomac.2017.05.042>

699 Heux, S., Meynial-Salles, I., O’Donohue, M.J., Dumon, C., 2015. White biotechnology: State  
700 of the art strategies for the development of biocatalysts for biorefining. *Biotechnol. Adv.* 33,  
701 1653–1670. <https://doi.org/10.1016/j.biotechadv.2015.08.004>

702 Heyer, R., Kohrs, F., Benndorf, D., Rapp, E., Kausmann, R., Heiermann, M., Klocke, M.,  
703 Reichl, U., 2013. Metaproteome analysis of the microbial communities in agricultural biogas  
704 plants. *N. Biotechnol.* 30, 614–622. <https://doi.org/10.1016/j.nbt.2013.01.002>

705 Hidayati N, Surtiningsih T, Matuzahroh N. 2014. Removal of heavy metals Pb, Zn and Cu  
706 from sludge waste of paper industries using biosurfactant. *J Bioremediation Biodegrad.* 5:  
707 255. <https://doi:10.4172/2155-6199.1000255>

708 Huang, S., Huang, D., Wu, Q., Hou, M., Tang, X., Zhou, J., 2017. The Effects of  
709 Environmental C/N on the Activities of Lignin-degrading Enzymes Produced by  
710 *Phanerochaete chrysosporium*. *Pedosphere.* [https://doi.org/10.1016/S1002-0160\(17\)60391-6](https://doi.org/10.1016/S1002-0160(17)60391-6)

711 Humbird, D., Davis, R., Tao, L., Kinchin, C., Hsu, D., Aden, A., 2011. Process Design and  
712 Economics for Biochemical Conversion of Lignocellulosic Biomass to Ethanol. *Renew.*  
713 *Energy.* NREL/TP-5100-47764. <https://doi.org/10.2172/1013269>

714Iandolo, D., Piscitelli, A., Sannia, G., Faraco, V., 2011. Enzyme production by solid substrate  
715fermentation of pleurotus ostreatus and trametes versicolor on tomato pomace. Appl.  
716Biochem. Biotechnol. 163(1), 40-51. <https://doi.org/10.1007/s12010-010-9014-0>

717Jegannathan, K.R., Nielsen, P.H., 2013. Environmental assessment of enzyme use in industrial  
718production – a literature review. J. Clean. Prod. 42, 228–240.  
719<https://doi.org/10.1016/J.JCLEPRO.2012.11.005>

720Junker, B., 2010. Minimizing the environmental footprint of bioprocesses - Part 2: Evaluation  
721of wastewater, electricity, and air emissions. Bioprocess Int. 2010, 62-70

722Kang, L., Wang, W., Lee, Y.Y., 2010. Bioconversion of kraft paper mill sludges to ethanol by  
723SSF and SSCF. Appl. Biochem. Biotechnol. 161, 53-66. <https://doi.org/10.1007/s12010-009-7248893-4>

725Kapahi, M., Sachdeva, S., 2017. Mycoremediation potential of Pleurotus species for heavy  
726metals: a review. Biores Bioproc. 2017; 4(1), 32. <https://doi.org/10.1186/s40643-017-0162-8>

727Khalil, M.I., Hoque, M.M., Basunia, M.A., Alam, N., Khan, M.A., 2011. Production of  
728cellulase by Pleurotus ostreatus and Pleurotus sajor-caju in solid state fermentation of  
729lignocellulosic biomass. Turkish J. Agric. For. 35, 333-341. <https://doi.org/10.3906/tar-1002-730684>

731Knežević, A., Milovanović, I., Stajić, M., Lončar, N., Brčeski, I., Vukojević, J., Čilerdžić, J.,  
7322013. Lignin degradation by selected fungal species. Bioresour. Technol. 138, 117-123.  
733<https://doi.org/10.1016/j.biortech.2013.03.182>

734Kurschner, K., Hoffer, A. 1993. Cellulose and cellulose derivative. Fresenius J Anal Chem.  
73592, 145–154.

- 736 Kurt, S., Buyukalaca, S., 2010. Yield performances and changes in enzyme activities of  
737 *Pleurotus* spp. (*P. ostreatus* and *P. sajor-caju*) cultivated on different agricultural wastes.  
738 *Bioresour. Technol.* 101, 3164-3169. <https://doi.org/10.1016/j.biortech.2009.12.011>
- 739 Kües, U., 2015. Fungal enzymes for environmental management. *Curr. Opin. Biotechnol.* 33,  
740 268–278. <https://doi.org/10.1016/j.copbio.2015.03.006>
- 741 Laemmli, U.K., 1970. Cleavage of structural proteins during the assembly of the head of  
742 bacteriophage T4. *Nature* 227, 680–685. <https://doi.org/10.1038/227680a0>
- 743 Lee, J.W., 2012. Advanced biofuels and bioproducts, *Advanced Biofuels and Bioproducts*.  
744 Springer, New York, pp. 1122. <https://doi.org/10.1007/978-1-4614-3348-4>
- 745 Lever, M., 1977. Carbohydrate determination with 4-hydroxybenzoic acid hydrazide  
746 (PAHBAH): effect of bismuth on the reaction. *Anal. Biochem.* 81, 21–27.
- 747 Liguori, R., Faraco, V., 2016. Biological processes for advancing lignocellulosic waste  
748 biorefinery by advocating circular economy. *Bioresour. Technol.* 215, 13–20.  
749 <https://doi.org/10.1016/j.biortech.2016.04.054>
- 750 Likon, M., Saarela, J., 2012. The Conversion of Paper Mill Sludge into Absorbent for Oil  
751 Spill Sanitation - The Life Cycle Assessment. *Macromol. Symp.* 320, 50–56.  
752 <https://doi.org/10.1002/masy.201251006>
- 753 Likon, M., Trebše, P., 2012. Recent advances in paper mill sludge management, in: Show KY,  
754 Guo X (Eds), *Industrial Waste*. InTech Open, London. pp. 73-90.  
755 <https://doi.org/10.5772/37043>
- 756 Liu, Y., Huang, J., Zhang, X., 2009. Decolorization and biodegradation of remazol brilliant  
757 blue R by bilirubin oxidase. *J. Biosci. Bioeng.* 108, 496–500.  
758 <https://doi.org/10.1016/j.jbiosc.2009.06.001>

759 Lorenzo, M., Moldes, D., Sanromán, M.Á., 2006. Effect of heavy metals on the production of  
760 several laccase isoenzymes by *Trametes versicolor* and on their ability to decolourise dyes.  
761 *Chemosphere* 63, 912–917.  
762 <https://doi.org/http://dx.doi.org/10.1016/j.chemosphere.2005.09.046>

763 Lowry, O.H., Rosebrough, N.J., Farr, A.L., Randall, R.J. 1951. Protein measurement with the  
764 Folin phenol reagent. *J Biol Chem.* 193, 265–275.

765 Machrafi, Y., Chalifour, F.-P., Wamegni, J., Beauchamp, C.J., 2008. Short-term effects of  
766 deinking paper sludge on the dynamics of soil carbon, nitrogen, and phenolic compounds. *J.*  
767 *Agric. Food Chem.* 56, 11399–11406. <https://doi.org/10.1021/jf801443a>

768 Manavalan, T., Manavalan, A., Heese, K., 2015. Characterization of Lignocellulolytic  
769 Enzymes from White-Rot Fungi. *Curr. Microbiol.* 70, 485-498.  
770 <https://doi.org/10.1007/s00284-014-0743-0>

771 Meng, X., Ragauskas, A.J., 2014. Recent advances in understanding the role of cellulose  
772 accessibility in enzymatic hydrolysis of lignocellulosic substrates. *Curr. Opin. Biotechnol.* 27,  
773 150–8. <https://doi.org/10.1016/j.copbio.2014.01.014>

774 Mohan, S.V., Butti, S.K., Amulya, K., Dahiya, S., Modestra, J.A., 2016. Waste Biorefinery: A  
775 New Paradigm for a Sustainable Bioelectro Economy. *Trends Biotechnol.* 34, 852–855.  
776 <https://doi.org/10.1016/j.tibtech.2016.06.006>

777

778 Monod, M., Capoccia, S., Léchenne, B., Zaugg, C., Holdom, M., Jousson, O., 2002. Secreted  
779 proteases from pathogenic fungi. *Int. J. Med. Microbiol.* 292, 405–419.  
780 <https://doi.org/10.1078/1438-4221-00223>

781Monte, M.C., Fuente, E., Blanco, A., Negro, C., 2009. Waste management from pulp and  
782paper production in the European Union. *Waste Manag.* 29, 293–308.  
783<https://doi.org/10.1016/j.wasman.2008.02.002>

784More, S.S., Renuka, P.S., Pruthvi, K., Swetha, M., Malini, S., Veena, S.M., 2011. Isolation,  
785purification, and characterization of fungal laccase from *Pleurotus* sp. *Enzyme Res.* 2011, 1-7.  
786<https://doi.org/10.4061/2011/248735>

787Naghdi, M., Taheran, M., Brar, S.K., Kermanshahi-pour, A., Verma, M., Surampalli, R.Y.,  
7882018. Removal of pharmaceutical compounds in water and wastewater using fungal  
789oxidoreductase enzymes. *Environ. Pollut.* 234, 190–213.  
790<https://doi.org/10.1016/J.ENVPOL.2017.11.060>

791Pereira, J., Giese, E.C., de Souza, M.M., Gomes, A.C., Perrone, O.M., Boscolo, P., Roberto da  
792Silva, Gomes, E., Bocchini Martins, E., 2017. Effect of Metal Ions, Chemical Agents and  
793Organic Compounds on Lignocellulolytic Enzymes Activities, in: Şentürk (Ed.), *Enzyme*  
794*Inhibitors and Activators*. InTech Open, London. <https://doi.org/10.5772/65934>.

795Pointner, M., Kuttner, P., Obrlik, T., Jäger, A., Kahr, H., 2014. Composition of corncobs as a  
796substrate for fermentation of biofuels. *Agron. Res.* 12(2), 391-396.

797Prasetyo, J., Park, E.Y., 2013. Waste paper sludge as a potential biomass for bio-ethanol  
798production. *Korean J. Chem. Eng.* 30, 253-261.

799Quevillon, E., Silventoinen, V., Pillai, S., Harte, N., Mulder, N., Apweiler, R., Lopez, R.,  
8002005. InterProScan: Protein domains identifier. *Nucleic Acids Res.* 33, W116-120.  
801<https://doi.org/10.1093/nar/gki442>

802Robus, C.L.L., Gottumukkala, L.D., van Rensburg, E., Görgens, J.F., 2016. Feasible process  
803development and techno-economic evaluation of paper sludge to bioethanol conversion:

804 South African paper mills scenario. *Renew. Energy*, 92, 333-345.

805 <https://doi.org/10.1016/j.renene.2016.02.017>

806 Sato, S., Liu, F., Koc, H., Tien, M., 2007. Expression analysis of extracellular proteins from

807 *Phanerochaete chrysosporium* grown on different liquid and solid substrates. *Microbiology-*

808 *Sgm* 153, 3023-3033. <https://doi.org/10.1099/mic.0.2006/000513-0>

809 Schomburg, I., Chang, A., Ebeling, C., Gremse, M., Heldt, C., Huhn, G., Schomburg, D.,

810 2004. BRENDA, the enzyme database: updates and major new developments. *Nucleic Acids*

811 *Res.* 2, D431-D433. <https://doi.org/10.1093/nar/gkh081>

812 Schweikl, H., Klein, U., Schindlbeck, M., Wieczorek, H., 1989. A vacuolar-type ATPase,

813 partially purified from potassium transporting plasma membranes of tobacco hornworm

814 midgut. *J. Biol. Chem.* 264(19), 11136-11142.

815 Steffen, F., Janzon, R., Saake, B., 2017a. Enzymatic treatment of deinking sludge – effect on

816 fibre and drainage properties. *Environ. Technol.* 2017, 1-17.

817 <https://doi.org/10.1080/09593330.2017.1365948>

818 Steffen, F., Janzon, R., Wenig, F., Saake, B., 2017b. Valorization of waste streams from

819 deinked pulp mills through anaerobic digestion of deinking sludge. *BioResources*, 12, 4547-

820 4566. <https://doi.org/10.15376/biores.12.3.4547-4566>

821 Teather, R.M., Wood, P.J., 1982. Use of Congo red-polysaccharide interactions in enumeration

822 and characterization of cellulolytic bacteria from the bovine rumen. *Appl. Environ. Microbiol.*

823 43, 777-780. <https://doi.org/10.1099/aem.1982.43.4.777-780>

824 Tóth, G., Hermann, T., Da Silva, M.R., Montanarella, L., 2016. Heavy metals in agricultural

825 soils of the European Union with implications for food safety. *Environ. Int.* 88, 299–309.

826 <https://doi.org/10.1016/j.envint.2015.12.017>

827 Velázquez-Cedeño, M.A., Mata, G., Savoie, J.M., 2002. Waste-reducing cultivation of  
828 *Pleurotus ostreatus* and *Pleurotus pulmonarius* on coffee pulp: Changes in the production of  
829 some lignocellulolytic enzymes. *World J. Microbiol. Biotechnol.* 18, 201-207.  
830 <https://doi.org/10.1023/A:1014999616381>

831 Virk, A.P., Sharma, P., Capalash, N., 2012. Use of laccase in pulp and paper industry.  
832 *Biotechnol. Prog.* 28(1), 21-32. <https://doi.org/10.1002/btpr.727>

833 Vodovnik, M., Logar, R.M., 2012. Expression patterns of *Ruminococcus flavefaciens* 007S  
834 cellulases as revealed by zymogram approach. *Folia Microbiol.* 57, 367-370.  
835 <https://doi.org/10.1007/s12223-012-0144-3>

836 Vodovnik, M., Logar, R.M., 2010. Cellulosomes - Promising supramolecular machines of  
837 anaerobic cellulolytic microorganisms. *Acta Chim. Slov.* 57(4), 767-774.

838 Vodovnik, M., Duncan, S.H., Reid, M.D., Cantlay, L., Turner, K., Parkhill, J., Lamed, R.,  
839 Yeoman, C.J., Miller, M.E.B., White, B.A., Bayer, E.A., Marinšek-Logar, R., Flint, H.J.,  
840 2013. Expression of Cellulosome Components and Type IV Pili within the Extracellular  
841 Proteome of *Ruminococcus flavefaciens* 007. *PLoS One* 8(6), 1-11.  
842 <https://doi.org/10.1371/journal.pone.0065333>

843 Wan, C., Li, Y., 2012. Fungal pretreatment of lignocellulosic biomass. *Biotechnol. Adv.* 30,  
844 1447–1457. <https://doi.org/10.1016/j.biotechadv.2012.03.003>

845 Wheeler, D.L., Church, D.M., Lash, A.E., Leipe, D.D., Madden, T.L., Pontius, J.U., Schuler,  
846 G.D., Schriml, L.M., Tatusova, T.A., Wagner, L., Rapp, B.A., 2002. Database resources of the  
847 National Center for Biotechnology Information: 2002 update. *Nucleic Acids Res.* 30(1), 13-  
848 16. <https://doi.org/10.1093/nar/gkh073>

849 Williams A. The production of bioethanol and biogas from paper sludge. Chemical  
850 Engineering, Stellenbosch University. 2017 (Masters' thesis).

851 Willis, J.D., Klingeman, W.E., Oppert, C., Oppert, B., Jurat-Fuentes, J.L., 2010.  
852 Characterization of cellulolytic activity from digestive fluids of *Dissosteira carolina*  
853 (Orthoptera: Acrididae). *Comp. Biochem. Physiol. Part B Biochem. Mol. Biol.* 157, 267–272.  
854 <https://doi.org/10.1016/j.cbpb.2010.06.012>

855 Wise, L.E., Karl, H.L. 1962. Cellulose and hemicellulose. In: Earl, L.C. (Ed.) *Pulp and paper*  
856 *science and technology*. vol. 1. McGraw Hill Book Co., New York, pp.54-73.

857 Yang, S., Sun, X., Shen, Y., Chang, C., Guo, E., La, G., Zhao, Y., Li, X., 2017. Tolerance and  
858 Removal Mechanisms of Heavy Metals by Fungus *Pleurotus ostreatus* Haas. *Water, Air, Soil*  
859 *Pollut.* 228, 130-139. <https://doi.org/10.1007/s11270-016-3170-y>

860

861 Yoon, L.W., Ang, T.N., Ngoh, G.C., Chua, A.S.M., 2014. Fungal solid-state fermentation and  
862 various methods of enhancement in cellulase production. *Biomass and Bioenergy*. 67, 319-  
863 338. <https://doi.org/10.1016/j.biombioe.2014.05.013>

864 Zdobnov, E.M., Apweiler, R., 2001. InterProScan - An integration platform for the signature-  
865 recognition methods in InterPro. *Bioinformatics* 17, 847-848.  
866 <https://doi.org/10.1093/bioinformatics/17.9.847>

867 Zhang, H., Zhang, S., He, F., Qin, X., Zhang, X., Yang, Y., 2016. Characterization of a  
868 manganese peroxidase from white-rot fungus *Trametes* sp.48424 with strong ability of  
869 degrading different types of dyes and polycyclic aromatic hydrocarbons. *J. Hazard. Mater.*  
870 320, 265–277. <https://doi.org/10.1016/j.jhazmat.2016.07.065>

871 Zheng, F., An, Q., Meng, G., Wu, X.-J., Dai, Y.-C., Si, J., Cui, B.-K., 2017. A novel laccase  
872 from white rot fungus *Trametes orientalis*: Purification, characterization, and application. *Int.*  
873 *J. Biol. Macromol.* 102, 758–770.

874 <https://doi.org/http://dx.doi.org/10.1016/j.ijbiomac.2017.04.089>

875 Zheng, Y., Zhao, J., Xu, F., Li, Y., 2014. Pretreatment of lignocellulosic biomass for enhanced  
876 biogas production. *Prog. Energy Combust. Sci.* 42, 35-53.

877 <https://doi.org/10.1016/j.pecs.2014.01.001>

878 Zhuo, R., Yuan, P., Yang, Y., Zhang, S., Ma, F., Zhang, X., 2016. Induction of laccase by metal  
879 ions and aromatic compounds in *Pleurotus ostreatus* HAUCC 162 and decolorization of  
880 different synthetic dyes by the extracellular laccase. *Biochem. Eng. J.* 117:62-72.

881 <https://doi.org/10.1016/j.bej.2016.09.016>

882 Žužek, M.C., Maček, P., Sepčić, K., Cestnik, V., Frangež, R., 2006. Toxic and lethal effects of  
883 ostreolysin, a cytolytic protein from edible oyster mushroom (*Pleurotus ostreatus*), in rodents.

884 *Toxicon* 48, 264-271. <https://doi.org/10.1016/j.toxicon.2006.05.011>

885

886

887

888

889

890

891

892

## 893 FIGURE CAPTIONS

894 **Fig. 1.** Schematic representation of experimental set-up.

895 **Fig. 2.** A heatmap representing the ability/efficiency of selected fungal strains to grow on  
896 autoclaved (A-DIS) and non-autoclaved (NA-DIS) deinking paper sludge. 0 – no growth  
897 observed in any of the inoculated samples, 1 – mycelium observed in less than half of the  
898 samples, 2- mycelium overgrew more than 50% of the samples, 3- mycelium overgrew more  
899 than 75% of the samples (N=4).

900 **Fig. 3.** Degradation/oxidation zone diameters around secreted enzyme extracts (SEEs) from  
901 different fungal strains applied on screening plates indicating endoglucanase (red), xylanase  
902 (grey) and laccase (blue) activities. Results represent an average values  $\pm$  standard deviations  
903 from 3 biological replicates. Different capital letters above solid bars denote significant ( $P <$   
904  $0.05$ ) differences in endoglucanase activity, while different lowercase letters above hollow  
905 bars denote significant ( $P < 0.05$ ) differences in laccase activity.

906 **Fig. 4.** Activities of target lignocellulolytic enzymes (EA) in *P. ostreatus* PLAB at different  
907 stages of its growth. A: glycoside hydrolase (endoglucanases, xylanases), B: laccase (ABTS-  
908 oxidases). Results represent an average values  $\pm$  standard deviations from 3 biological  
909 replicates.

910

911

912

913

914

915 **Table 1.** Properties (composition) of deinking sludge (DIS) used as a substrate for selected  
 916 fungal strains.

917

<b>Parameter</b>	<b>Value</b>
pH	8.6
Dry content (DC)	61.35 %
Total organic carbon (TOC)	34.53 % (w/w <sub>DC</sub> )
Total nitrogen (TON)	1.48 % (w/w <sub>DC</sub> )
Cellulose content	34.2 % (w/w <sub>DC</sub> )
Lignin content	6.2 % (w/w <sub>DC</sub> )
Hemicellulose content	3.7 % (w/w <sub>DC</sub> )
Copper (Co)	64.99 mg/kg DC
Zinc (Zn)	84.90 mg/kg DC
Cadmium (Cd)	0.40 mg/kg DC
Chromium (Cr)	11.60 mg/kg DC
Nickel (Ni)	2.20 mg/kg DC
Lead (Pb)	17.04 mg/kg DC
Mercury (Hg)	0.09 mg/kg DC

918

919

920

921

922

923

924

925

926

927 **Table 2.** Major fungal proteins likely involved in the degradation of deinking sludge (DIS)  
 928 identified in the secretome extracts of *P. ostreatus* PLAB grown on A-DIS.

<b>Protein hits (fungal enzymes)</b>	<b>Accession No. (NCBI nr)</b>	<b>Mw [kDa]</b>	<b>pI</b>	<b>Detected in band(s)</b>	<b>Matching Peptides #</b>	<b>Peptide scores</b>
<b>Laccase</b>	CAC69853	57.4	5.7	1, 2, 3	8	308.2
<b>Laccase 2</b>	Q12739	56.7	4.7	1, 2	1	132.7
<b>Cellobiohydrolase</b>	AAA50608	46.1	4.8	2	1	90.0
<b>Bilirubin oxidase</b>	BAA85185	56.8	4.6	1, 2	3	158.8
<b>Feruloyl esterase</b>	CDI44666	59.3	6.9	1	2	159.6
<b>Peptidase 1</b>	ACR25273	38.7	7.9	3	3	84.5

929

930

931

932

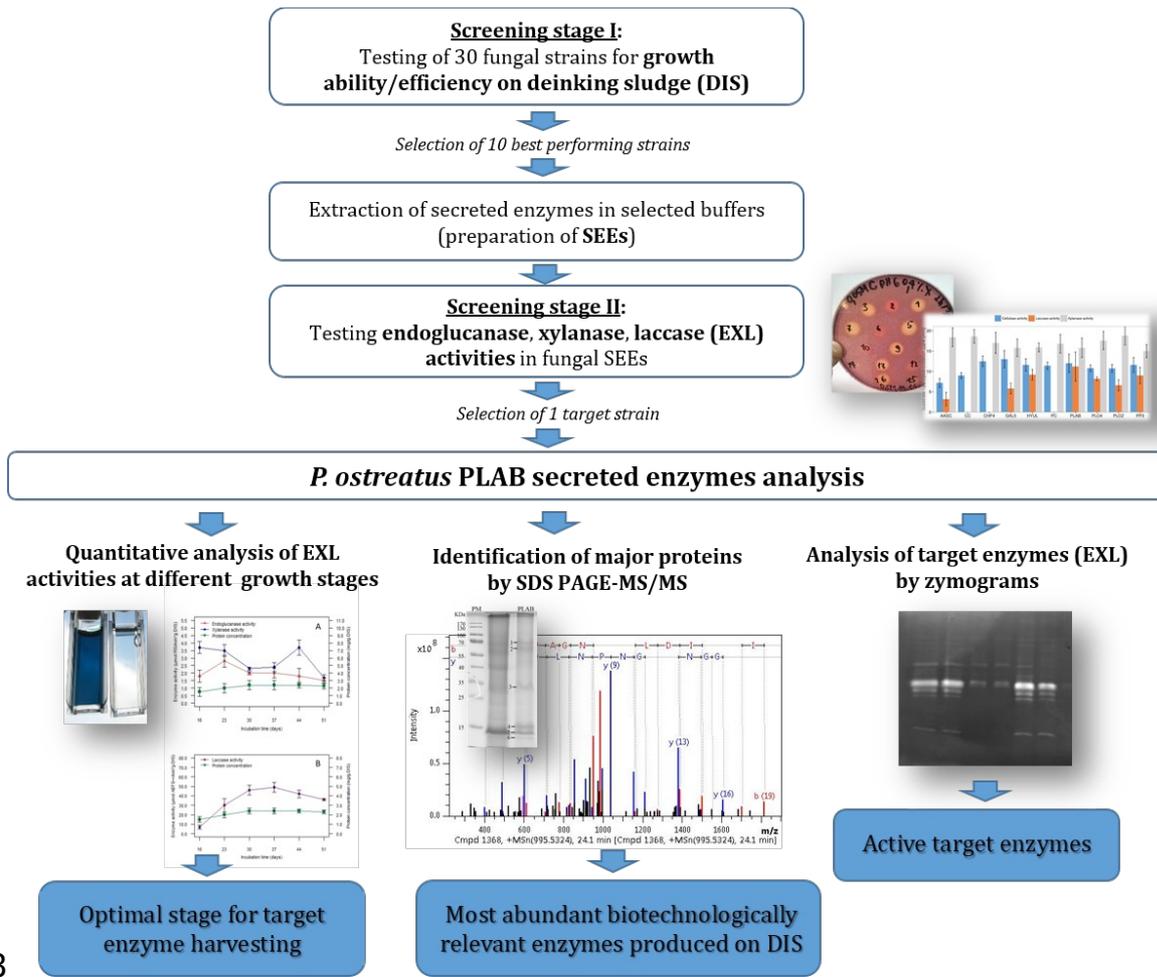
933

934

935

936

937 **Fig. 1**



938

939

940

941

942

943

944

945

946 Fig. 2

<i>Species</i>	<b>Strain accesion code</b>	<b>A-DIS</b>	<b>NA-DIS</b>
<i>Agaricus bisporus</i>	AGB1	1	0
<i>Agrocybe aegerita</i>	AASC	2	1
<i>Auricularia auricula</i>	AAP3	1	0
<i>Auricularia sambucina</i>	AS	0	0
<i>Chondrostereum purpureum</i>	CHP4	2	0
<i>Coprinus comatus</i>	CC	2	0
<i>Cordyceps militaris</i>	CM2	0	0
<i>Fomes fomentarius</i>	FF366	1	0
<i>Ganoderma lucidum</i>	GAL5	2	1
<i>Gleophyllum trabeum</i>	GT	0	0
<i>Grifola frondosa</i>	GF5	0	0
<i>Grifola frondosa</i>	GF2	0	0
<i>Hypsizygus ulmarius</i>	HYUL1	2	1
<i>Inonotus obliquus</i>	IO	1	0
<i>Lentinula edodes</i>	LE3770	0	0
<i>Lentinula edodes</i>	LE7401	0	0
<i>Meripilus giganteus</i>	MG2	1	0
<i>Monascus ruber</i>	M2	0	0
<i>Pleurotus eryngii</i>	POE1	1	1
<i>Pleurotus ostreatus</i>	PLO4	2	1
<i>Pleurotus ostreatus</i>	PLAB	3	1
<i>Pleurotus ostreatus</i>	H35	2	1
<i>Pleurotus pulmonarius</i>	PP3	2	0
<i>Pleurotus sp.</i>	PLOZ	2	1
<i>Polyporus umbellatus</i>	PLUI	0	0
<i>Polyporus ciliatus</i>	PC	2	0
<i>Poria vaillantii</i>	PV2	0	0
<i>Psilocybe cubensis</i>	PSCUII.	0	0
<i>Trametes versicolor</i>	TVER	1	0
947 <i>Wolfiporia cocos</i>	DSM1216	0	0

948

949

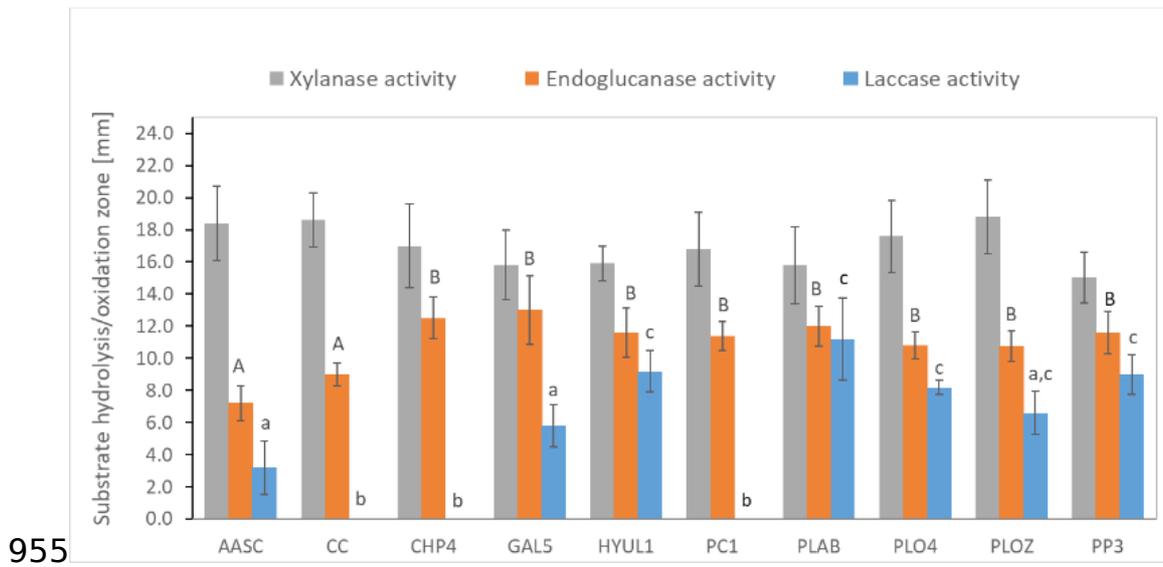
950

951

952

953

954 Fig.3



955

956

957

958

959

960

961

962

963

964

965

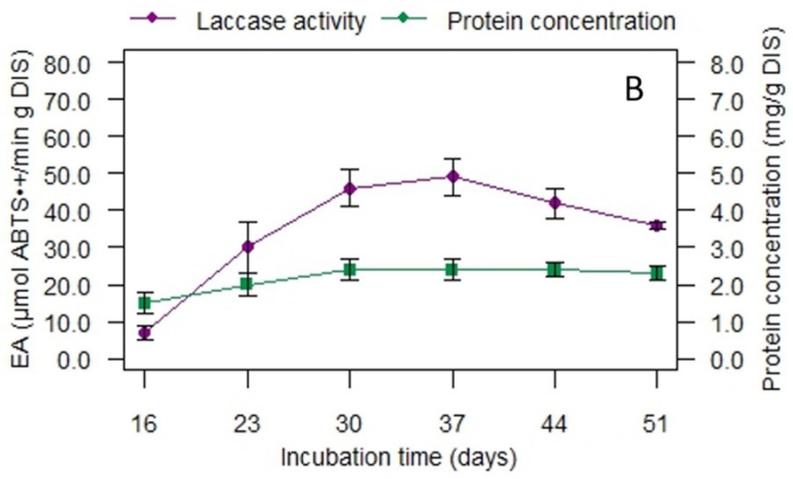
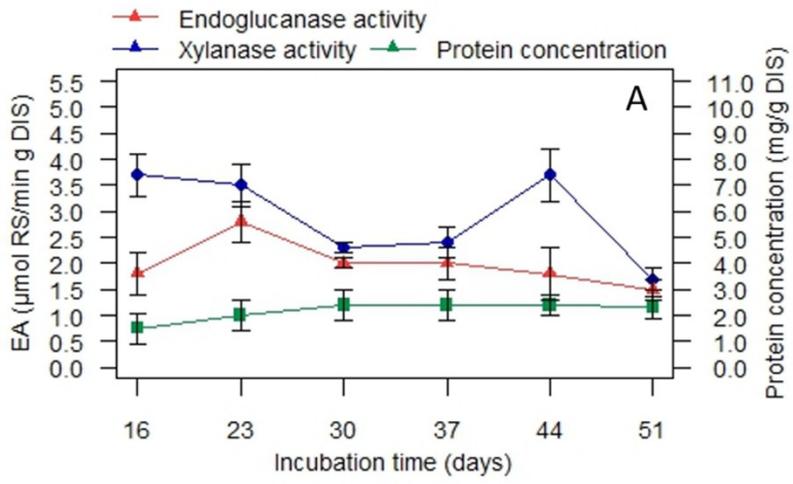
966

967

968 Fig. 4

87

88



969

970