

## *Tansley review*

# Plant phosphorus-use and -acquisition strategies in Amazonia

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### Notes S1. Leaf P<sub>i</sub>-resorption efficiency

Leaf P<sub>i</sub>-resorption efficiency is an essential metric for dynamic vegetation models (Fleischer *et al.*, 2019); it is defined as the percentage of remobilized nutrients in fully senesced leaves compared with those in mature green leaves (Van Heerwaarden *et al.*, 2003; Killingbeck, 2004). However, the relationship between leaf P<sub>i</sub>-resorption efficiency and soil P availability is still ambiguous due to contrasting observations. Various field studies and meta-analyses have shown a robust negative relationship between P<sub>i</sub>-resorption efficiency and soil P concentrations (Yuan & Chen, 2009; Hidaka & Kitayama, 2011; Han *et al.*, 2013; Hayes *et al.*, 2014; Tsujii *et al.*, 2017; Wang *et al.*, 2018; He *et al.*, 2020), while others showed weak or no relationships (Aerts, 1996; Vitousek, 1998; Aerts & Chapin III, 1999; Eckstein *et al.*, 1999; Wright & Westoby, 2003; Tang *et al.*, 2013). This controversy might have resulted from methodological inconsistencies among studies, which have been extensively covered in the literature (Van Heerwaarden *et al.*, 2003; Killingbeck, 2004; Luysaert *et al.*, 2005; Ares & Gleason, 2007). Most importantly, Van Heerwaarden *et al.* (2003) show leaf mass and leaf area loss during senescence should be accounted for in leaf P<sub>i</sub>-resorption efficiency as these losses can underestimate resorption efficiency by 20% 10%, respectively.

Most studies involving natural soil P gradients show evidence of soil P effects on leaf P<sub>i</sub>-resorption efficiency, at least partially. Along a strong soil P gradient in an Australian dune chronosequence, mean PRE across different species varied from zero in high-P soils to 79% in low-P soils (Hayes *et al.*, 2014). Similarly, a strong negative relationship between leaf P<sub>i</sub>-resorption efficiency and soil P concentrations was found in the rainforests of Borneo (Hidaka & Kitayama, 2011; Tsujii *et al.*, 2017). For instance, Tsujii *et al.* (2017) found community-level mean leaf P<sub>i</sub>-resorption efficiency varied from 24 to 93% and estimated that 20 to 37% of this variation was explained by the site, while 25 to 43% was explained by the genus, highlighting the role of adaptations in nutrient resorption. At the species level, variation in leaf P<sub>i</sub>-resorption efficiency of a common species in neotropical forests was explained by soil P<sub>i</sub> availability and the reproductive status of the plant, as individuals with high reproductive demands resorbed greater P amounts than those with no reproductive demands (Tully *et al.*, 2013).

Although leaf P<sub>i</sub>-resorption efficiency has not been measured along fertility gradients in Amazonia, several small-scale studies show high variability across species and communities. In central Amazonia, mean leaf P<sub>i</sub>-resorption efficiency among species varied from 53% to 74% (Gomes & Luizão, 2012; Machado *et al.*, 2016), in north Amazonia from 41% to 82% (Scott *et al.*, 1992; Reich *et al.*, 1995), and in northeastern Amazonia, PRE varied from 26% to 89% (Hättenschwiler *et al.*, 2011). At the community level, in a forest succession chronosequence (6 to 200 years) in eastern Amazonia, PRE varied from 53% to 73% (Reed *et al.*, 2012). Thus, leaf P<sub>i</sub>-resorption efficiency in Amazonia is similar to the range of 24% and 93% found along the P gradient of tropical forests of Borneo (Tsujii *et al.*, 2017). We hypothesize that community-level leaf P<sub>i</sub>-resorption

efficiency in Amazonia follows the same negative relationship with soil P concentrations. Studies focused on the relationships between leaf  $P_i$ -resorption efficiency and soil P concentrations in Amazonia, following standardized methodologies, have the potential to offer valuable insights into this plant P-use strategy.

## Notes S2. Effects of N on acid phosphatase exudation and activity

Root expression of phosphatases is regulated by plant  $P_i$  demand (McGill & Cole, 1981). Therefore, it is expected that as soil  $P_i$  availability decreases, C and N investments in phosphatases increase. An increase in root PME activity with N fertilization has been observed in tropical forests in Hawaii (Treseder & Vitousek, 2001) and other ecosystems (Marklein & Houlton, 2012). In contrast, N fertilization did not affect root PME activity in forests of central Amazonia (Lugli *et al.*, 2021) and Borneo (Yokoyama *et al.*, 2017), suggesting N is not a limiting factor for PME release in these forests. N and P co-limitation may happen in montane Andean forests, in the western region of Amazonia in Ecuador and Peru, where N limitation appears to increase with increasing elevation (Wullaert *et al.*, 2010; Homeier *et al.*, 2012; Fisher *et al.*, 2013), and might happen in white-sand forests (Vitousek & Sanford Jr, 1986; Martinelli *et al.*, 1999). However, currently, there is no evidence of widespread N and P co-limitation in Amazonia (Quesada *et al.*, 2010).

It has been suggested that N-fixing plants might have an advantage in P acquisition by investing excess N in the expression of phosphatases (Houlton *et al.*, 2008). In soils beneath N-fixing species, phosphatase activity was three times greater than that of other species (Houlton *et al.*, 2008). Similarly, in tropical forests in Costa Rica and Panama and a coastal dune ecosystem in Australia, phosphatase activity was significantly greater in roots of N-fixing than non-N-fixing species (Nasto *et al.*, 2014; Png *et al.*, 2017; Batterman *et al.*, 2018). However, high root phosphatase activities in N-fixing species were most likely a result of a phylogenetically conserved strategy (Png *et al.*, 2017; Batterman *et al.*, 2018). In Amazonia, legume species are mostly facultative N-fixers and do not seem to be fixing substantial amounts of N (Vitousek *et al.*, 2002; Nardoto *et al.*, 2014). Rates of biological N fixation in a primary forest in southeastern Amazonia were about 20% of the average of other tropical forests (Wong *et al.*, 2019). This might be explained by the high N availability in most soils of Amazonia (Vitousek *et al.*, 2002; Quesada *et al.*, 2010). Alternatively, low P and low molybdenum availability could be the reason for slow rates of biological N fixation, as these might be limiting factors (Reed *et al.*, 2013; Wong *et al.*, 2019). However, increased soil P and molybdenum concentrations did not alter rates of biological N fixation in low-P soils of southeastern Amazonia (Wong *et al.*, 2019).

Currently, the few studies on root PME activity in Amazonia suggest that soil N availability is not a limiting factor in most areas. However, more studies are needed to clarify the role of soil N availability in root PME expression and the role of plant species, such as N-fixing species.

### Notes S3. Cluster roots in Amazonia

Little is currently known about the formation of cluster roots in Amazonia and other tropical forests. One of the plant families most commonly cited for cluster-root formation, Proteaceae (Lambers *et al.*, 2008), has at least 13 species described in Amazonia in the Database of Brazilian plant species (REFLORA, <http://floradobrasil.jbrj.gov.br/>). For instance, *Roupala montana*, a widespread species in Amazonia and other neotropical forests, did not form cluster roots in soils of the Brazilian Cerrado with over 220 mg P<sub>t</sub> kg<sup>-1</sup> (mean resin-P<sub>i</sub> varied from 5.2 to 6.6 mg kg<sup>-1</sup>) but formed associations with arbuscular mycorrhizal fungi (da Silva Coutinho Detmann *et al.*, 2019). The authors suggested that soil P availability might have been high enough for these plants to still benefit from the symbiosis (da Silva Coutinho Detmann *et al.*, 2019). Although it is not typical for Proteaceae species to form associations with mycorrhizas, it has been previously observed in earlier studies (Boulet & Lambers, 2005; Lambers *et al.*, 2015). In another study, cluster roots in *R. montana* only developed in the treatment with no added P. Roots of *R. montana* were associated with significantly higher phosphatase activity and use of phytate when compared with three mycorrhizal species (Steidinger *et al.*, 2014). Although root organic acid exudation was not measured in Steidinger *et al.* (2014), the superior ability of *R. montana* to use phytate suggest that rapid rates of root organic acid exudation could have been the reason, even without the formation of cluster roots, as phytate is thought to be strongly adsorbed to the soil matrix (see Box 4; Gerke, 2015). Moreover, another species known to form cluster roots, *Euplassa cantareirae* (Proteaceae), most commonly found in the Brazilian Atlantic tropical forest, has been studied in a greenhouse experiment, and formation of its cluster roots did not depend on P supply (de Britto Costa *et al.*, 2016). More studies are necessary to find out how common the formation of cluster roots is in Amazonia and other tropical forests, under what conditions they form, and their role in P acquisition.

**Table S1. Summary table of the cited studies along soil P gradients.** \*Mehlich-1. \*\*Resin-P<sub>i</sub>. \*\*\*Soil P<sub>t</sub> concentration data.

Strategy/trait	Place	Soil P <sub>t</sub> (mg kg <sup>-1</sup> )	Results	Reference
Leaf P <sub>i</sub> resorption proficiency	Northeastern Amazonia	7 to 600	Senesced leaf P concentrations decreased with decreasing soil P <sub>t</sub>	Soong <i>et al.</i> (2020)
	Northwestern Amazonia	40 to 480		Lips and Duivenvoorden (1996)
	Hawaii	280 to 980		Vitousek (1998); Olander and Vitousek (2000)***
	Mount Kinabalu, Borneo	20 to 417		Hidaka and Kitayama (2011)
	Mount Kinabalu, Borneo	20 to 417		Tsujii <i>et al.</i> (2017)
	Eastern China	378 to 1290		Tang <i>et al.</i> (2013)
	New Zealand	~100 to 900		Richardson <i>et al.</i> (2005); Richardson <i>et al.</i> (2004)***
	Southwestern Australia	6.6 to 432.2		Hayes <i>et al.</i> (2014)
SRL	Atlantic forest, Brazil	5.4 to 8.9*	Higher in the low-P sites	Zangaro <i>et al.</i> (2008)
	Northeastern Amazonia	3 to 36*		Metcalf <i>et al.</i> (2008)
	New Zealand	108 to 804		Holdaway <i>et al.</i> (2011); Richardson <i>et al.</i> (2004)***
SRA	Borneo	20 to 417	Increased with declining soil P	Ushio <i>et al.</i> (2015)
	Northeastern Amazonia	3 to 36*		Metcalf <i>et al.</i> (2008)
Fine root length	Costa Rica, Panama, Peru	200 to 1552	Increased with declining soil P	Powers <i>et al.</i> (2005)
	Hawaii	280 to 980		Ostertag (2001)
	Northwest Borneo	83 to 151		Kochsiek <i>et al.</i> (2013)
	Atlantic forest, Brazil	5.4 to 8.9*		Zangaro <i>et al.</i> (2008)
Fine root tissue density	New Zealand	108 to 804	Increased with declining soil P	Holdaway <i>et al.</i> (2011); Richardson <i>et al.</i> (2004)***
	Borneo	20 to 417	Unchanged	Ushio <i>et al.</i> (2015)
Root diameter	New Zealand	108 to 804	Declined with declining soil P	Holdaway <i>et al.</i> (2011); Richardson <i>et al.</i> (2004)***
	Atlantic forest, Brazil	5.4 to 8.9*		Zangaro <i>et al.</i> (2008)
	Borneo	20 to 417		Ushio <i>et al.</i> (2015)

**Table S1.** Continuation.

Strategy/trait	Place	Soil P <sub>t</sub> (mg kg <sup>-1</sup> )	Results	Reference
AMF root colonization	Hawaii	280 to 980	Unchanged	Treseder and Allen (2002); Olander and Vitousek (2000)***
	Costa Rica	665 and 1601		Nasto <i>et al.</i> (2014)
	Atlantic forest	5.4 to 8.9*	Increased with declining soil P	Zangaro <i>et al.</i> (2008)
AMF Abundance in soil	Costa Rica, Panama, Peru	1552 to 600	Unchanged	Powers <i>et al.</i> (2005)
	Hawaii	280 to 980	Declined with declining soil P	Treseder and Allen (2002); Olander and Vitousek (2000)***
	Australian Dune Chronosequence	456 to 4		Teste <i>et al.</i> (2016)
	Hawaii	280 to 980		Balser <i>et al.</i> (2005)
	Northeastern Amazonia	7 to 600		Soong <i>et al.</i> (2020)
Root PME activity	Costa Rica	665 and 1601	Increased with declining soil P	Nasto <i>et al.</i> (2014)
	Borneo	20 to 417		Ushio <i>et al.</i> (2015)
	Puerto Rico	60 to 570		Cabugao <i>et al.</i> (2017); Cabugao <i>et al.</i> (2021)
	Panama	1.1 to 19.4**		Guilbeault-Mayers <i>et al.</i> (2020)
Root LMWOA exudation	Borneo	20 to 417	Increased with declining soil P	Aoki <i>et al.</i> (2012)

**Table S2. Phosphorus concentrations in leaf litter in different regions of Amazonia.** \*Species-level measurements; unmarked references refer to community-level measurements. Although soil P concentrations were not reported in some of the studies in the north, east, and central Amazonia, these areas are considered to have very low P concentrations overall (Quesada et al., 2010). Note that these studies used different methodological approaches, e.g., litter collection timing from traps, most notably, which can affect results as litter can be rapidly decomposed.

Amazon forest region	Min (mg g <sup>-1</sup> )	Max (mg g <sup>-1</sup> )	Soil P <sub>t</sub> (mg kg <sup>-1</sup> )	Author
Northeast (French Guiana)	0.18 ± 0.01	0.56	10	Fanin <i>et al.</i> (2012)*
Northeast (French Guiana)	0.09	0.6	23	Hättenschwiler <i>et al.</i> (2008)*
North (Roraima, BR)	0.4	0.6	61	Scott <i>et al.</i> (1992)
Northwest (Colombia)	0.1	0.4	40 to 480	Lips and Duivenvoorden (1996)
North (Venezuela)	0.2	0.5	Not reported	Cuevas and Medina (1986)
East (Pará, BR)	0.2	0.25	Not reported	Hayashi <i>et al.</i> (2012)
East (Pará, BR)	0.41	0.75	Not reported	Dantas and Phillipson (1989)
Central (Manaus, BR)	0.2	0.6	Not reported	Klinge (1977)



**Table S3. Summary table of results from P<sub>i</sub> fertilization studies.** Total soil P concentrations refer to control plots. \*Soil P<sub>t</sub> concentration data.

Indicators	Location	Soil P <sub>t</sub> (mg kg <sup>-1</sup> )	Results	References
Leaf Pi-resorption proficiency	Indonesia	80 to 237	Decreased	Mirmanto <i>et al.</i> (1999)
	Hawaii	280		Vitousek (1998); Olander and Vitousek (2000)*
	Panama	600	Tended to decrease	Mayor <i>et al.</i> (2014); Wright <i>et al.</i> (2011)*
	Hawaii	980	Unchanged	Vitousek (1998); Olander and Vitousek (2000)*
	Costa Rica	1690		Alvarez-Clares and Mack (2015)
Specific root length	Central Amazonia	85	Unchanged	Lugli <i>et al.</i> (2021)
	Subtropical China	410		Liu <i>et al.</i> (2015)
	Panama	600		Wurzburger and Wright (2015); Wright <i>et al.</i> (2011)*
Fine root length	Panama	600	Tended to increase	Wurzburger and Wright (2015); Wright <i>et al.</i> (2011)*
	Hawaii	700 and 280		Ostertag (2001)
	Ecuador	450 to 525	Unchanged	Camenzind <i>et al.</i> (2016); Dietrich <i>et al.</i> (2016)*
	Subtropical China	410	Decreased	Liu <i>et al.</i> (2015)
Fine root tissue density	Central Amazonia	85	Unchanged	Lugli <i>et al.</i> (2021)
	Subtropical China	410		Liu <i>et al.</i> (2015)
	Panama	600		Wurzburger and Wright (2015); Wright <i>et al.</i> (2011)*
Diameter	Central Amazonia	85	Increased	Lugli <i>et al.</i> (2021)
	Subtropical China	410	Unchanged	Liu <i>et al.</i> (2015)
AMF root colonization	Central Amazonia	85	Unchanged	Lugli <i>et al.</i> (2021)
	Ecuador	344 ± 31		Camenzind <i>et al.</i> (2014)
	Ecuador	450 to 525		Camenzind <i>et al.</i> (2016); Dietrich <i>et al.</i> (2016)*
	Panama	600	Increased	Wurzburger and Wright (2015); Wright <i>et al.</i> (2011)*
	Hawaii	280	Tended to decrease	Treseder and Allen (2002); Olander and Vitousek (2000)*
	Subtropical China	410		Liu <i>et al.</i> (2015)
	Hawaii	980		Treseder and Allen (2002); Olander and Vitousek (2000)*

**Table S3.** Continuation.

Indicators	Location	Soil P <sub>t</sub> (mg kg <sup>-1</sup> )	Results	References
AMF abundance in soil	Hawaii	280	Increased	Treseder and Allen (2002); Olander and Vitousek (2000)*
	Ecuador	450 to 525		Camenzind <i>et al.</i> (2016); Dietrich <i>et al.</i> (2016)*
	Panama	600	Decreased	Sheldrake <i>et al.</i> (2018); Wright <i>et al.</i> (2011)*
	Hawaii	980		Treseder and Allen (2002); Olander and Vitousek (2000)*
Root PME activity	Central Amazonia	85	Decreased	Lugli <i>et al.</i> (2021)
	Hawaii	980 to 280		Treseder and Vitousek (2001)
	Borneo	1.0 and 1.5 (Bray P)		Yokoyama <i>et al.</i> (2017)

**Table S4. Root morphological traits in Amazonia, other tropical forests, and meta-analyses.** Root morphological traits: specific root length (SRL,  $\text{km kg}^{-1}$ ), specific root area (SRA,  $\text{m}^2 \text{kg}^{-1}$ ), and root tissue density (RTD,  $\text{mg cm}^{-3}$ ). N/A, not applicable (meta-analysis). \*Values refer to Mehlich-1 P.

Trait	Region	Soil $P_t$ ( $\text{mg kg}^{-1}$ )	Trait variation	References
SRL	Eastern Amazon	3 to 36*	8 to 10	Metcalfe <i>et al.</i> (2008)
SRL	Central Amazon	118 to 217.4	5.9 to 41.5	Lugli <i>et al.</i> (2020)
SRL	Tropical forest in Panama	443	8 to 19.89	Wurzburger and Wright (2015)
SRL	Tropical Atlantic forest	8.9*	18.3	Zangaro <i>et al.</i> (2008)
SRL	Peruvian elevation gradient	628 to 1154	20.3 to 39.8	Girardin <i>et al.</i> (2013)
SRL	Tropical forests	N/A	7.4 to 79.3	Addo-Danso <i>et al.</i> (2020)
SRL	Tropical forests	N/A	12.2	Jackson <i>et al.</i> (1996)
SRA	Eastern Amazon	3 to 36*	24 to 34	Metcalfe <i>et al.</i> (2008)
SRA	Central Amazon	118 to 217.4	14 to 56	Lugli <i>et al.</i> (2020)
SRA	Amazon elevation gradient	628 to 1154	44 to 76	Girardin <i>et al.</i> (2013)
SRA	Tropical forests	N/A	7.9 to 87.9	Addo-Danso <i>et al.</i> (2020)
RTD	Central Amazon	118 to 217.4	141.78 to 419.22	Lugli <i>et al.</i> (2020)
RTD	Tropical forest Borneo	20 to 417	~280	Ushio <i>et al.</i> (2015)
RTD	Tropical forest Panama	443	71.8 to 328.1	Wurzburger and Wright (2015)
RTD	Tropical forests	N/A	130 to 680	Addo-Danso <i>et al.</i> (2020)

**Table S5. Potential root acid phosphomonoesterase activity (PME) in Amazonia and other tropical forests.** PME activity is given in  $\mu\text{mol}$  of substrate cleaved  $\text{g}^{-1}$  root  $\text{hr}^{-1}$ . Para-nitrophenyl phosphate (pNPP); 4-methylumbelliferyl phosphate (MUF). Note that phosphatase activity seasonally fluctuates with changes in precipitation (Turner & Wright, 2014), which may confound comparisons among one-point-in-time measurements. \*Approximate mean values for  $\text{N}_2$ -fixing and non-fixing species, respectively.

Location	PME activity	Substrate	Soil $\text{P}_t$ ( $\text{mg kg}^{-1}$ )	References
Mount Kinabalu, Borneo	94 to 180	pNPP	35 to 92	Kitayama (2013)
Central Amazon	40.8	MUF	85	Lugli <i>et al.</i> (2021)
Central Amazon	36.05	MUF	148	Lugli <i>et al.</i> (2020)
Mount Kinabalu, Borneo	118 to 164	pNPP	123 to 170	Kitayama (2013)
Puerto Rico	60	pNPP	170	Cabugao <i>et al.</i> (2017)
Mount Kinabalu, Borneo	105	pNPP	274	Kitayama (2013)
Puerto Rico	36	pNPP	290	Cabugao <i>et al.</i> (2017)
Puerto Rico	25	pNPP	410	Cabugao <i>et al.</i> (2017)
Pacific, Costa Rica	~8.6 and ~6.8*	MUF	665	Nasto <i>et al.</i> (2014)
Caribbean, Costa Rica	~6.6 and ~5.3*	MUF	1601	Nasto <i>et al.</i> (2014)

**Table S6. Plant and mycorrhizal responses to eCO<sub>2</sub> in meta-analyses;** AMF (Arbuscular mycorrhizas), ECM (Ectomycorrhizas). \*Mycorrhizal abundance refers to a group of indices that include percent root length/tip colonized, spore count, and extraradical hyphal length.

Mycorrhiza	Measure	eCO <sub>2</sub> Effect (%)	References
AMF and ECM	Mycorrhizal abundance*	+47	Treseder (2004)
AMF and ECM	Root colonization	+36	Treseder (2004)
AMF	Fungal response ratio	+21	Alberton <i>et al.</i> (2005)
ECM	Fungal response ratio	+34	Alberton <i>et al.</i> (2005)
AMF	Host-plant response ratio	+25	Alberton <i>et al.</i> (2005)
ECM	Host-plant response ratio	+26	Alberton <i>et al.</i> (2005)
AMF and ECM	Host-plant biomass	+20 ± 3	Terrer <i>et al.</i> (2016)
AMF	Host-plant biomass	+7 ± 4	Terrer <i>et al.</i> (2016)
ECM	Host-plant biomass	+30 ± 3	Terrer <i>et al.</i> (2016)
AMF and ECM	Host-plant biomass	+26	Dong <i>et al.</i> (2018)
AMF and ECM	Extraradical hyphal length	+23	Dong <i>et al.</i> (2018)
AMF and ECM	Colonization length	+15	Dong <i>et al.</i> (2018)
AMF and ECM	Fungal biomass	+22	Dong <i>et al.</i> (2018)
AMF	Fungal biomass	+7	Dong <i>et al.</i> (2018)
ECM	Fungal biomass	+30	Dong <i>et al.</i> (2018)
AMF	Host-plant biomass	+34	Dong <i>et al.</i> (2018)
ECM	Host-plant biomass	+20	Dong <i>et al.</i> (2018)
AMF	Host-plant N and P content	+22 and +19	Dong <i>et al.</i> (2018)
ECM	Host-plant N and P content	-4 and -13	Dong <i>et al.</i> (2018)

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